

1 **Reduced representation sequencing to understand the evolutionary history of Torrey pine**  
2 **(*Pinus torreyana* Parry) with implications for rare species conservation**

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## Abstract

Understanding the contribution of neutral and adaptive evolutionary processes to population differences is often necessary for better informed management and conservation of rare species. In this study, we focused on *Pinus torreyana* Parry (Torrey pine), one of the world's rarest pines, endemic to one island and one mainland population in California. Small population size, low genetic diversity, and susceptibility to abiotic and biotic stresses suggest Torrey pine may benefit from inter-population genetic rescue to preserve the species' evolutionary potential. We leveraged reduced representation sequencing to tease apart the respective contributions of stochastic and deterministic evolutionary processes to population differentiation. We applied these data to model spatial and temporal demographic changes in effective population sizes and genetic connectivity, to assess loci possibly under selection, and evaluate genetic rescue as a potential conservation strategy. Overall, we observed exceedingly low standing variation reflecting consistently low effective population sizes across time and limited genetic differentiation suggesting maintenance of gene flow following divergence. However, genome scans identified more than 2000 SNPs candidates for divergent selection. Combined with previous observations indicating population phenotypic differentiation, this indicates that natural selection has likely contributed to population genetic differences. Thus, while reduced genetic diversity, small effective population size, and genetic connectivity between populations suggest genetic rescue could mitigate the adverse effect of rarity, divergent selection between populations indicates that genetic mixing could disrupt adaptation. Further work evaluating the fitness consequences of inter-population admixture is necessary to empirically evaluate the trade-offs associated with genetic rescue in Torrey pine.

**Keywords:** Rare species conservation, demographics, adaptation, genome scans, genetic rescue.

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## Introduction

Conservation biology aims to preserve rare species and determine the appropriate management strategies necessary for long-term persistence and maintenance of evolutionary potential (Di Santo & Hamilton, 2020; Ralls et al., 2018; Swarts, Sinclair, Krauss, & Dixon, 2009; Young, Brown, & Zich, 1999). Rare species may have reduced effective population sizes ( $N_e$ ), impeding populations' ability to adapt to change (e.g.,  $N_e < 1000$ ) or increasing probability of inbreeding (e.g.,  $N_e < 100$ ) (Frankham, Bradshaw, & Brook, 2014), ultimately increasing the risk of local extirpation. Combined, rarity and isolation are often associated with stochastic loss of genetic variation (Aguilar, Quesada, Ashworth, Herrerias-Diego, & Lobo, 2008; Hague & Routman, 2016; Young, Boyle, & Brown, 1996). Genetic rescue is one conservation strategy that has been successfully used in both animals and plants to mitigate consequences of severely reduced genetic diversity (Bossuyt, 2007; Hedrick, Peterson, Vucetich, Adams, & Vucetich, 2014; W. E. Johnson et al., 2010; Madsen, Shine, Olsson, & Wittzell, 1999; Westemeier et al., 1998; Willi, Van Kleunen, Dietrich, & Fischer, 2007). Genetic rescue introduces or restores gene flow between populations to alleviate the fitness consequences of inbreeding through the introduction of genetic variation. However, while rare species may exhibit small effective population sizes and reduced adaptive potential, disruption of local adaptation may lead to outbreeding depression, or reduced fitness of progeny following admixture between genetically differentiated lineages (Hufford & Mazer, 2003). Thus, the contribution of natural selection to the evolution of population genetic differences is a consideration for genetic rescue, as it may ultimately lead to migrants or translocated individuals being maladapted (Lowry, Rockwood, & Willis, 2008; Nosil, Vines, & Funk, 2005). For rare species conservation, an understanding of contemporary effective population size is therefore required to assess immediate genetic threats

58 to population persistence. Following these threats, however, understanding the history of  
59 population connectivity, the distribution of genetic variation within and among populations, and  
60 the role of selection in shaping population differences will be critical to informed management  
61 decisions.

62 In addition to guiding conservation management strategies, understanding rare species'  
63 demographic and evolutionary history may prove valuable to optimizing strategies necessary to  
64 preserve neutral and nonneutral genetic diversity *ex situ*. *Ex situ* conservation collections, or the  
65 preservation of species outside their natural range of occurrence, can complement *in situ*  
66 conservation strategies (Cavender et al., 2015; Potter et al., 2017; Pritchard, Fa, Oldfield, &  
67 Harrop, 2012), providing a critical resource for the preservation of genetic variation, restoration  
68 or reintroduction (Guerrant Jr, Havens, & Vitt, 2014; Potter et al., 2017). *Ex situ* sampling  
69 designs traditionally rely on neutral population genetic structure to guide sampling decisions  
70 (Caujapé-Castells & Pedrola-Monfort, 2004; Gapare, Yanchuk, & Aitken, 2008; Hoban, 2019;  
71 Hoban & Schlarbaum, 2014). However, concerns exist regarding the sole use of neutral genetic  
72 variability for species conservation, as variation at neutral loci is unlikely to reflect adaptive  
73 genetic diversity (Bonin, Nicole, Pompanon, Miaud, & Taberlet, 2007; Holderegger, Kamm, &  
74 Gugerli, 2006; McKay & Latta, 2002; Teixeira & Huber, 2021). *Ex situ* population sampling  
75 may need to evaluate the impact different evolutionary processes have had on population genetic  
76 structure to optimize neutral and adaptive variation collected. Thus, an understanding of  
77 population connectivity and the impact of selection across populations can inform *ex situ*  
78 collection design. If empirical or simulated data suggest populations are genetically connected  
79 and genetic differentiation is low, then most neutral genetic variation may be captured within one  
80 or a few populations. However, if selection overcomes the homogenizing effects of gene flow,

81 ensuring adaptive genetic differences are preserved for all populations will require the inclusion  
82 of diverse population origins, separation of such populations *ex situ*, and consideration of  
83 population origin in breeding programs.

84 With the advent of next-generation sequencing (NGS) and the ever-decreasing costs  
85 associated with these technologies, genome-wide estimates of genetic diversity can be readily  
86 assessed and used to guide conservation management strategies. Combined with statistical and  
87 simulation-based tools, these data provide a powerful and timely means to evaluate aspects of  
88 population genetic variation and spatial genetic structure critical to informing genetic rescue and  
89 *ex situ* conservation plans, including both populations' demographic and adaptive history  
90 (Abebe, Naz, & Léon, 2015; Liu, Zhang, Wang, & Ma, 2020; Wang, Bernhardsson, &  
91 Ingvarsson, 2020; Xia et al., 2018). However, in conifers, despite extensive use in characterizing  
92 genomes as well as neutral and adaptive variation (Eckert et al., 2010; Namroud, Beaulieu, Juge,  
93 Laroche, & Bousquet, 2008; Nystedt et al., 2013; Stevens et al., 2016; Tyrmi et al., 2020; Wang  
94 et al., 2020), these data have only rarely been used to inform conservation management  
95 decisions.

96 Torrey pine (*Pinus torreyana* Parry) is a critically endangered pine (IUCN, 2021),  
97 endemic to California. One of the rarest pine species in the world (Critchfield & Little, 1966;  
98 Dusek, 1985), Torrey pine's distribution spans one island population (*Pinus torreyana* subsp.  
99 *insularis*) of approximately 3,000 reproductive individuals (Santa Rosa Island, CA), and one  
100 mainland population (*Pinus torreyana* subsp. *torreyana*) of approximately 4,000 reproductive  
101 individuals (Torrey Pine State Reserve in La Jolla, CA) (J. Franklin & Santos, 2011; Hall &  
102 Brinkman, 2015). In addition to low population size, and despite current *in situ* and *ex situ*  
103 conservation efforts, Torrey pine suffers from exceedingly low genetic variation and faces both

104 anthropogenic and environmental disturbances (J. Franklin & Santos, 2011; Hamilton, Royauté,  
105 Wright, Hodgskiss, & Ledig, 2017; Ledig & Conkle, 1983; Waters & Schaal, 1991; Whittall et  
106 al., 2010). For these reasons, the species may be at imminent risk for population-scale extirpation  
107 events and thus a potential candidate for genetic rescue. Inter-population admixture may increase  
108 population genetic diversity, alleviating potential fitness consequences associated with Torrey  
109 pine's low genetic diversity (Hamilton et al., 2017), and increase evolutionary potential  
110 necessary to respond to current and future ecological challenges (Carlson, Cunningham, &  
111 Westley, 2014). Previous research observed heterosis following one generation of admixture  
112 between island and mainland individuals, suggesting that genetic rescue may alleviate fitness  
113 consequences associated with reduce genetic variation (Hamilton et al., 2017). However, if  
114 adaptive genetic differences have evolved between island and mainland populations, fitness  
115 consequences following the disruption of co-adapted gene complexes may not be observed in the  
116 first generation cross. Thus, although the combination of exceedingly low genetic diversity and  
117 conservation status suggest Torrey pine may be a candidate for genetic rescue, evaluation of the  
118 species' demographic and adaptive evolutionary history will be necessary prior to inform  
119 conservation management decisions.

