1 2 3 4	<i>Title</i> Copepod functional traits and groups show contrasting biogeographies in the global ocean
5 6	Running title Global marine copepod traits biogeography
7 8 9	<i>Authors</i> Fabio Benedetti ^{1*} (0000-0002-7554-3646)
10 11 12	Jonas Wydler ¹ Meike Vogt ¹ (0000-0002-0608-1935)
13 14 15	¹ Environmental Physics, Institute of Biogeochemistry and Pollutant Dynamics, ETH Zürich, 8092 Zürich, Switzerland. *Corresponding author: <u>fabio.benedetti@usys.ethz.ch</u>
16 17 18	Acknowledgements We thank all contributors involved in the plankton species field sampling and identification
19 20	throughout the world and we acknowledge the efforts made to deposit the data on publicly available online archives. This project has received funding from the European Union's
21 22	Horizon 2020 research and innovation program under grant agreement No. 862923. This output reflects only the author's view and the European Union cannot be held responsible for
23 24	any use that may be made of the information contained therein. We thank Luke Gregor for editing the language of an early version of the manuscript.
25 26	Conflict of interest

- 27 The authors declare no conflict of interests.
- 28

29 Abstract

30 Aim: To define global zooplankton functional groups (FGs) and to estimate their 31 environmental niche and habitat distribution. We model the spatial patterns of copepod FGs

- 32 habitat and identify regions sharing similar functional trait expression at the community level.
- 33 **Taxon**: Marine planktonic Neocopepoda.

34 **Location**: Global ocean.

Methods: Factor analysis on mixed data and hierarchical clustering were used to identify copepod FGs based on five species-level functional traits. An ensemble of species distribution models was used to estimate the environmental niches of the modelled species, project the mean annual habitat suitability of the FGs, and to estimate the community weighted mean values of the traits studied. Ocean regions were defined based on their community-level mean trait expression using a principal component analysis and hierarchical clustering.

41 **Results**: Eleven global copepod FGs were identified. They displayed contrasting latitudinal 42 patterns in mean annual habitat suitability that could be explained by differences in 43 environmental niche preferences: two FGs were associated with polar conditions, one 44 followed the global temperature gradient, five were associated with tropical oligotrophic 45 gyres, and the remaining three with boundary currents and counter currents. Four main 46 regions of varying community weighted mean trait values emerged: the Southern Ocean, the 47 northern and southern high latitudes, the tropical gyres, and the boundary currents and 48 upwelling systems.

49 **Conclusions**: We build on an exhaustive species trait dataset to put forward novel FGs that 50 will improve the representation of zooplankton in global marine ecosystem models. Our 51 results contribute to our understanding of the spatial patterns and drivers of marine plankton 52 trait biogeography and will serve as a basis for studying the links between zooplankton 53 biodiversity and ecosystem functioning and how they might evolve in the context of climate 54 change.

55

56 Keywords

57 Planktonic copepods, Functional groups, Trait-based approach, Species Distribution58 Modelling, Global Ocean, Community weighted mean trait

60 **1. Introduction**

61 Copepods are crustaceans that dominate the biomass of the mesozooplankton size class (0.2-62 2.0 mm) and rank amongst the most abundant animals in the oceans (Turner, 2004; Kiørboe, 63 2011a). They are morphologically and functionally diverse and are adapted to almost all 64 marine ecosystems (Kiørboe, 2011a; Bron et al., 2011). Copepods play a pivotal role in the 65 food web, both as microplankton grazers and prey for small pelagic fishes and other larger 66 animals (Beaugrand, Edwards, & Legendre, 2010; Steinberg & Landry, 2017). Copepods play 67 a critical role in the marine biological carbon pump, the fixation of inorganic carbon to 68 organic matter by photosynthesis and consequent sequestration away from the surface into the 69 deep ocean. Planktonic copepods export carbon from the euphotic layer to the deep ocean in 70 several ways, such as the grazing of phytoplankton cells in the sunlit layers, followed by the 71 excretion of relatively fast-sinking fecal pellets, or by extensive vertical migration (Turner, 72 2015; Steinberg & Landry, 2017).

