Machine learning models based on molecular descriptors to predict human and environmental toxicological factors in continental freshwater

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Highlights:

- Characterization factors (for human health and ecotoxicological impacts) were predicted using molecular descriptors.
- Several linear or non-linear machine learning methods were compared.
- The non-linear methods tend to outperform the linear ones using a train and test procedure. Cluster-then-predict approaches often show the best performances, highlighting their usefulness.
- This methodology was then used to derive characterization factors that were missing for more than a hundred chemicals in USEtox®.

Abstract: It is a real challenge for life cycle assessment practitioners to identify all relevant substances contributing to the ecotoxicity. Once this identification has been made, the lack

1 of corresponding ecotoxicity factors can make the results partial and difficult to interpret. So,
 2 it is a real and important challenge to provide ecotoxicity factors for a wide range of
 3 compounds. Nevertheless, obtaining such factors using experiments is tedious, time-
 4 consuming, and made at a high cost. A modeling method that could predict these factors
 5 from easy-to-obtain information on each chemical would be of great value. Here, we present
 6 such a method, based on machine learning algorithms, that used molecular descriptors to
 7 predict two specific endpoints in continental freshwater for ecotoxicological and human
 8 impacts. The different tested machine learning algorithms show good performances on a
 9 learning database and the non-linear methods tend to outperform the linear ones. The
 10 cluster-then-predict approaches usually show the best performances which suggests that
 11 these predicted models must be derived for somewhat similar compounds. Finally,
 12 predictions were derived from the validated model for compounds with missing
 13 toxicity/ecotoxicity factors.

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15 Graphical abstract:

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1 **Keywords:** machine learning, Life Cycle Assessment, characterisation factors, toxicity,
2 ecotoxicity, continental freshwater.

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4 1. Introduction

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6 Recent legislations such as the Registration, Evaluation, Authorization and restriction of
7 Chemicals (REACH) regulation in the EU requires that manufacturers of substances and
8 formulators register to provide toxicological and ecotoxicological data for substances with
9 volume higher than one metric ton per year. Furthermore, the U.S. Environmental Protection
10 Agency (EPA) has more than 85,000 chemicals listed under the Toxic Substances Control
11 Act (Hinds and Weller, 2016). Therefore, robust toxicological and ecotoxicological data are
12 quickly needed to make informed decisions on how to regulate new chemicals. These data
13 must also be coupled with environmental exposures and sources data, to better understand
14 the impact on the environment.

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16 To address the cause-effect relationships between the flow of molecules emitted by human
17 activities and the consequences for ecosystems and humans, Life Cycle Assessment (LCA)
18 offers a structured, operational, and standardized (Finkbeiner et al., 2006) methodological
19 framework. Two main steps are at the core of this approach:

- 20 • Quantification of the masses of substances emitted into the environment through the
21 Life Cycle Inventory (LCI). While it is possible to rely on databases that facilitate this
22 inventory work for the background of the system under study, this task must
23 nevertheless be carried out on a case-by-case basis to represent all the specificities
24 of the foreground elements. This is the task of the LCA practitioner.
- 25 • Calculation of the impacts on ecosystems and human health of these emitted
26 masses. Due to the complexity of environmental mechanisms, it is not possible to
27 (re)model impact pathways on a case-by-case basis. Therefore, LCA uses
28 characterization factors (CF) to assess the potential impacts of a compound.

1 Concretely, if two compounds are emitted with the same mass, the one with the
2 higher CFs will have the higher impact. Then, CFs are multiplied by the emitted
3 masses of each compound to determine the impacts. CFs are not recalculated for
4 each study but provided within a Life Cycle Impact Assessment (LCIA) method.