120         With this study, we use genomic data to quantify and model aspects of populations'  
121 demographic and evolutionary history necessary to preserve rare species' evolutionary potential.  
122 Specifically, we delineate the contribution of stochastic and deterministic processes to genomic  
123 differentiation in *Pinus torreyana*, asking three questions: (i) what are current effective  
124 population sizes and how have they changed over time, (ii) have populations remained  
125 genetically connected following isolation, and (iii) is there evidence of adaptive divergence  
126 between populations that may indicate distinct evolutionary trajectories? We identify the most

127 probable demographic scenario for Torrey pine using Approximate Bayesian Computing (ABC)  
128 and coalescent simulations. To evaluate the role of selection, we identify loci that may be  
129 important to adaptation using various  $F_{ST}$  outlier methods and assess the functional significance  
130 of these loci by annotating candidates with the gene ontology (GO) resource. Using this  
131 knowledge, we discuss conservation strategies for Torrey pine, including *ex situ* sampling to  
132 preserve neutral and nonneutral processes within species collections, and possible risks  
133 associated with genetic rescue. This study demonstrates the benefits and necessity of  
134 understanding the demographic and adaptive history of rare species to guide conservation.

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## Material and Methods

### 137 Population sampling and DNA extraction

138 Between June and July 2017, needle tissue was collected from individuals spanning the  
139 entire natural distribution of *Pinus torreyana* (Torrey pine; Fig. 1). A total of 286 individuals  
140 were sampled, including 146 individuals from the mainland population at the Torrey Pine State  
141 Reserve (TPSR) near La Jolla, CA and 140 individuals from the island population on Santa Rosa  
142 Island, CA (SRI), one of the Channel Islands. Needles were dried in silica gel following which  
143 genomic DNA was extracted using between 25-30 mg of dry needle tissue and a modified CTAB  
144 protocol (Doyle & Doyle, 1987). To reduce DNA shearing, slow manual shaking of tubes was  
145 used. Following extraction, the concentration and purity of DNA extracted was quantified for  
146 each sample using a NanoDrop 1000 Spectrophotometer (Thermo scientific) to ensure all  
147 samples had a concentration of approximately 85 ng/ $\mu$ l and purity ratios of 1.4 and above  
148 (average across samples; 260/280 = 1.85, 260/230 = 1.96).

### 149 Genomic library preparation and ddRAD sequencing

150 Genomic libraries were prepared for all 286 individuals following the protocol of  
151 Parchman & colleagues (2012). Briefly, 510 ng (6  $\mu$ l at 85 ng/ $\mu$ l) of DNA was digested using  
152 endonucleases *Eco*RI and *Mse*I (New England BioLabs, Inc.) after which barcoded (*Eco*RI cut  
153 site) and non-barcoded (*Mse*I cut site) adapters compatible with Illumina sequencing were  
154 ligated to each end of DNA fragments using T4 ligase (New England BioLabs, Inc.). A different  
155 barcode sequence was used for each of the 286 samples. Due to the large, highly repetitive nature  
156 of pines' genome (Stevens et al., 2016), we used the methylation-sensitive enzyme *Eco*RI, as it  
157 effectively reduces the presence of repetitive and non-coding DNA sequences in genomic  
158 libraries (Parchman et al., 2012). Restriction-ligation products were amplified using two  
159 successive PCR reactions to produce genomic libraries with concentrations necessary for  
160 sequencing (> 2nM). All PCR-amplified genomic libraries were subsequently pooled and sent to  
161 the Genomic Sequencing and Analysis Facility (GSAF; Austin, TX) for size selection of  
162 fragments within the range of 450-500 bp and sequenced on 5 lanes of an Illumina HiSeq 2500  
163 using the 100 bp single-end sequencing protocol (1 x 100 bp).

#### 164 **De novo assembly and SNP calling**

165 Demultiplexing of sequence files was performed using *ipyrad* v0.9.12 (Eaton & Overcast,  
166 2020) allowing one mismatch in the barcode sequence. Reads were filtered, assembled *de novo*,  
167 and used to call SNPs within the *dDocent* v2.7.8 pipeline (Puritz, Hollenbeck, & Gold, 2014;  
168 Puritz, Matz, et al., 2014). Reads were filtered by removing low-quality bases at the beginning  
169 and end of reads (PHRED score < 20), Illumina adapters, and trimmed when the average  
170 PHRED score fell below 10 within a 5 bp window using the program TRIMMOMATIC (Bolger,  
171 Lohse, & Usadel, 2014). As a contiguous genome assembly for *P. torreyana* is not available, we  
172 first generated a reference of genomic regions sampled with our sequencing design using a *de*



173 *de novo* approach. Reads were clustered based on sequence similarity and assembled into a  
174 reference assembly using the program CD-HIT (Fu, Niu, Zhu, Wu, & Li, 2012; W. Li & Godzik,  
175 2006). To be included in *de novo* assembly, reads had to have a minimum of 3x within-  
176 individual coverage and be present in at least 5 individuals. To form a cluster (locus), reads had  
177 to have a minimum of 86% sequence similarity, a cutoff previously used in pines (Menon et al.,  
178 2018). This threshold was chosen as a tradeoff to avoid the clustering of paralogous loci while  
179 still accounting for the presence of missing bases, errors, or polymorphisms between true  
180 homologous sequences. Finally, reads were mapped onto *de novo* assembled loci using BWA  
181 MEM (H. Li, 2013) and SNPs were called using the software FREEBAYES (Garrison & Marth,  
182 2012). Read mapping was performed using BWA default parameters, including a match value of  
183 1, a mismatch penalty of 4 and a gap penalty of 6. This yielded a set of 652,492 SNPs that were  
184 subjected to downstream filtering. Variants with genotype quality (GQ) < 20 and genotype depth  
185 < 3 were first marked as missing. Then, variants with PHRED scores (QUAL)  $\leq$  30, minor allele  
186 counts < 3, minor allele frequencies < 0.01, call rate across all individuals < 0.95, mean depth  
187 across samples > 57 (based on the equation:  $d + 4\sqrt{d}$ , where  $d$  is the average read depth across  
188 variants, H. Li, (2014),  $F_{IS}$  estimates < -0.5 or > 0.5, and linkage score ( $r^2$ ) > 0.5 within a 95 bp  
189 window were removed from the raw SNP dataset. Following filtering, a total of 93,085 biallelic  
190 SNPs were kept and used for analysis (hereafter referred to as the full dataset). Note that 16  
191 individuals with > 40% missing data were also discarded, leaving 270 genotyped individuals  
192 (SRI: 130 individuals, TPSR: 140 individuals) for inclusion in analyses.

### 193 **Population structure and genetic diversity analyses**

194 To describe and quantify contemporary genetic differences between Torrey pine  
195 populations, we first assessed genetic structure of populations using principal component

196 analysis (PCA) implemented in the R package ADEGENET (Jombart, 2008; Jombart & Ahmed,  
197 2011). Unless otherwise stated, analyses were performed in R version 3.6.2 and 4.0.2 (R Core  
198 Team, 2019, 2020). To quantify genetic differences between island and mainland populations,  
199 we calculated Nei's  $F_{ST}$  statistic (Nei, 1987) for each SNP using the HIERFSTAT R package  
200 (Goudet & Jombart, 2020) and averaged estimates across loci. A 95% confidence interval around  
201 the mean was constructed by bootstrapping the empirical  $F_{ST}$  distribution 10,000 times using R  
202 packages BOOT (Canty & Ripley, 2021; Davison & Hinkley, 1997) and SIMPLEBOOT (Peng,  
203 2019).

204 To determine the extent of contemporary within-population genetic structure, the most  
205 likely number of genetic clusters was independently evaluated for SRI and TPSR using the  
206 function *find.clusters()* implemented in the R package ADEGENET. This function transforms  
207 genomic data using principal component analysis and performs successive K-means clustering  
208 with an increasing number of clusters ( $k$ ). For each successive value of  $k$ , the Bayesian  
209 Information Criterion (BIC) is computed and was used to assess the optimal number of clusters  
210 ( $k$ ). For each population, we assessed between 1 to 10 clusters while retaining principal  
211 components necessary to explain 90% of the variation after ordination (SRI: 114, TPSR: 122).  
212 For TPSR, two individuals (TPSR5107, TPSR3189) clustered distantly from the population,  
213 which may mask subtle within-population genetic structure. Consequently, we re-ran the analysis  
214 excluding the two individuals while maintaining the same range for  $k$  (1 to 10) and retaining 121  
215 principal components (90.38% of variation explained after ordination).

216 To evaluate Torrey pine evolutionary potential, contemporary standing genetic variation  
217 within the species was estimated by calculating expected heterozygosity ( $H_E$ ), inbreeding  
218 coefficients ( $F_{IS}$ ), and coancestry coefficients ( $\theta$ ) for both island and mainland populations

219 independently. Values of  $H_E$  and  $F_{IS}$  were calculated for each SNP separately using the R  
220 package ADEGENET and averaged across loci to provide population-level estimates. To  
221 evaluate  $\theta$ , we used the R package RELATED (Pew, Wang, Muir, & Frasier, 2015) that  
222 estimates genetic relatedness between all possible pairs of individuals within a population.  
223 Specifically, we used the triadic likelihood (TrioML) estimate of relatedness assuming no  
224 inbreeding within populations. Averaging pairwise  $\theta$  values across all individuals within TPSR  
225 and SRI provided population estimates of genetic relatedness. To build 95% confidence intervals  
226 around  $H_E$ ,  $F_{IS}$  and  $\theta$  averages, the empirical distribution for each parameter within each  
227 population was bootstrapped 10,000 times in R using the BOOT and SIMPLEBOOT packages.

## 228 **ABC demographic modeling**

229 *Demographic models* – To evaluate the impact genetic drift and gene flow have had to  
230 patterns of neutral genetic variation in Torrey pine, we quantified changes in effective population  
231 size over time, interpopulation migration rate, and time since population divergence. To do so,  
232 we tested six distinct demographic models that were classified into two broad categories: (1)  
233 isolation with/without migration (Fig. 2A), and (2) two-population demic expansion (Fig. 2B, C).