73 The relative contribution of planktonic copepods to the above-mentioned processes is 74 mediated by their diversity and trait expression (Barton et al., 2013). For instance, larger 75 copepods perform stronger diel vertical migration (Ohman & Romagnan, 2016) and produce 76 larger fecal pellets that sink faster and thus increase the proportion of particulate organic 77 carbon sinking to depth (Stamieszkin et al., 2015; Brun et al., 2019). The dominance of 78 certain copepod feeding modes (i.e., the mechanism through which prey is captured) impacts 79 food-web dynamics as feeding modes affect copepod grazing and mortality rates (Kenitz, 80 Visser, Mariani, & Andersen, 2017; van Someren Gréve, Almeda, & Kiørboe, 2017). Yet, 81 mechanistic ecosystem models usually represent zooplankton only through a couple of size 82 classes (Le Quéré et al., 2005), which greatly oversimplifies the contribution of species 83 diversity and functional traits to the functioning of ecosystems and biogeochemical cycles 84 (Flynn et al., 2015). Multiple modelling studies show that the number of zooplankton 85 functional groups and their grazing characteristics exerts an important control on the biomass 86 and diversity of other trophic groups, with consequences for global biogeochemical cycles in 87 these models (Prowe et al., 2012; Sailley et al., 2013; Vallina et al., 2014; Le Quéré et al., 88 2016). Ecosystem models, however, rely strongly on parameterisations of specific traits 89 governing biological processes and food-web interactions (e.g. grazing rates or food 90 preferences) based on scarce information, as the empirical evidence included in current 91 models is often sourced from limited laboratory or field data from a narrow range of 92 culturable species (Barton et al., 2013). Consequently, investigating the potential links

93 between copepod trait distribution and ecosystem functioning contributes to improving
94 current ecosystem models (Stocker, 2014).

95 To better understand the role of planktonic copepods in marine systems, plankton ecologists 96 are increasingly adopting trait-based approaches (Litchman, Ohman, & Kiørboe, 2013; 97 Hébert, Beisner, & Maranger, 2017). Functional traits are species- or organism-level 98 characteristics (morphological, behavioral, physiological, or related to life history) that affect 99 their fitness and contribute to ecosystem functioning (Violle et al., 2007). According to 100 Litchman et al. (2013), functional traits can be classified by the main functions they 101 contribute to, namely, survival, feeding, growth, and reproduction. Organisms often cannot 102 maximize all four ecological functions at the same time since there are trade-offs between 103 these functions. For example, copepods that rely on passive feeding (i.e., copepods that 104 remain immobile and only move to capture prey) show 8.5 times lower mortality rates than 105 copepods that rely on active feeding (van Someren Gréve, Almeda, & Kiørboe, 2017). Yet, 106 adult male copepods often need to move actively to find a mate for reproduction, which 107 undermines the benefits of their feeding mode (Kiørboe, 2011b; Litchman et al., 2013).

108 Copepod functional traits can be investigated by grouping species into functional groups 109 (FGs) according to similarities in functional trait combinations (Benedetti, Gasparini, & 110 Ayata, 2016). Categorizing species according to their similarity in functional traits rather than 111 their taxonomic classification allows one to summarize highly diverse groups, and/or 112 communities, into more parsimonious categories defined by distinct ecological functions 113 (Barnett, Finlay, & Beisner, 2007; Benedetti, Vogt, Righetti, Guilhaumon, & Ayata, 2018). At 114 the regional scale, Pomerleau et al. (2015) showed that 42 zooplankton species from the North 115 East Pacific Ocean could be clustered into five groups based on their body length, feeding and 116 reproduction mode, and trophic group. Based on a similar set of traits, Benedetti, Vogt, et al. 117 (2018) grouped 106 marine copepod species that dominate Mediterranean communities into 118 seven FGs. They found that carnivorous species were associated with tropical oligotrophic 119 conditions, whereas current-feeding herbivores were associated with more productive and 120 seasonally varying conditions. Such groupings can improve the representation of zooplankton 121 in an ecosystem model since FGs increase the representation of ecological function without 122 adding taxonomic diversity and complexity. However, studies based on a regional pool of 123 species might underestimate the real range of functions ensured by zooplankton in the global 124 ocean, warranting the need for a study based on a global species pool.