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6 For a given impact, the LCIA method designer refers to the knowledge of the scientific
7 community to model the mechanisms involved. For human toxicity and freshwater
8 ecotoxicity, USEtox® (Rosenbaum et al., 2008), was developed by life cycle initiative under
9 the United Nations Environmental Programme (UNEP) and the Society for Environmental
10 Toxicology and Chemistry (SETAC) (Henderson et al. 2011) to produce a transparent and
11 consensus characterization model. USEtox® is also used for the European Product
12 Environmental Footprint (PEF) (Saouter et al., 2020). This model gathers in one single
13 characterization factor the chemical fate, the exposure, and the effect for each of the several
14 thousands of organic and inorganic compounds. Then, the USEtox® model results can be
15 extended to determine endpoint effects expressed as total (i.e. cancer and non-cancer)
16 disability-adjusted life years (DALY) for human health impacts and potentially disappeared
17 fraction of species (PDF) for ecotoxicological impacts. The PDF represents an increase in
18 the fraction of species potentially disappearing as a consequence of emission in a
19 compartment while the DALY represents an increase in adversely affected life years. These
20 endpoints are now consensual at an international level (Verones et al., 2017).

21 If the structure of the USEtox® multimedia model is always the same, to determine the CF of
22 a molecule, numerous physicochemical parameters (such as solubility, hydrophobicity,
23 degradability) and detailed toxicological and ecotoxicological data must be provided. For
24 example, EC50 values (i.e. the effective concentration required to have a 50% effect) for at
25 least three species from three different trophic levels are required for the ecotoxocological
26 effect factor.

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1 Over the past few decades, thousands of tests (in laboratory and field) have been carried out
2 to evaluate the potential hazard effects of chemicals (He et al., 2017). Usually, toxicity
3 testing has relied on *in vivo* animal models, which is extremely costly and time-consuming
4 (Xia et al., 2008). In recent years, under societal pressures, there has been a significant
5 paradigm shift in toxicity testing of chemicals from traditional *in vivo* tests to less expensive
6 and higher throughput *in vitro* methods (National Research Council, 2007). However, it is still
7 extremely difficult to test existing chemicals due to their large and ever increasing number,
8 which leaves their impacts largely unknown. For example, in a recent study, Aemig et al.
9 (2021) studied the potential impacts on Human health and aquatic environment of the
10 release of 286 organic and inorganic micropollutants at the scale of France. One of their
11 conclusion was that, due to a lack of characterization factors, these impacts could be
12 assessed only for 1/3 of these molecules. That is why more computational models are
13 needed to complement experimental approaches to decrease the experimental cost and
14 determine the prioritization for those chemicals which may need further *in vivo* studies. Such
15 models already exist, like QSAR models that are mostly linear models based on the
16 chemical structure of compounds (Danish QSAR database (DTU, 2015), ECOSAR (Mayo-
17 Bean et al., 2011), VEGA (Benfenati et al., 2013)) and are used to predict ecotoxicological
18 data (LC50) needed for REACH for example. Recently, machine learning algorithms have
19 been used to predict some midpoints based on molecular descriptors and environmental
20 parameters (Marvuglia et al., 2014 and 2015; Song et al., 2017; Lysenko et al 2018) and a
21 first review on this subject could be found in Wu and Wang (2018). After these first works,
22 predictions of hazardous concentration 50% (HC50) based on 14 physicochemical
23 characteristics (Hou et al., 2020a) or on 691 more various variables (Hou et al., 2020b) were
24 carried out. Nevertheless, their input variables need some experiments and could be difficult
25 to collect. This problem was tackled by Song et al. (2021) who predicted Lethal
26 Concentration 50 (LC50) based on 2000 easy-to-obtain molecular descriptors. In the case of
27 USEtox®, despite its wide use in LCA, it only offers characterization factors for

approximately 3000 chemicals and even for this limited number of compounds, 19% of ecotoxicity CFs and 67% of human toxicity CFs are missing.

The objective of this article is thus to propose a new way of calculating CFs using machine learning approaches to solve the problem of nonlinearity that could affect a linear QSAR method. This makes it possible, when the CFs are not determined due to lack of time or lack of data, to propose values based solely on easily identifiable molecular descriptors. Here, the main differences with the above-cited methods are twofold: first, our input variables are only molecular descriptors that could be easily collected for any newly available compounds; second, our output variables are directly the CFs that are closer to the endpoints (DALY and PDF) than the HC50 or the LC50 (i.e. the acute aquatic toxicity experimental threshold). These two specific endpoints will be studied in the present paper through the emission of compounds in continental freshwater and will be named CF_{ET} for ecotoxicological impacts and CF_{HT} for human ones. To address this aim, we will test different methods (linear and non-linear) and assess their performances, to build a robust model that could predict CFs that are currently lacking.