234 Models of isolation with or without migration (RPM1, RPM2) were developed to test a  
235 hypothesis formulated by Ledig & Conkle, (1983). This hypothesis predicts that there was once a  
236 single ancestral population of Torrey pine that diverged to form one island and one mainland  
237 population following tectonic movement (Fig. 2A). These models assume that an ancestral  
238 population with an effective size  $N_A$  diverged  $T_{Div}$  generations before present to form two  
239 populations with current effective sizes  $N_I$  (island, SRI) and  $N_M$  (mainland, TPSR). Following  
240 divergence, to assess whether gene flow has occurred between populations, bidirectional

241 migration was either prevented (RPM1,  $m_{IM}=m_{MI}=0$ ) or permitted (RPM2,  $m_{IM}=m_{MI}>0$ ). Both  
242 models assume constant island and mainland effective population size following divergence.

243 The remaining four models (CM1, CM2, CM3, CM4) tested two different hypotheses of  
244 land colonization where one population was founded by the other (Fig. 2B, C). Models CM1 and  
245 CM2 specifically test the hypothesis that Santa Rosa Island was colonized by a subset of  
246 mainland individuals (Ledig & Conkle, 1983). In this scenario, SRI was founded by  $N_C$  effective  
247 migrants from TPSR  $T_{Div}$  generations ago and grew exponentially at a rate  $r_I$  ( $r_I = \log\left(\frac{N_C}{N_I}\right)/T_{Div}$ )  
248 to form a population with an effective size  $N_I$ . TPSR effective population size ( $N_M$ ) was assumed  
249 constant. As above, bidirectional migration between populations was either prevented (CM1,  
250  $m_{IM}=m_{MI}=0$ ) or permitted (CM2,  $m_{IM}=m_{MI}>0$ ) to evaluate whether gene flow has occurred  
251 between populations following colonization. Models CM3 and CM4 test the hypothesis that the  
252 mainland population was founded by a subset of island individuals (Haller, 1986). This scenario  
253 assumes TPSR was founded by  $N_C$  effective migrants from SRI  $T_{Div}$  generations before present.  
254 SRI effective population size ( $N_I$ ) was assumed constant while TPSR effective population size  
255 ( $N_M$ ) grew exponentially at rate  $r_M$  ( $r_M = \log\left(\frac{N_C}{N_M}\right)/T_{Div}$ ). Once again, to test whether gene flow  
256 has occurred between populations since colonization, exchange of migrants between population  
257 was either prevented (CM3,  $m_{IM}=m_{MI}=0$ ) or permitted (CM4,  $m_{IM}=m_{MI}>0$ ).

258 For all six models, uniform priors were used except for  $T_{Div}$ ,  $N_C$ , and  $N_A$  for which we  
259 used log-uniform priors. Priors on a logarithmic scale increases the weight given to small values  
260 and is recommended when parameters' ranges span several orders of magnitude (Wegmann,  
261 Leuenberger, & Excoffier, 2009). Details on demographic parameters and their prior  
262 distributions are provided in Table 1.

263            *Additional filtering and down-sampling of genetic variants* – Accurate estimation of  
264 demographic parameters using coalescence simulations requires the use of neutrally evolving  
265 genetic markers. To ensure neutrality of SNPs, we filtered the full dataset for SNPs that did not  
266 deviate significantly from Hardy-Weinberg equilibrium (HWE). Since population structure may  
267 create departures from HWE, we applied this filter to island and mainland populations  
268 independently and removed SNPs that deviated significantly from HWE ( $P < 0.05$ ) in both  
269 populations. This was performed using a customized R function relying on R packages  
270 ADEGENET and PEGAS (Paradis, 2010). In total, 73,928 SNPs were retained following  
271 filtering for use in demographic modelling (hereafter referred to as the HWE-filtered dataset).

272            For computational efficiency, we down sampled the HWE-filtered dataset from 73,928 to  
273 9,795 variants first by generating bivariate bins based on observed heterozygosity and Nei's  $F_{ST}$   
274 (0.05-interval bins), and then by subsampling each bin proportionally to the number of SNPs  
275 they contained. In this way, each bin is subsampled to reflect its contribution to the HWE-filtered  
276 dataset (Appendix S1). Following subsampling, we conducted a principal component analysis  
277 using the down-sampled dataset to ensure patterns of genetic diversity and population structure  
278 were maintained between datasets (Appendix S2).

279            *Generating coalescent simulations and estimating summary statistics* – To evaluate and  
280 compare demographic scenarios, a set of 200,000 simulations was generated using  
281 ABCSAMPLER for each model, a wrapper program included in the package ABCTOOLBOX  
282 (Wegmann, Leuenberger, Neuenschwander, & Excoffier, 2010). For each simulation,  
283 ABCSAMPLER samples prior ranges of demographic parameters and uses these values as inputs  
284 for coalescence simulations within a user-defined simulation program. We used  
285 FASTSIMCOAL version 2.6.0.3 (Excoffier, Dupanloup, Huerta-Sánchez, Sousa, & Foll, 2013;

286 Excoffier & Foll, 2011) to simulate 9,795 unlinked SNPs with a minor allele frequency of 0.01,  
287 reflecting the composition of the down-sampled genetic dataset. For each model, simulated data  
288 were output as genotypes and fed to a user-defined program by ABCSAMPLER to estimate  
289 population genetic summary statistics. ARLEQUIN version 3.5.2.2 (Excoffier & Lischer, 2010)  
290 was used to calculate ten distinct population genetic summary statistics, specifically aiming at  
291 quantifying genetic variation and divergence within and between Torrey pine populations. These  
292 included genetic diversity (i.e., population-specific heterozygosity and number of alleles, average  
293 heterozygosity and number of alleles over loci and populations, and mean total heterozygosity),  
294 genetic differentiation (i.e., pairwise  $F_{ST}$ ), and variance (i.e., standard deviation over populations  
295 of the average heterozygosity and number of alleles) statistics. Finally, to obtain observed  
296 population genetic summary statistics, ARLEQUIN version 3.5.2.2 was rerun using the down-  
297 sampled dataset. Characterizing and summarizing the amount and distribution of genetic  
298 variation within each dataset, these statistics were used to calculate posterior probabilities and  
299 identify the demographic model with greatest support, as well as estimate demographic  
300 parameters associated with that model (see below).

301 ***ABC parameter estimation*** – Demographic parameters were estimated in R using the  
302 ABC package (Csilléry, François, & Blum, 2012). Cross-validation simulations were conducted  
303 first to evaluate the ability of summary statistics to distinguish between demographic models  
304 (Appendix S3). We performed leave-one-out cross-validations, consisting of selecting one  
305 simulation of a demographic scenario and then assigning it to one of the six models using  
306 posterior probabilities estimated from all remaining simulations. This was repeated one hundred  
307 times for each demographic model, generating a confusion matrix. If misclassification rates  
308 (proportions of simulations incorrectly assigned to a model) are low, then computed summary

309 statistics can distinguish between our different demographic scenarios. The posterior probability  
310 of each demographic model was approximated as the proportion of accepted simulations and  
311 used to select the best model following cross-validation. To ensure the best model provided a  
312 good fit to the data, we performed a goodness-of-fit test as implemented in the function *gfit()*.  
313 Finally, demographic parameters associated with the best model were estimated as the weighted  
314 medians of posterior distributions using the *weighted.median()* function implemented in the R  
315 package SPATSTAT (Baddeley, Rubak, & Turner, 2015). Posterior distributions were created  
316 from the set of accepted simulations using a non-linear postsampling regression adjustment  
317 conducted on log-transformed data (Blum & François, 2010). Ninety-five percent confidence  
318 intervals around weighted medians were estimated using 10,000 bootstrap replicates of posterior  
319 distributions in R, using BOOT, SIMPLEBOOT and SPATSTAT packages. The validity and  
320 accuracy of each estimated parameter were tested using additional, yet distinct, 100-fold leave-  
321 one-out cross-validation simulations (Appendix S4). The cross-validation process begins with  
322 the random selection of one simulation generated by the best demographic model. Summary  
323 statistics associated with that simulation are considered as pseudo-observed data and its  
324 demographic parameters are estimated using remaining simulations for the model. If pseudo-  
325 observed parameters can accurately be predicted, then inferred demographic parameters from  
326 true observed data can be considered valid and accurate. Cross-validation simulations, model  
327 selection, model validation, and parameters estimation were conducted using a tolerance  
328 threshold of 0.01, a tradeoff between retaining a reasonable number of simulations to estimate  
329 posterior distributions and keeping the tolerance value as low as possible (S. Li & Jakobsson,  
330 2012). Finally, as census sizes for Torrey pine populations are available (J. Franklin & Santos,  
331 2011; Hall & Brinkman, 2015), we calculated the proportion of the census size (N) to effective

332 population size ( $N_e$ ) for each population separately. Of all reproductive trees present within a  
333 population (census size), this ratio estimated the proportion contributing to the next generation  
334 (effective size).

### 335 **Simulating the null $F_{ST}$ distribution**

336 To evaluate the influence natural selection may have had on the genomic structure of  
337 Torrey pine populations, we compared the distribution of Nei's  $F_{ST}$  estimated from the full SNP  
338 dataset with a simulated distribution based on 93,085 independent SNPs from 270 individuals  
339 generated using SIMCOAL2 version 12.09.07 (Laval & Excoffier, 2004). We used weighted  
340 medians estimated from posterior distributions borrowed from the best demographic model (see  
341 Results) as input parameters for neutral simulations. For each simulated SNP, the minimum  
342 allele frequency was set to 0.01 to reflect filters applied to the full dataset. Locus-specific Nei's  
343  $F_{ST}$  values were estimated for both full and simulated datasets in R using the HIERFSTAT  
344 package (Appendix S5).