126 The functional composition of observed plankton communities can be investigated through 127 community weighted mean (CWM) values of traits (Ricotta, 2005; Pomerleau, Sastri, & 128 Beisner, 2015; Brun et al., 2016). This is done by calculating the proportion of species in a 129 community that exhibits a particular functional trait value by accounting for their relative 130 contribution to community composition or abundance. Thereby, CWM values of functional 131 traits help to identify the relationships between the emerging expression of functional traits on 132 a community-level, the environmental conditions associated to the spatial patterns of trait 133 expression (i.e., their biogeographical patterns), and the potential trade-offs underlying trait 134 expression. Like for FGs, empirical CWM trait values of marine zooplankton have been 135 explored on local or regional levels for the northern hemisphere mostly (Pomerleau et al., 136 2015; Pecuchet et al., 2018). They evidenced a negative impact of warming and seasonality 137 on community-level body length and offspring size. The first truly global study by Brun et al. 138 (2016) also found larger body sizes, higher proportions of myelinated species, and relatively 139 smaller offspring size in latitudes $>50^\circ$, because of gradients in temperature and seasonality in 140 phytoplankton production. To our knowledge, other qualitative traits, such as trophic groups 141 or spawning mode, have not been investigated through the CWM trait approach (McGinty et 142 al., 2018; Benedetti, Vogt et al., 2018). Therefore, their global spatial distribution remains poorly documented and a more complete understanding of trade-offs between the expression 143 144 of these diverse traits is still needed to better characterize and map ecosystem functions 145 worldwide. If geo-referenced trait information, such as patterns of CWM values, are clustered 146 in space and time, they can be used to define ocean regions sharing common ecological 147 characteristics (Longhurst, 2010; Reygondeau et al., 2017; Hofmann Elizondo et al., 2021). 148 Such regionalization schemes have never been applied on plankton communities' functional 149 trait expressions on a global scale, although this could be more informative of large-scale 150 marine food web and ecosystem dynamics than previous regionalization schemes based on 151 environmental parameters alone.

152 In this study, we address the following questions: (i) What are the global zooplankton FGs 153 emerging from the/a clustering of the most frequently sampled copepod species according to 154 their functional traits (Benedetti et al., 2016)? (ii) Which parameters structure FGs most 155 strongly in functional trait space and environmental niche space (Benedetti, Vogt, et al., 156 2018)? (iii) What are the habitat distribution patterns of these FGs and their community-level 157 trait expression (estimated through CWM trait values), and which regions of the global ocean 158 share similar trait expression? To do so, a new functional trait synthesis was carried out for a 159 global pool of >380 frequently observed copepod species, and five traits (body size, trophic

160 group, feeding mode, myelination, and spawning mode) were used to identify novel FGs 161 based on species' traits combinations. Occurrence-based species distribution models (SDMs) 162 were used to investigate the differences in the environmental niches and monthly habitat 163 distribution modelled for the new copepod FGs. Mean annual CWM trait values were derived 164 from the monthly SDMs-based habitat distribution maps, and hierarchical clustering was used 165 to identify regions that share similar community-level trait expression.

166

167 2. Materials and Methods

168 2.1. Species occurrence data

We used the zooplankton occurrences dataset (geo-localised and dated presences) of 169 170 Benedetti et al. (2021), which has been gathered to model global zooplankton species 171 biogeography, diversity and community composition. The dataset combines 165'716 172 occurrences binned into the 1° x 1° grid of the World Ocean Atlas (WOA; Boyer et al., 2013) 173 for 385 copepod species. The occurrences were initially retrieved from the Ocean 174 Biodiversity Information System (OBIS; https://www.obis.org), the Global Biodiversity 175 Information Facility (GBIF; https://www.gbif.org), as well as complementary datasets (see 176 Benedetti et al., 2021 for an exhaustive description and for the curation of species names). 177 The occurrences corresponding to benthic and parasitic taxa and those that were not fit to 178 model the habitat of copepod species in the surface open ocean were discarded as follows: 179 occurrences with missing spatial coordinates, sampling dates, sampling depth, or taxonomic 180 identification at species-level were removed. Furthermore, occurrences from drilling holes, 181 freshwater (sea surface salinity < 20 according to the WOA, grid cells within 25 km of the 182 nearest shoreline, or with a sampling depth >500 m were also discarded.