2. Materials & Methods

2.1. USEtox® database

The last version of the USEtox® database was downloaded, namely the corrective release 2.12 (USEtox®, 2020). The whole USEtox® 2.12 database contains 3076 compounds. The CFs were computed using the default landscape.

2.2. TyPol database

We recently developed TyPol (Typology of Pollutants), a classification method based on statistical analyses combining several environmental parameters (i.e., sorption coefficient,

1 degradation half-life, Henry constant) and an ecotoxicological parameter (bioconcentration
2 factor BCF), and structural molecular descriptors (i.e., number of atoms in the molecule,
3 molecular surface, dipole moment, energy of orbitals). Molecular descriptors are calculated
4 using an *in silico* approach (combining Austin Model1 and Dragon software). In the present
5 paper, we only extracted and used the molecular descriptors from the TyPol database, as
6 this information could be easily collected for any new compound. The 40 descriptors
7 included in the TyPol database have been selected based on a literature review on QSAR
8 equations used to predict the main environmental processes as degradation, sorption,
9 volatilization. These 40 descriptors were the ones most frequently used in the equations,
10 meaning describing the best the behaviour of organic compounds in the environment. By
11 consequence, even if no environmental parameters were directly incorporated as input in our
12 model, some information that is directly linked to them were included in the 40 molecular
13 descriptors. These descriptors are constitutional, geometric, topological, and quantum-
14 chemical descriptors (see Table 1); 35 described the 2D-structure of the compound while the
15 other five are linked to its 3D-structure. An important advantage of the unique use of
16 molecular descriptors is that they are easily and quickly computable for not yet synthesized
17 compounds. For more details, we refer the interested reader to Servien et al. (2014) and to
18 Mamy et al. (2015) where the choice of the 40 molecular descriptors is described in details.
19 Now, TyPol gathers 526 compounds, including pesticides, persistent chemicals,
20 pharmaceuticals and their transformation products (Benoit et al. 2017, Traoré et al. 2018).

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22 **Table 1** – List of the 40 molecular descriptors in TyPol

Category	Molecular descriptors		
Constitutional	Number of atoms	Number of non-H atoms	Number of hydrogen atoms
	Number of chlorine atoms	Number of carbon atoms	Number of nitrogen atoms

	Number of oxygen atoms	Number of phosphorus atoms	Number of sulfur atoms
	Number of fluorine atoms	Number of circuits	Number of halogen atoms
	Number of bonds	Number of non-H bonds	Number of double bonds
	Number of triple bonds	Number of multiple bonds	Number of rotatable bonds
	Number of aromatic bonds	Sum of conventional bond order	Number of rings
	Molecular weight		
Geometric	Connolly molecular surface area		
Topological	Connectivity index of order 0	Connectivity index of order 1	Connectivity index of order 2
	Connectivity index of order 3	Connectivity index of order 4	Connectivity index of order 5
	Valence connectivity index of order 0	Valence connectivity index of order 1	Valence connectivity index of order 2
	Valence connectivity index of order 3	Valence connectivity index of order 4	Valence connectivity index of order 5
Quantum-chemical	Polarizability	Electric dipole moment	HOMO energy
	LUMO energy	Total energy	

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2.3. Machine learning methods

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2 To predict the CFs using the molecular descriptors we used three modeling methods
3 combined. The first method is a linear well-known prediction method namely the Partial
4 Least Squares (PLS) (Wold, 1985). It finds the multidimensional directions in the observable
5 variable (molecular descriptor) space that explains the maximum multidimensional variance
6 direction in the predicted variable (CF) space. That provides a linear regression model based
7 on the observable variables to predict the predicted variable. We also chose to compare two
8 machine learning methods adapted to non-linear problems: the random forest (Breiman
9 2001) and the support vector machines (SVM) (Drucker et al. 1996). Random forests are a
10 machine learning method, for classification or, in our case, regression, that operate by
11 constructing a multitude of decision trees that uses a random subset of the training data and
12 limits the number of variables used at each split and outputting the mean prediction
13 (regression) of the individual trees. SVM constructs a hyperplane or set of hyperplanes in a
14 high- or infinite-dimensional space in which the problem is linearly separable.