### 345 **Outlier detection analyses**

346 To estimate the potential contribution of local adaptation to genomic differentiation  
347 between Torrey pine populations, we used the full dataset of 93,085 SNPs to identify loci  
348 putatively under selection using three distinct methods: BAYESCAN version 2.1 (Foll &  
349 Gaggiotti, 2008), OUTFLANK version 0.2 (Whitlock & Lotterhos, 2014) and PCADAPT  
350 version 4.3.3 (Privé, Luu, Vilhjálmsson, & Blum, 2020). These three approaches were selected  
351 for their ability to account for neutral population structure (BAYESCAN, OUTFLANK) and to  
352 handle genetic admixture between individuals (PCADAPT). For both BAYESCAN and  
353 OUTFLANK, we grouped all 270 individuals by populations (mainland: 140 individuals, island:



354 130 individuals). Below is a brief description of all three methods used to identify candidate  
355 SNPs.

356 The first approach we used was BAYESCAN. For each locus, BAYESCAN uses a  
357 Bayesian approach to decompose  $F_{ST}$  coefficients into population- and locus-specific  
358 components using a logistic regression. Loci are identified as putatively under selection if the  
359 locus-specific component is needed to explain the observed distribution of genetic diversity. Our  
360 analysis was conducted using BAYESCAN default parameters. Results were visualized and  
361 analyzed in R. Only SNPs with a false discovery rate of 1% or below were retained and  
362 considered as candidate loci. The second approach we used was OUTFLANK, an R package  
363 which identifies outliers by inferring a null  $F_{ST}$  distribution approximated from empirical  $F_{ST}$   
364 values. This distribution was produced by removing loci with an expected heterozygosity below  
365 0.1 ( $H_{min} = 0.1$ ), trimming 5% of lowest and highest empirical  $F_{ST}$  values ( $RightTrimFraction =$   
366  $LeftTrimFraction = 0.05$ ), and retaining only loci passing a built-in neutrality test with a false  
367 discovery rate of 0.1% ( $qthreshold = 0.001$ ). Using the latter threshold provided a conservative  
368 estimate for neutral  $F_{ST}$ . Outlier loci were identified by comparing the empirical  $F_{ST}$  distribution  
369 to the inferred null  $F_{ST}$  distribution using a built-in chi-squared test with a false discovery rate of  
370 1% ( $qthreshold = 0.01$ ). Note that loci with  $H_E < 0.1$  were excluded while conducting the chi-  
371 squared test and therefore could not be identified as potential outliers ( $H_{min} = 0.1$ ). The third  
372 and last approach we used was PCADAPT. Also implemented in R, PCADAPT is a package that  
373 assesses population structure using PCA. Consequently, this approach does not require  
374 individuals to be grouped into populations. Following PCA, candidate loci are identified as those  
375 substantially correlating with population structure. We ran PCADAPT retaining the first axis of  
376 differentiation ( $K = 1$ ) and SNPs with a minor allele frequency of 0.01 ( $min.maf = 0.01$ ). We

377 only considered the first principal component to calculate the test statistic, as additional axes did  
378 not ascertain population structure (Appendix S6). Candidate loci were identified as the set of  
379 SNPs with a false discovery rate of 1% or below. To minimize the presence of false positives  
380 within our dataset, only outlier SNPs common to all three approaches were considered as  
381 candidate loci and included in subsequent analyses. Finally, to both visualize and quantify  
382 genetic structure at putatively adaptive loci, we conducted a principal component analysis based  
383 only on candidate SNPs shared by all three methods using the R package ADEGENET.

#### 384 **Functional categorization of candidate loci**

385 To identify biological processes or molecular functions that may play a role in adaptation  
386 to mainland or island environments, *de novo* assembled sequences containing candidate SNPs  
387 common to all genome scans were first extracted and then annotated using BLASTx version  
388 2.6.0+. Sequences were blasted against the UniProt protein database filtered for sequences from  
389 species within the *Pinaceae* family (Taxon identifier 3318). Sequence similarity was assessed  
390 using BLASTx default parameters, an e-value hit filter of  $10^{-3}$  (-evalue 0.001), and a number of  
391 database hits to retain of 1 (-max\_target\_seqs 1). Gene ontology terms were mapped onto  
392 annotated sequences in R using the UNIPROTR package (Soudy et al., 2020).

393

## 394 **Results**

### 395 **Population genetic structure and variation**

396 Using all 93,085 SNP markers, the principal component analysis revealed little genome-  
397 wide differentiation in Torrey pine with the first two principal components (PC1, PC2)  
398 explaining only 1.9 and 0.6% of observed genetic differences, respectively (Fig. 3). Nonetheless,  
399 PC1 unambiguously separated island from mainland individuals, suggesting some level of  
400 genetic differentiation exists between populations. These results were further supported by the

401 low average coefficient of genetic differentiation found across loci (Nei's  $F_{ST}$ ), estimated at  
402 0.013 (95% CI: 0.012 - 0.013). In addition to population-scale genetic differentiation, within-  
403 population genetic differentiation was estimated to evaluate whether local inbreeding, possibly  
404 increased in small populations, could have contributed to fine-scale genetic structure in Torrey  
405 pine. For both island and mainland populations, we found no evidence of within-population  
406 genetic structure. Population-specific principal component analyses identified no clear genetic  
407 clusters and revealed that, combined, the first two axes of differentiation (PC1, PC2) only  
408 explained approximately 1.9% and 2.2% of mainland and island within-population genetic  
409 differences, respectively (Fig. 4). In addition, evaluating the likelihood between 1 to 10 genetic  
410 clusters ( $k$ ) within each population using BIC indicated that populations appear largely  
411 homogeneous (most likely  $k = 1$ ) (Appendix S7). Interestingly, average estimates of inbreeding  
412 and coancestry coefficients across loci were low for both the island and the mainland population  
413 (Table 2). While low inbreeding coefficients support the lack of observable within-population  
414 genetic structure, this also indicates that reproduction among relatives or unequal reproductive  
415 success among individuals is unlikely to have contributed to low expected heterozygosity  
416 observed within populations (Table 2). Combined, our results indicate that Torrey pine exhibits  
417 exceedingly low genetic diversity, with the majority of variation distributed within genetically  
418 unstructured populations.

### 419 **Demographic history of Torrey pine**

420 Of the six demographic models evaluated (Fig. 2), the isolation with migration model  
421 (RPM2) received the most support with a posterior probability of 92.18%. The remaining five  
422 models exhibited lower posterior probabilities ranging from 0% to 3.98%. With low  
423 misclassification rates, cross-validations indicated that simulated summary statistics were able to

424 confidently distinguish between different demographic scenarios (Appendix S3). The goodness-  
425 of-fit test revealed that simulated summary statistics did not significantly differ from observed  
426 ones ( $P = 0.76$ ), providing a good fit to the data.

427         Based on RPM2, an ancestral Torrey pine population with an effective size of  
428 approximately 1,124 individuals (95% CI: 939 - 1,213) diverged during the early Pleistocene  
429 approximately 1.2 million YBP (95% CI: 1,195,367 - 1,296,186, assuming a generation time of  
430 ten years) to form one island and one mainland population with effective sizes of approximately  
431 2,305 (95% CI: 2,166 - 2,338) and 1,715 (95% CI: 1,616 - 1,759) individuals, respectively (Fig.  
432 5). This resulted in a 0.75 ( $N_I/N = 2,305/3,063$ ) proportion of the census to effective population  
433 size on the island, and a 0.45 ( $N_M/N = 1,715/3,806$ ) proportion of the census to effective  
434 population size on the mainland. Following divergence, some gene flow was maintained between  
435 populations with an estimated migration rate of  $8.34 \times 10^{-3}$  (95% CI:  $8.15 \times 10^{-3} - 8.76 \times 10^{-3}$ ) per  
436 generation. In general, cross-validation simulations indicated low prediction errors (Appendix  
437 S4), suggesting high accuracy of inferred parameters. Nonetheless, note that for  $T_{Div}$ , the  
438 associated prediction error was higher than for other parameters.

#### 439 **Divergent selection between island and mainland populations**

440         Despite limited genetic variation within populations, we found some evidence for the  
441 evolution of genetic differences among populations at a subset of loci. We compared the neutral  
442  $F_{ST}$  distribution generated from the simulated RPM2 demographic model with the empirical  
443 distribution based on our full dataset of SNP variants (Appendix S5). For select SNPs, moderate  
444 to high empirical  $F_{ST}$  values (from approximately 0.2 to 0.65) could not be generated through  
445 neutral simulations, suggesting they may be candidate for selection. PCADAPT, BAYESCAN,  
446 and OUTFLANK identified 3,138 (3.37%), 2,163 (2.32%), and 3,942 (4.23%) outlier SNPs,

447 respectively (Fig. 6A). Of these outlier loci, 2,053 (2.21%) were common to all three methods  
448 and contribute to genomic structure between Torrey pine populations (Fig. 6B). Indeed, the  
449 principal component analysis revealed that the first axis of differentiation (PC1) unambiguously  
450 differentiated island from mainland populations based on common outlier SNPs, explaining over  
451 20% of observed genetic differences. Consistent with these results,  $F_{ST}$  values for putatively  
452 adaptive loci ranged between 0.1 and 0.63 and either could not be generated or could only be  
453 generated at low frequency through neutral simulations, representing only 0.078% of all  
454 simulated  $F_{ST}$  values.