183 The copepod occurrence data displayed sampling effort biases (Appendix S1) that could 184 inflate metrics of empirical models' performance or over-represent portions of the 185 environmental space and thus hinder model predictability and interpretation (Veloz, 2009; 186 Hijmans, 2012). We addressed these potential sampling biases by further thinning the species-187 level occurrence data (Aiello-Lammens, Boria, Radosavljevic, Vilela, & Anderson, 2015). 188 For each month and each species separately, the occurrences were thinned by applying a 189 randomization algorithm (30 randomizations per species dataset) that returned a 'thinned' 190 dataset where monthly species occurrences were at least 500 km apart (Aiello-Lammens, 191 Boria, Radosavljevic, Vilela, & Anderson, 2015). Only those species displaying at least 50 192 occurrences (n = 385) were retained and constituted the final list of species for which 193 functional traits data were searched for in the literature.

194

195 2.2. Species functional trait data

The five following species-level functional traits were included in our analyses, as detailed
below. These traits were chosen based on data availability from previous traits compilations
and extended literature survey (see Appendices S2 and S3):

Body size (quantitative continuous): mean maximum adult female body size (i.e. length of the cephalothorax) in mm. Body size is considered a master functional trait as it impacts all life functions and scales most physiological rates (Kiørboe & Hirst, 2014; Hébert, Beisner, & Maranger, 2017) and influences predator-prey interactions (Hansen, Bjornsen, & Hansen,

203 1994).

Trophic group (categorical): most marine planktonic copepods are omnivorous, yet they can be grouped according to their preference in food sources (Kiørboe, 2011a; Pomerleau et al., 2015; Benedetti et al., 2016). Here, the following five groups were defined: Omnivore-Herbivore, Omnivore-Carnivore, Omnivore-Detritivore, strict Carnivore, and Omnivore. The trophic group describes the primary food source of a species and therefore its role in foodweb dynamics.

210 Feeding mode (categorical): copepods have developed various strategies to detect and capture 211 food items. The present feeding modes followed the definitions of Kiørboe (2011a): ambush-212 feeding, current-feeding, cruise-feeding, particle-feeding, current-cruise feeding, and current-213 ambush feeding (the last two referred to those mixed-feeding species that perform both 214 strategies; Kiørboe, 2011a; Litchman et al., 2013). Ambush-feeding copepods lurk stationarily 215 in the water column and detect the vibration generated by motile preys thanks to specialized 216 appendages (i.e., hydromechanical perception) and capture prey through quick jumps. 217 Copepods perform active ambush-feeding according to Kiørboe (2011a), which is considered 218 a more passive feeding mode than cruise-feeding and current-feeding (Brun et al., 2017). 219 Current-feeding describes the use of a scanning current that can detect and capture food items. 220 This strategy is believed to favor a more efficient capture of many immobile preys like 221 phytoplankton cells compared to ambush-feeding (Kiørboe, 2011a). Lastly, cruise-feeders 222 correspond to copepods that swim actively through the water column in search of their prey.

Myelination (binary): myelinated copepod species have a lipid-rich myelin sheath around their nerves that increases the speed of nervous signals transmission and therefore enables faster attack or evasive reactions (Lenz, 2012). Myelin sheaths play a key role in modulating mortality and feeding rates and improving energy savings under low food conditions.

Spawning mode (binary): eggs are either released into the open water after fertilization (freespawning) or remain attached to the female in specialized egg-sacs until hatching (sacspawning). Sac spawning copepods display lower fecundity rates and longer hatching times

- compared to free-spawners (Kiørboe & Sabatini, 1994).
- 231
- 232

2.3. Definition of functional groups

233 We used multivariate analysis and hierarchical clustering to identify functional groups of 234 copepod species that show similar functional trait combinations and therefore constitute 235 different functional units in marine ecosystems. Only those copepod species with a body size 236 value and no missing data for three out of five functional traits (343 out of the 385 species) 237 were retained in the analysis (Appendix S2). The species' trophic groups and feeding modes 238 were re-coded binarily to accurately represent species that belong to several feeding modes/ 239 trophic groups (Appendix S2). This means that the analyses described below were performed 240 on 343 species based on ten trait dimensions instead of the initial five dimensions.