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16 These choices allowed us to compare several ideas. The PLS is a simple linear method that
17 will not exhibit good performances if the underlying relationship is not linear. The SVM and
18 RF methods are well-known non-linear machine learning algorithms that used to show good
19 results in this kind of problem (Hou et al., 2020a).

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21 All the models were computed in the freeware R (R core team, 2019). The PLS has been
22 computed using the package mixOmics (Rohart et al., 2017), the random forests using the
23 package randomForest (Liaw et al., 2002), and the SVM using the package e1071 (Meyer et
24 al., 2019). These 3 modeling methods have some parameters that needed to be fixed: the
25 number of latent components for the PLS (fixed using the tune.pls function), the number of
26 variables randomly sampled as candidates at each split for the random forests (selected
27 using the tune.randomForest function) and, for the SVM, the gamma parameter of the radial
28 kernel and the cost of constraints violation (using the tune.svm function). All these different

tune functions are based on cross-validation (i.e. a training/test procedure to find the best value for the parameters) using default function values.

2.4. Clustering-based model

A recent popular way to make predictions is to use a cluster-then-predict approach. That is, clustering is used for pre-classification which is to arrange a given collection of input patterns into natural meaningful clusters. Then, the clustering results are used to construct a predictor in each cluster. The main idea of the cluster-then-predict approach is that if the clustering performs well, the prediction will be easier by modeling only similar compounds. If a new compound with no CF_{ET} and/or CF_{HT} is investigated, the clustering can easily be applied to it before the prediction model itself. The cluster-then-predict approach has already been applied with success in various domains such as sentiment prediction (Sony et al., 2015), finance (Tsai et al., 2014), chemometrics (Minh Maï Le et al., 2018). So we decided to use the clustering given by the TyPol application (more details in Servien et al., 2014). Note that the TyPol clustering has already been shown relevant on various occasion: in combination with mass spectrometry to categorize tebuconazole products in soil (Storck et al., 2016), to explore the potential environmental behaviour of putative chlordecone transformation products (Benoit et al., 2017) or to classify pesticides with similar environmental behaviors (Traore et al., 2018; Mamy et al., 2021). So, the clustering procedure of TyPol was applied on the whole database of 526 compounds using the 40 molecular descriptors. This approach provided us a global clustering based on all the available information contained in the TyPol database. It is based on PLS, hierarchical clustering and an optimal choice of the number of clusters and is detailed in Servien et al. (2014). The obtained clustering is given in Supplementary Figure S1 and relies on 5 different clusters. The Supplementary Figure S2 represents this clustering restricted to the common molecules between TyPol and USEtox®. We could see that, as the cluster 5 is only constituted of one compound, the cluster-then-predict-models cannot be applied.

Based on this clustering, we then defined three other competing methods. For these methods, a different model (with different parameters) was derived for the compounds in each cluster. Consequently, six different models were calibrated and tested for each CF prediction: global PLS, global SVM, global random forest, cluster-then-PLS, cluster-then-SVM and cluster-then-Random Forest.

2.5. Comparison procedure of the models

To assess the performances of the different models we used the following procedure:

1. Split each cluster (the whole dataset if the model is global, only the data lying in the dedicated cluster if that is a cluster-then-predict model) between a training set (85% of the dataset) and a test set (15%). The test set is not used for any step of the procedure (such as the imputation of the missing data, the calibration of the parameters ...).
2. Imputation of the NA (Not Available, *i.e.* missing) values (less than 1%) in the descriptor matrix using the NIPALS algorithm (Wold, 1985).
3. Tune the parameters and train the specific models by performing cross-validation on the training set. We have 3 global models to train (PLS, random forest, and SVM) and the cluster-then-test models (PLS, random forest and SVM for each cluster).
4. Test the different models on the test set. Compute the absolute error.
5. Back to step 1.