455       Functional categorization of common outlier loci was performed by blasting *de novo*  
456 assembled contigs carrying outlier SNPs against the *Pinaceae* UniProt protein database and  
457 retrieving each hit's Gene Ontology terms. Overall, 110 (7.51%) contigs were annotated. After  
458 accounting for redundancy in the data (i.e., different contigs aligning to the same locus), we  
459 identified a total of 80 putative adaptive genes with homologous sequences in *Larix*, *Picea*, or  
460 *Pinus* species (Appendix S8) that may be targets of selection. Functionally, these genes are  
461 primarily encoded in mitochondria, involved in the process of DNA integration, or with  
462 processes associated with molecular functions such as RNA-directed DNA polymerase activity  
463 or nucleic acid binding (Table 3).

464

465

## Discussion

466       An understanding of demographic and adaptive evolutionary history is invaluable for rare  
467 species of conservation concern, particularly when management decisions impact populations at  
468 risk of extinction. Teasing apart the contribution of both stochastic and deterministic  
469 evolutionary processes to population genomic differentiation over time and space can be used to

470 inform species conservation decisions, including the potential consequences of genetic rescue  
471 (Frankham et al., 2011; Hufford & Mazer, 2003; Ralls et al., 2018). Here, we evaluated the  
472 genomics of Torrey pine, a critically endangered endemic isolated to two populations in  
473 California. We modeled demographic change and connectivity over time and tested the influence  
474 of neutral and selective processes have had on contemporary population genomic structure. We  
475 observed that Torrey pine populations exhibited exceedingly low genetic variation, particularly  
476 for a conifer (see below), with little within- or among-population structure. Although some  
477 connectivity has been maintained between island and mainland populations, demographic  
478 modeling indicates that Torrey pine has consistently suffered from low effective population size.  
479 Genome scans revealed over 2000 loci that were candidates for selection. Consistent with  
480 previous observations indicating phenotypic differences between island and mainland  
481 populations, these data suggest adaptive genetic differences have evolved among populations (Di  
482 Santo et al., *in review*; Haller, 1986; Hamilton et al., 2017). From a conservation standpoint these  
483 results lead to contrasting recommendations with respect to genetic rescue. A history of reduced  
484 effective population size and low genome-wide differentiation at neutral loci indicate little  
485 genetic differentiation among populations that may impact a genetic rescue program. However,  
486 previous observations of phenotypic differences paired with loci associated with divergent  
487 selection point towards the importance of adaptive evolution among Torrey pine populations.  
488 These results suggest increased genetic variation via inter-population crosses may be needed in  
489 this species, but admixture should be evaluated first to quantify its fitness impact.

#### 490 **Standing genetic variation**

491 Torrey pine populations exhibited extremely low genetic variation in comparison with  
492 other conifers which often exhibit an expected heterozygosity around 0.3 within populations

493 (e.g., Namroud et al., 2008; Tsumura, Uchiyama, Moriguchi, Ueno, & Ihara-Ujino, 2012).  
494 Average expected heterozygosity ( $H_E$ ) and contemporary effective population size ( $N_e$ ) were  
495 estimated at 0.185 and 2,305 individuals for the island population, and 0.184 and 1,715  
496 individuals for the mainland population, respectively (Table 2; Fig. 5). Although low, this study  
497 is the first to find genetic variability within island and mainland populations of Torrey pine.  
498 While previous genetic analyses using allozymes and chloroplast DNA markers identified fixed  
499 genetic differences between populations, they failed to observe genetic variation within  
500 populations (Ledig & Conkle, 1983; Whittall et al., 2010). Ledig & Conkle, (1983) hypothesized  
501 reduced genetic diversity was attributable to drastically reduced mainland and island populations  
502 in the recent geological past. They suggested that Torrey pine populations declined to fewer than  
503 50 individuals, following which a recovery led to approximately 3,000 to 4,000 reproductively  
504 mature trees (J. Franklin & Santos, 2011; Hall & Brinkman, 2015). Demographic models herein,  
505 however, suggest that effective population size has always been low for Torrey pine (Fig. 5). The  
506 best fit demographic model predicted an ancestral effective population size ( $N_A$ ) of 1,124 and  
507 only limited change following population divergence (Fig. 5). Given these observations, long-  
508 term reduced effective population size has likely exacerbated the consequences of genetic drift  
509 leading to an extreme lack of genetic variation within Torrey pine populations.

510         Despite extremely low genetic variation and small effective population sizes, there was  
511 no evidence for inbreeding ( $F_{IS}$ ) or excessive relatedness ( $\theta$ ) within populations (Table 2). These  
512 findings support high ratios between effective and census population size ( $N_e/N$ ) found in both  
513 Torrey pine populations (island = 0.75, mainland = 0.45) and may, at least partly, explain why  
514 these estimates were higher than ratio averages of 0.1 to 0.2 typically recommended for  
515 conservation management (Frankham et al., 2014). Indeed, it is not uncommon for plants to

516 exhibit  $N_e/N$  ratios greater than 0.2 (Hoban et al., 2020; Waples, Luikart, Faulkner, & Tallmon,  
517 2013). Combined with a lack of within population genetic structure (Fig. 4, Appendix S7), these  
518 results also indicate that neither reproduction among relatives nor unequal reproductive success  
519 has likely contributed to reduced genetic variation within populations. Wind pollination and  
520 zoochorous seed dispersal have likely contributed to homogenizing the gene pool within  
521 populations (M. Johnson, Vander Wall, & Borchert, 2003; Loveless & Hamrick, 1984). Pines  
522 also possess mechanisms that can reduce the probability of self-fertilization, including the  
523 embryo lethal system (Williams, 2009; Williams, Zhou, & Hall, 2001). This self-incompatibility  
524 system selectively induces death in embryos resulting from self-fertilization (Bramlett &  
525 Popham, 1971; Williams, Barnes, & Nyoka, 1999). Consequently, increased dispersal potential  
526 paired with post-zygotic barriers limiting the probability of mating among relatives have likely  
527 reduced within-population genetic structure in Torrey pine.

#### 528 **Neutral genetic differences across populations over time**

529 Demographic modelling using neutral genomic variation supports the maintenance of  
530 some genetic connectivity following population divergence approximately 1.2 MYA (Fig. 5),  
531 estimating the probability of gene exchange at  $8.34 \times 10^{-3}$  per generation. Despite geographic  
532 isolation among populations and reduced potential for inter-population gene flow, contemporary  
533 estimates of  $F_{ST}=0.013$  indicate only subtle genome-wide differentiation between island and  
534 mainland populations. Reduced genetic differentiation is typical of many conifers (Eckert et al.,  
535 2010; Namroud et al., 2008; Tyrmi et al., 2020), as pollen may maintain connectivity over very  
536 long distances (Campbell, McDonald, Flannigan, & Kringayark, 1999; Varis, Pakkanen, Galofré,  
537 & Pulkkinen, 2009; Williams, 2010). Gene flow between populations may also have been  
538 maintained via seed dispersal. Birds represent potential seed dispersers for Torrey pine and may



539 play a prominent role in long-distance seed dispersal (M. Johnson et al., 2003; Pesendorfer,  
540 Sillett, Koenig, & Morrison, 2016; Viana, Gangoso, Bouten, & Figuerola, 2016). Interestingly,  
541 higher estimates of contemporary effective population size ( $N_I$ ,  $N_M$ ) relative to the ancestral  
542 population size ( $N_A$ ) suggest that island and mainland populations have experienced genetic  
543 bottlenecks following one or multiple moderate population expansion events (Fig. 5). Overall,  
544 our findings indicate that despite the increased probability of genetic drift due to genetic  
545 bottlenecks and low population sizes, gene flow maintained between island and mainland  
546 populations may have been sufficient to prevent extensive genomic differentiation at neutral loci  
547 following population isolation. Note, however, that coalescent simulations assume non-  
548 overlapping generations, which may limit their ability to accurately estimate demographic  
549 parameters in long-lived species, including conifers. Consequently, gene flow estimated between  
550 island and mainland populations may have been overestimated or may possibly be an artefact  
551 resulting from an attempt of the demographic model to account for shared ancestral genetic  
552 variation among populations.

### 553 **Evidence for local adaptation**

554 Phenotypic monitoring using common garden experiments or *in situ* morphological  
555 observations for cone, seed, and needle morphology have previously suggested genetically-based  
556 phenotypic divergence among Torrey pine populations (Di Santo et al., *in review*; Haller, 1986;  
557 Hamilton et al., 2017). Thus, subtle genetic differentiation observed among populations likely  
558 reflects the fact that most genome divergence among populations is the result of neutral rather  
559 than adaptive differentiation among populations. To test for the role of selection across loci, we  
560 simulated a null  $F_{ST}$  distribution to compare with our empirical  $F_{ST}$  distribution, which indicated  
561 that a few thousand loci may be under divergent selection (Appendix S5). Genome scans further

562 supported this observation, identifying 2,053 (2.21%) candidate SNPs with accentuated  
563 divergence between island and mainland populations (Fig. 6). Annotation of sequences  
564 containing these outliers SNPs suggested that adaptive evolution in Torrey pine may not only  
565 result from genetic differentiation at the nuclear level, but also at the mitochondrial level (Table  
566 3). This could be consistent with previous observations of the importance of cytoplasmic genetic  
567 differences as a factor contributing to local adaptation in plants (Hamilton & Aitken, 2013;  
568 Leinonen, Remington, Leppälä, & Savolainen, 2013; Leinonen, Remington, & Savolainen, 2011;  
569 for a review see Bock, Andrew, & Rieseberg, 2014). For example, Leinonen et al., (2011) found  
570 using a reciprocal transplant experiment that individuals of *Arabidopsis lyrata* harboring the  
571 local cytoplasmic genome had higher fitness than individuals harboring the non-local  
572 cytoplasmic genome, suggesting that cytoplasmic genetic variation may contribute to local  
573 adaptation.