The whole algorithm was repeated 200 times. All the performances were compared in terms of absolute error. The absolute error is the absolute difference between the prediction and the true value. It has been shown to be the most natural and unambiguous measure of error (Willmott et Matsuura, 2005) and is chosen to be easily comparable to the assumed error on the experimental CFs (2-3 logs, see Rosenbaum, 2008). For each cluster, we chose the model with the lowest median absolute error.

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2.6 Predictions

Then, the best model was calibrated and computed on the whole cluster. Finally, it was applied to the compounds, according to their clusters, with a CF_{ET} (or a CF_{HT}) equals to NA to provide a prediction. For the compounds in cluster 5, this best model cannot be a cluster-then-predict one and, by consequence, is a global one. To assess the robustness of our prediction we derived a 95% prediction interval for each prediction. The type of model and its corresponding parameters were fixed during this process, according to the best model of the cluster. For example, if the best model of cluster 1 was the random forest approach, random forest models are used with the parameters optimized during the previous step. Then, we performed a leave-one-out bootstrap on the dataset that was used to compute the model (the whole dataset if the model is global, only the data lying in the dedicated cluster if that is a cluster-then-predict model) and a new model was computed on this leave-one-out sample. A prediction was carried for each leave-one-out model (*i.e.* $n-1$ models if n is the number of compounds of the, eventually global, cluster) and the 2.5% and 97.5% quantile of these predictions were computed and considered as the prediction interval (Hou et al., 2020a). The whole modeling process is summarized in the following Figure 1.

Figure 1 Schematic representation of the modeling procedure adopted in the paper.

The five molecular descriptors contributing the most to the prediction were then derived for each chosen model to assess the differences between models and to interpret their relevance. For a random forest model, these descriptors are calculated using variable permutations (Breiman, 2001), for the SVM they are the descriptors with the higher coefficients in absolute value.

3. Results

3.1. Descriptive analysis of the intersection of the TyPol and the USEtox® databases

As the objective of this proof-of-concept study was to predict USEtox® CF_{ET} and CF_{HT} using the molecular descriptors contained in TyPol, we could only use the compounds that are present in both databases. This resulted in 274 compounds that are detailed in Table S1 in supplementary material and the range of their CF_{ET} and CF_{HT} values are summarized in the

- 1 boxplots in Figures 2 and S3. Note that for the 274 common compounds there are 15 NA
- 2 values for the CF_{ET} and 102 for the CF_{HT} .

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- 4 **Figure 2** Boxplots of the CF_{HT} for the USEtox® database and the common molecules
- 5 between the USEtox® and the TyPol databases. This CF_{HT} is equal to $\log_{10}((DALY+\epsilon).kg^{-1})$.
- 6 The ϵ is needed as some values of the DALY are exactly equal to zero. ϵ has been chosen
- 7 equal to $1e-10$ to be below the minimum of the USEtox® database ($5e-9$).
- 8
- 9 We could see on these two figures that the common compounds present higher CF_{ET} and
- 10 CF_{HT} values than the one of the complete USEtox® database: it focuses on the more

1 dangerous compounds as their boxplots are above the USEtox® counterparts and they
2 cover the whole order of magnitude of the CFs of the USEtox® database.

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4 **3.2. Clustering of the compounds**

5 The global TyPol clustering of Supplementary Figure S1 with only the common compounds
6 is plotted in Supplementary Figure S2 and the boxplots of each molecular descriptor per
7 cluster are given in Supplementary Figure S4 with different indicators in Table S2. We could
8 see that they are clustered in 5 groups with different sizes (respectively 33 compounds in the
9 first black cluster, 122 compounds in the second red cluster, 91 compounds in the third
10 green cluster, 27 compounds in the fourth blue cluster, and one compound in the fifth brown
11 cluster). Cluster 1 grouped compounds with a high number of aromatic bonds, double bonds,
12 rotatable bonds, and multiple bonds. Cluster 2 is an intermediate one between clusters 1
13 and 3, with less extreme values. Cluster 3 is made of compounds with the lowest molecular
14 mass. Cluster 4 gathered compounds presenting a high number of halogens, rings, and
15 circuits. The unique compound in the fifth cluster is erythromycin (highest molecular mass
16 and number of H and C, lowest number of rings) and, obviously, no cluster-then-predict
17 model could be built for this cluster

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19 As a first analysis of the clustering given by TyPol, we could see in Figure 3 below the
20 boxplots of the CF_{ET} and CF_{HT} within the 5 clusters.

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Figure 3- Boxplot by cluster for the CF_{ET} and CF_{HT} values. Note that the unique compound of Cluster 5 has no CF_{HT} value. The size of the clusters and the numbers of NA are gathered in the legend.

The predictions will be made difficult for the CF_{ET} of cluster 1 as it covers a wide range whereas it includes a relatively small number of compounds. On the contrary, cluster 3 covers a small range with no extreme values and includes a high number of compounds, for this cluster the cluster-then-predict approach could produce interesting results.

3.3. Performances of the machine learning methods

3.3.1. For the prediction of the CF_{ET}

The methodology described in the previous section was applied to our dataset and gave the results gathered in Figure 4 for each cluster.

Figure 4 - Performances of the different methods (RF: random forest, PLS: partial least squares, SVM: support vector machines) in terms of the log of the absolute error of the CF_{ET} with respect to the different clusters for 200 repetitions. In each cluster, the models are coloured from dark blue (best) to clear blue (worst) according to their median of the absolute error. The red dotted line represents an absolute error of 1 log that is considered as acceptable (Rosebaum et al. 2008 ; Douziech et al., 2019).

The performances were not similar from one cluster to another. For example, performances of all methods for cluster 1 were very poor (median absolute error above 1) whereas performances for cluster 4 seemed good despite its smallest size (median absolute error

1 around 0.6). Therefore, a future prediction of an unknown compound which lies in cluster 1
2 will be less reliable than in other clusters. Note that we could not test this in the next section
3 as no NA value is present in this cluster 1.

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5 The cluster-then-predict methods seemed more appropriate in each cluster. The cluster-
6 then-RF approach had the best performances (with a global median absolute error equals to
7 0.64 and the best performances on clusters 2 and 3), even if there was not a big difference
8 between the different methods. The cluster-then-SVM was also the best method for the two
9 clusters 1 and 4. The linear methods (PLS and cluster-then-PLS) had higher absolute errors
10 but were competitive. The individual predictions of the best method in each cluster are
11 reported in Figure S5.

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15 **3.3.2. Prediction for the CF_{HT}**

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17 Let us recall that we have more NA values for the CF_{HT} (102) than for the CF_{ET} (15). The
18 performances of the methods are illustrated in the following figure.

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Figure 5- Performances of the different methods (RF: random forest, PLS: partial least squares, SVM: support vector machines) in terms of the log of the absolute error of the CF_{HT} with respect to the different clusters for 200 repetitions. In each cluster, the models are coloured from dark blue (best) to clear blue (worst) according to their median of the absolute error. The red dotted line represents an absolute error of 1 log that is considered as acceptable (Rosebaum et al. 2008 ; Douziech et al., 2019).

We observed that, despite its small size (11 compounds), the CF_{HT} of the first cluster were well predicted (with the best performance for the cluster-then-RF approach). It could be explained by the small range of the CF_{HT} values of this cluster, as illustrated on the boxplot in Figure 3. The performances of all the methods were comparable on clusters 2 and 3 where the best method was the SVM. Cluster 4 seemed to be the most difficult to predict: all the methods had their worst results on this cluster and, if the SVM had an acceptable median absolute error of 0.82, all the medians of the other methods were above 1.3. Global

performances of the different methods were given in Supplementary Figure S6. Note that, as for CF_{ET} , the linear methods based on PLS were outperformed by the other ones.

3.4. Best model predictions

3.4.1. Best models for the CF_{ET}

Using our methodology, we could exhibit a median absolute error of 0.62 log for the prediction of the CF_{ET} on the whole dataset using the best models. If we looked closer on Table S3, we could see that the median of our estimations is below 0.6 log except for cluster 1 (above 1 log).