574 Overall, GO annotation of both nuclear and mitochondrial encoded candidate genes  
575 indicated that genes important for mechanisms such as DNA integration, methylation, gene  
576 silencing, carbohydrate transport and metabolic processes, and defense against pathogens  
577 (bacteria and fungi) were candidates for selection (Table 3). This suggests that between the  
578 island and mainland environments, modification of genetic composition and architecture  
579 following DNA integration, changes in gene expression or protein function following  
580 methylation and gene silencing, and direct or indirect selection against pathogens may have  
581 played an important role in population divergence following isolation. For example, a candidate  
582 gene associated with defense against bacteria and fungi (UniProt accession: B8LLJ5, GO terms:  
583 GO:0042742, GO:0050832, GO:0031640) suggests that phenotypic differentiation may have  
584 evolved in response to pests or pathogens. Indeed, the mainland population of Torrey pine may

585 have faced substantial selection associated with the recent outbreak of the California five-spined  
586 ips (*Ips paraconfusus* Lanier) (J. Franklin & Santos, 2011; Shea & Neustein, 1995), whereas the  
587 island population may not have been exposed to that selective pressure. Noteworthy with these  
588 results is the fact that pine genomes are enormous (Grotkopp, Rejmánek, Sanderson, & Rost,  
589 2004; Stevens et al., 2016), and our sequencing approach (reduced representation sequencing;  
590 see Material and Methods) represents only a fraction of the Torrey pine genome. This suggests,  
591 despite the differences observed, most likely some variation has been overlooked that plays a  
592 critical role in local adaptation for this species.

### 593 **Applying neutral and adaptive evolutionary processes to rare species conservation**

594 While identification of the appropriate effective population size necessary to protect  
595 adaptive evolutionary potential for rare species is still debated, recommendations generally range  
596 between 500 to 5000 individuals (Frankham et al., 2014; I. R. Franklin & Frankham, 1998;  
597 Lynch & Lande, 1998). Torrey pine, critically endangered and endemic to just two native  
598 populations, suffers from extremely low effective population size ( $N_I = 2,305$ ,  $N_M = 1,715$ )  
599 relative to other pines (Menon et al., 2018; Xia et al., 2018). In addition, there is clear evidence  
600 of the impact of adaptive evolutionary processes alongside neutral processes structuring genomic  
601 variation within and among populations. Given historical and contemporary estimates of  
602 effective population size as well as contemporary estimates of expected heterozygosity, our  
603 results indicate that Torrey pine may not retain the genetic variation within populations needed to  
604 adapt to change. Current monitoring within the mainland population suggests that a lack of  
605 recruitment (personal observation), infestation by *Ips* beetles (personal observation), and climate  
606 warming (Diffenbaugh, Swain, & Touma, 2015) may increase extinction risk. Thus, for Torrey  
607 pine, increased  $N_e$  and greater genetic diversity may be required for long-term persistence.

608           As genetic variation is extremely low within populations, one conservation strategy that  
609    may facilitate the maintenance of genetic variation within populations at risk is a genetic rescue  
610    program. A genetic rescue program would facilitate inter-population breeding as a means to  
611    increase heterozygosity, increasing rates of inter-population gene flow. Indeed, demographic  
612    modeling indicates that following population isolation some gene flow has been maintained and  
613    there are low levels of genome-wide differentiation among populations (Nei's  $F_{ST} = 0.013$ ).  
614    However, the combination of observed phenotypic differences and large number of genes that  
615    appear targets of selection suggest island and mainland populations of Torrey pine have  
616    undergone distinct evolutionary trajectories necessary for adaptation following isolation. Thus, a  
617    genetic rescue program should be considered with caution as gene flow between populations  
618    may disrupt local adaptation and further reduce population performance (Goto, Iijima, Ogawa, &  
619    Ohya, 2011; Hufford & Mazer, 2003; Montalvo & Ellstrand, 2001). Despite this word of caution,  
620    preliminary data comparing mainland, island, and F1 individuals from a common garden  
621    experiment planted outside the species natural distribution indicate that F1s exhibit increased  
622    fitness relative to mainland and island populations (Hamilton et al., 2017). Consequently, future  
623    monitoring is needed to empirically quantify fitness consequences of advanced-generation  
624    admixture (F2, Backcross-Island (BC-I), Backcross-Mainland (BC-M)) following early-  
625    generation heterosis.

626           Given the challenge to conserve and manage rare species in a rapidly changing  
627    environment, the use of genomic data to model evolutionary history, assess demographic change,  
628    and tease apart the contributions of neutral and adaptive processes will be critical. For Torrey  
629    pine, the fact that there is low genome-wide differentiation among populations, a consistent  
630    history of low effective population size, and indications that some gene flow is maintained

631 among populations may suggest that one population (island or mainland) could be targeted to  
632 effectively preserve neutral genetic variation. However, the combination of outlier loci and  
633 previously observed phenotypic differences suggest if the goal is to preserve adaptive genetic  
634 variation, a strategy that favors conservation efforts across both mainland and island populations  
635 will be needed. If conservation strategies such as genetic rescue are considered, assessment of  
636 multiple admixed generations within a common environment will provide the necessary  
637 empirical test to evaluate the consequences of enhancing genetic exchange among populations.

638

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655

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970 **Tables**  
 971 **Table 1.** Demographic parameters with their prior distributions and occurrence in each of the six  
 972 models simulated.

Present in models	Parameter	Symbol	Prior distribution	Unit
RPM1, RPM2	Ancestral effective population size	$N_A$ <sup>†</sup>	logU(100:100,000)	ind.
All	Island effective population size	$N_I$ <sup>‡</sup>	U(100:6,000)	ind.
All	Mainland effective population size	$N_M$ <sup>‡</sup>	U(100:6,000)	ind.
CM1, CM2, CM3, CM4	Initial effective population size following colonization	$N_C$ <sup>§</sup>	logU(2:300)	ind.
All	Time of population divergence	$T_{Div}$ <sup>¶</sup>	logU(400:50,000)	gen.
RPM2, CM2, CM4	Migration from island to mainland and vice-versa.	$m_{IM}, m_{MI}$ <sup>§</sup>	U(0.001:0.01)	-
All	Minor allele frequency	maf <sup>i</sup>	0.01	-

973 <sup>†</sup> A wide prior range for Torrey pine ancestral effective population size was used to enable the simulation  
 974 of complex demographic histories, including population size expansion, population size reduction or a  
 975 combination of both.

976 <sup>‡</sup> Census size of reproductively mature trees is of approximately 3,000 to 4,000 individuals in island and  
 977 mainland population respectively (J. Franklin & Santos, 2011; Hall & Brinkman, 2015). We selected a  
 978 wide prior range around those estimates to account for the fact that effective and not census population  
 979 sizes were simulated.

980 <sup>§</sup> Number of potential effective migrants were chosen based on seed capacity of a Torrey pine cone,  
 981 which can hold up to several hundred seeds (unpublished data).

982 <sup>¶</sup> Time since populations' isolation was estimated between 4,300 and 430,000 years ago (Ledig & Conkle,  
 983 1983). Generation time for Torrey pine being approximately 10 years, this translates into a time since  
 984 divergence between 430 and 43,000 generations.

985 <sup>§</sup> Low genome-wide differentiation observed between populations in this study ( $F_{ST} = 0.013$ ) suggest that  
 986 gene flow may have occurred between populations of Torrey pine. We believe a probability of migration  
 987 between 0.001 and 0.01 should appropriately reflect low to moderate gene exchange expected between  
 988 geographically distant populations (approximately 280 km) of this wind-pollinated, long-lived, and  
 989 primarily outcrossing pine species ( $F_{IS} = -0.127$  and  $-0.124$  for island and mainland population  
 990 respectively).

991 <sup>i</sup> Minor allele frequency of 0.01 was chosen to reflect empirical genetic data, filtered for variants with  
 992 frequencies of or above 0.01 (see Material and Methods).

993 **Table 2.** Average genetic summary statistics across loci, including expected heterozygosity ( $H_E$ ),  
994 inbreeding coefficient ( $F_{IS}$ ), and coancestry (relatedness) coefficient ( $\theta$ ) for island (SRI) and  
995 mainland (TPSR) Torrey pine populations. Ninety-five percent confidence interval (95% CI)  
996 around mean estimates were obtained by bootstrapping.

<b>Population</b>	<b><math>H_E</math> (95% CI)</b>	<b><math>F_{IS}</math> (95% CI)</b>	<b><math>\theta</math> (95% CI)</b>
Island	0.185 (0.184, 0.186)	-0.127 (-0.128, -0.126)	0.023 (0.023, 0.024)
Mainland	0.184 (0.183, 0.184)	-0.124 (-0.125, -0.124)	0.022 (0.022, 0.023)

997

998 **Table 3.** Functional categorization of 80 putatively adaptive genes between Torrey pine  
 999 populations. Listed are the ten most frequent GO terms within each of the three GO classes  
 1000 (Biological Process, Molecular Function, Cellular Component).

GO class	GO term	Frequency (%)
Biological Process	DNA integration	6.2
	Methylation	2.5
	Actin filament polymerization	1.2
	Arp2/3 complex-mediated actin nucleation	1.2
	Carbohydrate metabolic process	1.2
	Carbohydrate transport	1.2
	Cell differentiation	1.2
	Defense response to bacterium	1.2
	Defense response to fungus	1.2
	Gene silencing by RNA	1.2
Molecular Function	RNA-directed DNA polymerase activity	11.2
	Nucleic acid binding	7.5
	ATP binding	3.8
	ADP binding	2.5
	Magnesium ion binding	2.5
	O-methyltransferase activity	2.5
	Oxidoreductase activity	2.5
	Protein kinase activity	2.5
	4-lactone oxidase activity	1.2
	Actin binding	1.2
Cellular Component	Mitochondrion	38.8
	Integral component of membrane	7.5
	Arp2/3 protein complex	1.2
	Cell wall	1.2
	Chloroplast	1.2
	Chloroplast thylakoid membrane	1.2
	Cytoplasm	1.2
	Extracellular region	1.2
	Membrane	1.2
	Nucleus	1.2

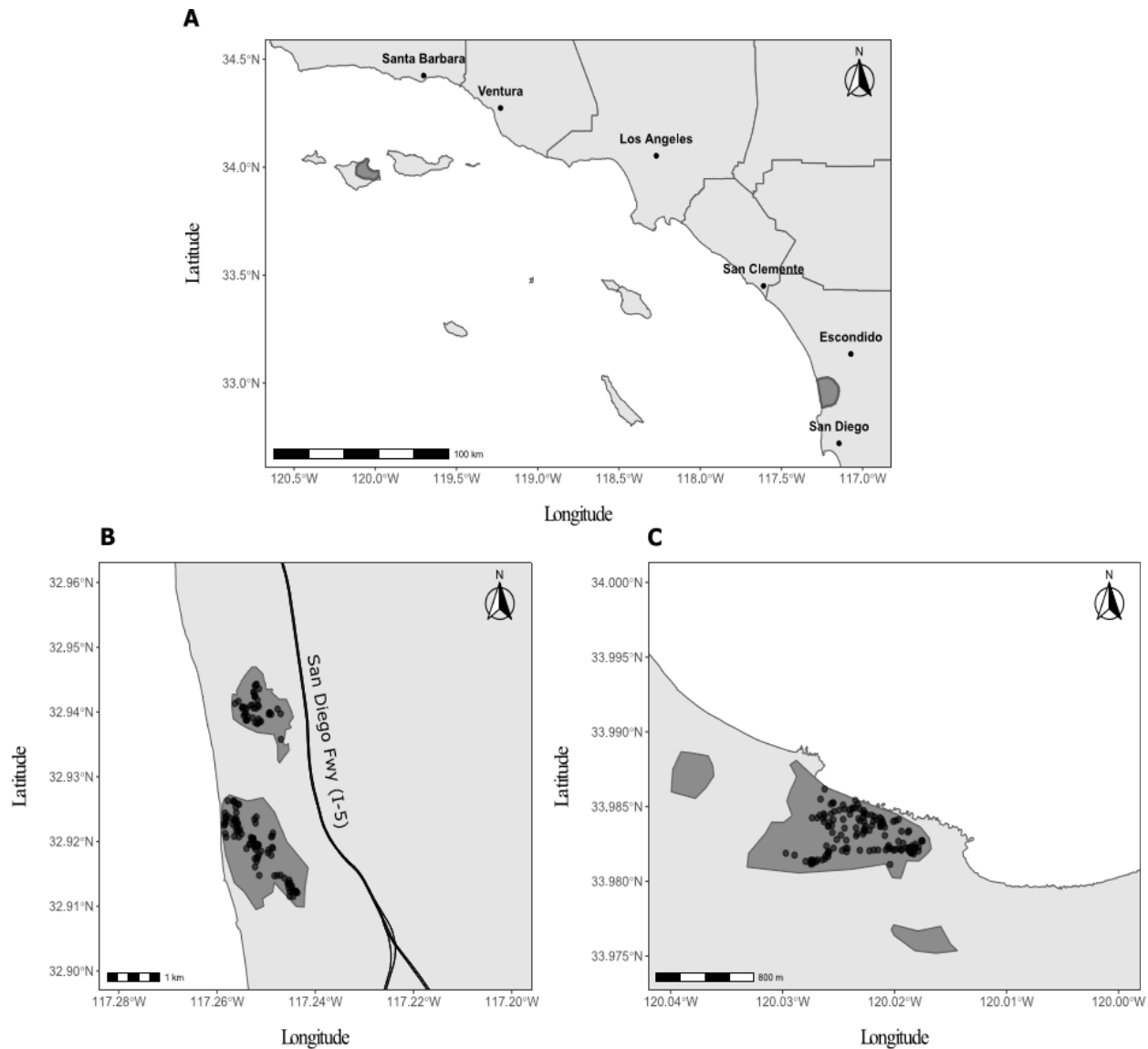
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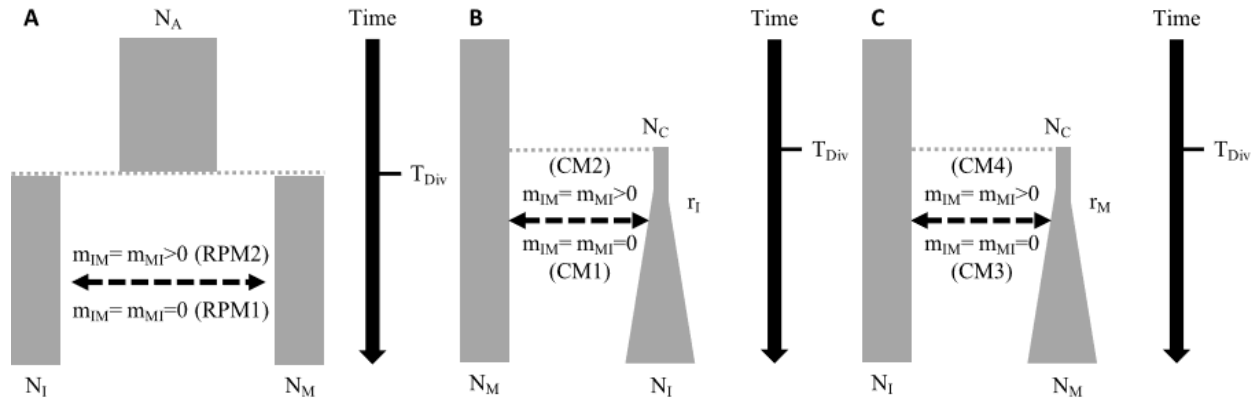
## Figures



1004

1005 **Figure 1.** (A) Distribution of the last two remnant populations of Torrey pine (dark grey  
1006 shades). Top left: Santa Rosa Island, Channel Islands, CA. Bottom right: Torrey Pine State  
1007 Reserve, La Jolla, CA. (B-C) Population-specific distribution of Torrey pine (dark grey shades)  
1008 and trees sampled for needle tissue (circles) at the Torrey Pine State Reserve (TPSR, B) and on  
1009 Santa Rosa Island (SRI, C).

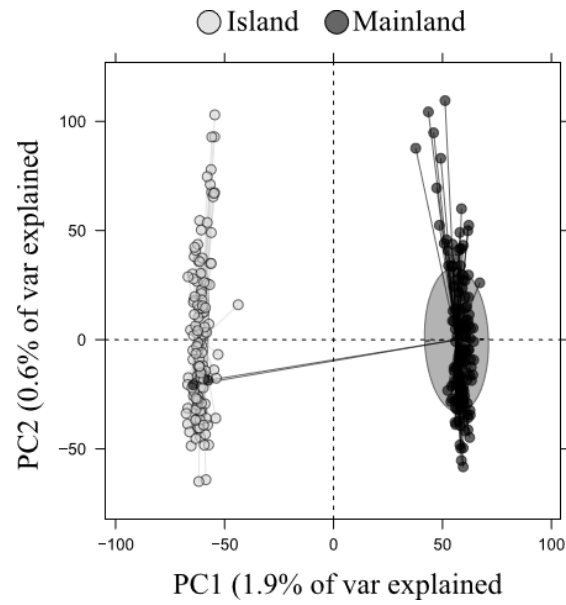
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1011

1012 **Figure 2.** Schematics of demographic scenarios simulated. Rectangles represent current or  
 1013 ancestral populations, dashed arrows represent migration between population, and solid arrows  
 1014 represent time. (A) Scenarios of isolation without (RPM1) or with (RPM2) migration between  
 1015 populations. (B) Island colonization scenarios without (CM1) or with (CM2) subsequent  
 1016 migration between populations. (C) Mainland colonization scenarios without (CM3) or with  
 1017 (CM4) subsequent migration between populations.  $N_A$ , ancestral effective population size;  $N_I$ ,  
 1018 island effective population size;  $N_M$ , mainland effective population size;  $N_C$  initial effective  
 1019 population size following migration (number of migrants);  $m_{IM}$ , migration probability from  
 1020 island to mainland;  $m_{MI}$ , migration probability from mainland to island;  $T_{Div}$ , time of population  
 1021 divergence;  $r_I$ , island (exponential) growth rate;  $r_M$ , mainland (exponential) growth rate