Then we calibrated the best models on the whole dataset of each cluster: a cluster-then-predict approach using SVM for clusters 1 and 4 and using random forest for clusters 2 and 3. To compare the different models in each cluster and give an idea of what were the important molecular descriptors we provided the five most important molecular descriptors for each cluster in the Table S4. We could see in this table that the important molecular descriptors strongly differ from one cluster to another.

Then the models were used to predict the missing CF_{ET} of the common compounds between USEtox® and TyPol databases. These values were by consequence new estimations of the CF_{ET} for compounds on which we had no information. The prediction intervals were relatively small: less than 0.5 log₁₀ in a log scale which highlighted the robustness of the estimation. They are given in Table S5. No NA value was present in cluster 1 with no prediction for this cluster. For cluster 2 gathering molecules with intermediate molecular mass, 9 CF_{ET} values were predicted for various kinds of compounds. One value concerns the antibiotic sulfamethazine and its value is quite near to the one of sulfamethoxazole and sulfadiazine of the same sulphonamide antibiotic family constituted of the sulphonamide group ($-S(=O)_2-NR_2R_3$). Cluster 3 grouped compounds with the lowest molecular mass and the lowest median CF_{ET} like ibuprofen, phthalates, cresol constituted of monoaromatic ring substituted

with methyl, carboxylic groups. The CF_{ET} prediction for acetylsalicylic acid seemed coherent with the value of the nearest compounds (herbicides mecoprop) of this group. Cluster 4 gathered compounds with the highest median CF_{ET} and that presented a high number of rings halogenated or not, like PAH and hormones. The 5 CF_{ET} predicted concerned 4 PAHs and 1 hormone. By comparison to the 2 other PAHs present in this cluster, the 4 predicted CF_{ET} were quite similar and higher. Concerning the prediction for the hormone, the CF_{ET} was intermediate between the CF_{ET} of the 3 other hormones in the cluster. It seems that all these 5 predicted values were very closed, falling near the median value of this cluster.

3.4.2. Best models for the CF_{HT}

The best models had a global median of 0.75 log for the prediction of the CF_{HT} . We could see on the results gathered on Table S6 that the best performances are for cluster 1 and that they are comparable for the other clusters. Then, the global SVM model was calibrated and computed on the whole dataset. It was used to predict the compound of clusters 2, 3, 4, and 5. Let us recall that there was a single molecule in cluster 5 and, as it has a NA value for its CF_{HT} , the best global model (SVM) was used. For cluster 1, a cluster-then-RF model was computed. The most important descriptors of these two models are gathered in the Table S7 and, as for CF_{ET} , were strongly different between the different best models.

Then, this model was used to predict the CF_{HT} value for the 102 common compounds without a CF_{HT} value. These predictions are reported in Supplementary Table S8. As for the CF_{ET} , the small width of the prediction interval (less than a $\log_{10}$ in a log scale) highlighted the robustness of the approach even with a relatively small number like estimations made for compounds that lie in cluster 1. In this cluster 1, CF_{HT} for a phthalate (DEHP) was already known, but the one for diisodecyl and diisononyl phthalate was predicted with value in the

same range. The 3 cyclines (tetracycline, aureomycin, and oxytetracycline) grouped in cluster 1, presented also similar predicted CF_{HT} . This was also the case for triclosan and triclocarban in cluster 2. Similar predicted and known CF_{HT} were found for four herbicides from the substituted urea family (linuron, diuron, monolinuron, isoproturon) in cluster 3. Cluster 4 gathered a small number of molecules but with the highest median CF_{HT} , the predicted CF_{HT} of the organochlorine insecticide isodrin was similar to another congener of the same family, aldrin.

4. Discussion

It is a real and important challenge to provide characterization factors for a wide range of compounds. Obviously, it is expected that these new calculated factors have an acceptable margin of error. As reported in UNEP/SETAC (2019), it is commonly assumed that the uncertainty of the characterization factors can vary by approximately 2-3 orders of log-magnitude (Rosenbaum et al. 2008) or significantly higher (up to 7 orders) if all sources of uncertainty are considered (Douziech et al. 2019). The results obtained in the previous section are very promising as they are below the level of uncertainty commonly assumed and as they are based on molecular descriptors that could be easily obtained for each compound without ecotoxicity factor. Based on this fact we could already provide 15 new CF_{ET} and 102 new CF_{HT} for the common molecules between USEtox® and TyPol without a previous value.