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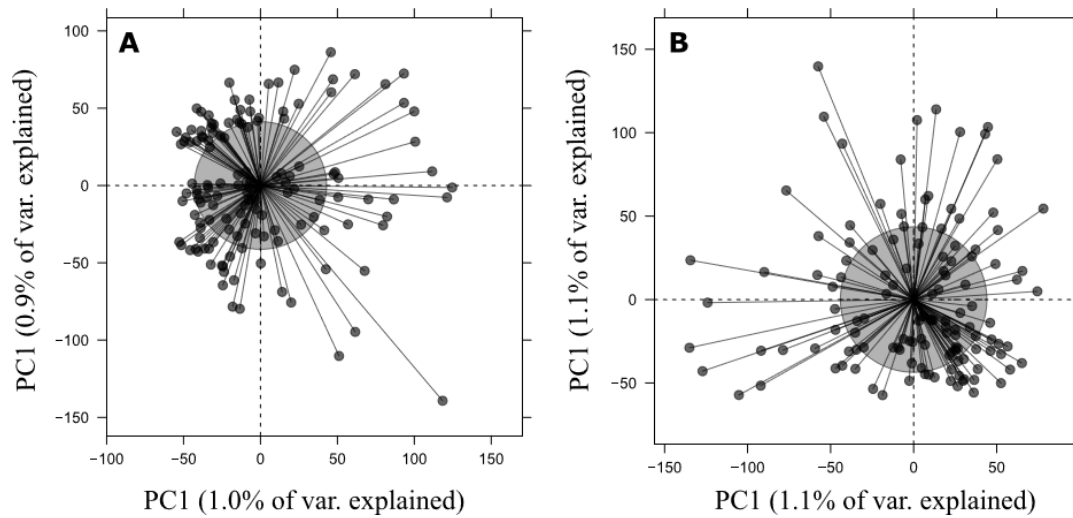
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1024 **Figure 3.** Principal component analysis using 93,085 SNPs for 270 Torrey pine individuals,

1025 including individuals from both mainland (black) and island (grey) populations. Variation

1026 explained by the first two principal components is provided in parentheses.

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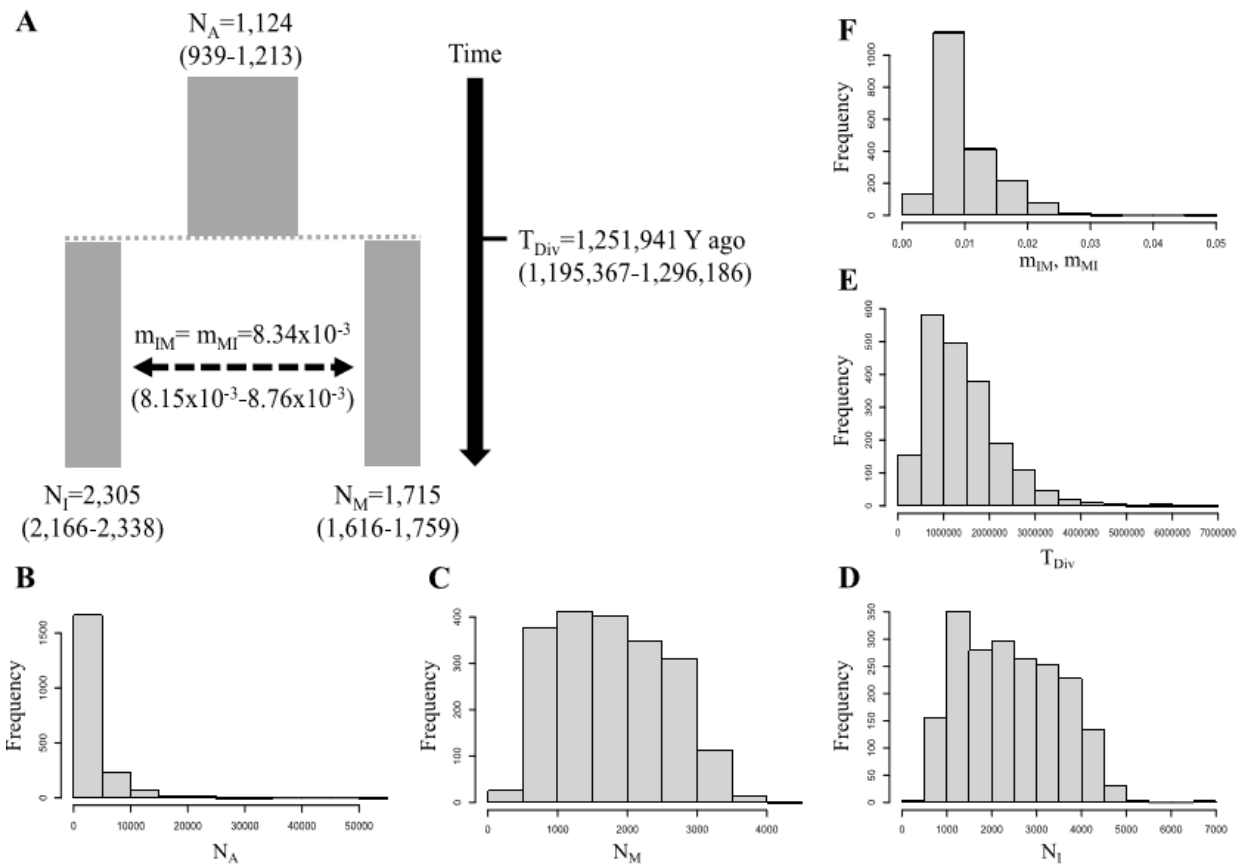
1029 **Figure 4.** Population-specific principal component analysis based on 93,085 SNP variants. (A)

1030 Mainland population (TPSR, n = 138). Note that two individuals were removed from analysis to

1031 visualize within-population structure on a finer scale (see Material and Methods). (B) Island

1032 population (SRI, n = 130).

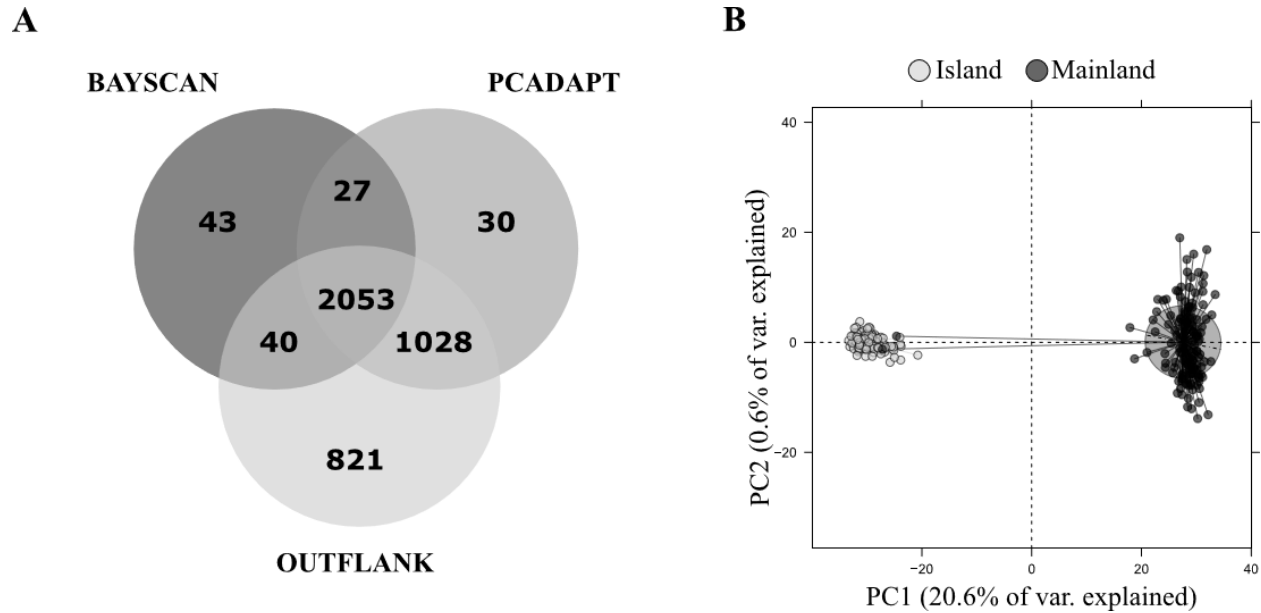
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1034

1035 **Figure 5.** (A) Demographic parameters (weighted medians) and 95% confidence intervals (in  
 1036 parentheses) estimated using RPM2. Each rectangle represents a population, either contemporary  
 1037 or ancestral. The solid arrow represents time, while the dashed arrow indicates gene flow  
 1038 between populations. (B-F) Posterior distribution of each demographic parameter inferred using  
 1039 a tolerance rate of 0.01.  $N_A$ , ancestral effective population size;  $N_I$ , island effective population  
 1040 size;  $N_M$ , mainland effective population size;  $m_{IM}$ , migration probability from island to mainland;  
 1041  $m_{MI}$ , migration probability from mainland to island;  $T_{Div}$ , time of population divergence.

1042



1043

1044 **Figure 6.** (A)Venn diagram showing the number of putatively unique and shared adaptive SNPs  
1045 detected by BAYESCAN, PCADAPT, and OUTFLANK. (B) Principal component analysis  
1046 based on all 2,053 shared candidate SNPs for Torrey pine individuals, including individuals from  
1047 both mainland (black, n = 140) and island (grey, n = 130) populations. Variation explained by  
1048 the first two principal components is provided in parentheses.