The idea of predicting ecotoxicity characterization factors for chemicals using machine learning algorithms has already been used (Hou et al., 2020a and 2020b). But, here, our findings go further. Indeed, we show that we could directly obtain accurate estimations of endpoint values from easy-to-obtain molecular descriptors. This will open the door to the fast characterization of each new unknown compound that appears, including transformation products. We also show that the cluster-then-predict approach can give better performances

1 than the approach without the clustering step. This local (*i.e.* cluster-then-predict) approach
2 confirms that local models could be an efficient prediction method when heterogeneity of
3 data generates nonlinear relations between the response and the explanatory variables
4 (Lesnoff et al., 2020).

5 Across the clusters and models, there is a general trend that the non-linear models tend to
6 outperform the linear ones. This suggests that a linear model is not fully adequate to capture
7 the complexity of the relationship between the molecular descriptors and the CFs. However,
8 the use of linear model for e.g. a QSAR is likely due to the ease of interpreting its
9 coefficients, while interpretation is much more challenging for machine learning approaches
10 such as random forest or SVM. Thus, the advantages or drawbacks of linear/non-linear
11 approaches must be balanced according to the final goal of each study. Here, as the main
12 goal is to calculate the most accurate CFs, non-linear models seem more suited. We must
13 also mention that a new emerging field is developing tools needed to help making black-box
14 models (e.g. random forest) more interpretable (Bénard et al., 2021).

15 The difficult interpretability of the machine learning models used in this study can thus be
16 viewed as a limitation. On another side, even if we already had an acceptable number of
17 compounds in our training datasets, the model accuracies would benefit of the inclusion of
18 new compounds. These compounds could be carefully chosen to improve the models where
19 there is a clear need (*i.e.* where the performances of the models are not good enough), for
20 example in the cluster 1 for CF_{ET} or in the cluster 4 for CF_{HT}.

21 One of the interests of USEtox® and its three-step structure (fate - exposure - effect) is that
22 it can be adapted to some specific contexts (a more accurate and spatialized fate model, a
23 different exposure...) while keeping the steps that are not modified. However, these
24 adaptations of USEtox® are not widely used and are reserved for advanced users. Our
25 approach does not allow this, with a direct one-step estimation of CFs. It was designed to
26 provide default CF values for molecules where information is missing. We have chosen to
27 directly predict the CF by simplicity, as the first tests revealed that doing three models (for
28 the three steps) and then calculating the CFs produced less accurate results. It would

1 however be an interesting perspective to estimate only some of the stages by these learning
2 approaches and to combine them with stages modelled in a classical way in USEtox®

3

4 **5. Conclusion**

5

6 This paper presents a modeling method to derive characterization factors from easily
7 obtainable molecular descriptors. The results presented here show that models that can
8 handle non-linearity and that could be adapted to a small number of compounds (using the
9 cluster-then-predict approaches) are the best suited. The cluster-then-predict approaches
10 could also be more accepted by the users, as they allow to consider mainly the compounds
11 similar to the one under investigation. By consequence, the missing characterization factors,
12 as well as those of new molecules, could now be quickly estimated with an overall good
13 precision as performed in Servien et al. (2021). More generally, one of the key factors in the
14 evaluation of toxicity and ecotoxicity in LCA lies in the construction of the characterization
15 factors: a task requiring a large amount of data and a consequent investment of time. The
16 use of machine learning allows us to go beyond these constraints. This makes it possible to
17 obtain characterization factor values in a fast and simple way, which can be used as long as
18 conventionally established CFs are not available.

19

20 **Declaration of Competing Interest**

21

22 The authors declare that they have no known competing financial interests or personal
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7 **Supplementary materials**

8

9 Supplementary material associated with this article can be found, in the online version.

10 Scripts and data used are provided.

11

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