241 Since the functional traits used represented both quantitative and qualitative variables, a factor 242 analysis on mixed data (FAMD; Pagès, 2004) was used to investigate inter-species functional 243 traits variance and estimate the functional distances between species. The use of a FAMD 244 represented an improvement to the Multiple Correspondence Analysis (MCA) used in 245 Benedetti et al. (2016) as it allowed to retain body size as a quantitative continuous variable 246 instead of having to convert this trait into size classes. The FAMD reduced the dimensionality 247 of the ten traits dataset by synthetizing them through a set of principal components (PCs) that 248 described the main modes of functional trait variance. The number of retained FAMD 249 components was chosen based on a leave-one-out cross-validation method where each object 250 (i.e., species) is alternatively removed and predicted with a new FAMD model and a mean 251 square error of prediction is calculated for each n components ranging from one to ten 252 (Audigier et al., 2016). The mean square error was lowest for n =four components (error = 253 0.028, whereas error > 0.029 for all other values of n) and those first four FAMD components 254 explained 80.15% of the variability in inter-species trait variance. An Euclidean distance 255 matrix was calculated from these four components to represent the inter-species functional 256 distance resulting from the FAMD. To assess the quality of this dimension reduction analysis, 257 we compared this Euclidean distance to a Gower distance matrix calculated directly from the 258 trait data (10 dimensions) following the guidelines of Mouillot et al. (2021), We calculated 259 the co-ranking matrix between the two distance matrices and determined how the dimension 260 reduction distorted the initial inter-species functional distance using the area under the curve

(AUC) criterion, which should be > 0.7 according to Mouillot et al. (2021). We found an
AUC criterion of 0.81 which supports the "excellent" quality of our dimension reduction
analysis (Mouillot et al., 2021). We used hierarchical clustering and Ward's agglomeration
link to generate the functional dendrogram. The number of clusters (i.e., FGs) was then
chosen by the height at which the dendrogram is "cut".

266 We cared to investigate the sensitivity of the resulting FGs to the main parameters of our 267 clustering approach (see Appendix S4 for results not reported below): (i) the effect of 268 including species with missing values in the FAMD compared to restricting the pool to those 269 species without any missing trait value (283 instead of 343; Appendix S2), (ii) the choice of 270 the distance metric (Euclidean distance matrix based on FAMD components versus Gower's 271 distance matrix based on the trait values), and the agglomeration link (Ward's aggregation 272 link versus average link; Mouchet et al., 2008). Therefore, four combinations of distance 273 matrices and agglomeration links were considered for two subsets of the species traits data. 274 Including 60 copepod species with some missing trait values affected the structure FAMD 275 space only marginally (Fig. S4.1) and did not change the number of components to be 276 retained (n = 4), nor did it change their amount of explained variance (80.15%). Therefore, 277 the FAMD including 343 species was kept as the standard since it allowed for a larger pool of 278 species. To quantify the quality of the clustering procedure, four internal cluster stability 279 indices (Calinski-Harabasz index, Connectivity, Dunn's index, and average silhouette width) 280 were calculated for each dataset, distance matrix choice, agglomeration link choice and for a 281 number of clusters ranging from three to 13 (Figs. S4.2 and S4.3). Plus, Baker's Gamma 282 correlation coefficients (Baker, 1974) were computed between each functional dendrogram 283 stemming from the four different combinations of distance metrics and agglomeration links to 284 examine their pairwise similarity (Table S4.1). Lastly, we performed a careful expert 285 inspection of each dendrogram and resulting FGs to further evaluate their similarity and 286 ensure the ecological relevance of the final FGs (i.e., avoid few large groups that are 287 functionally heterogenous or numerous small groups that are functionally redundant). 288 Ultimately, eleven FGs were identified (Table S4.2).

- 289
- 290 2.4. Species distribution modelling
- 291 2.4.1. Environmental predictor selection

We considered 14 environmental predictors that drive the spatial ranges of copepod species in the open ocean and that are commonly used to model their abiotic habitats (see Benedetti et al., 2021 and references therein). A more exhaustive description of the data coverage and 295 initial resolution is given in Appendix S5. First, five predictors were retrieved from the World 296 Ocean Atlas (WOA, version 2013v2): sea surface temperature (SST, °C), dissolved oxygen at 175 m depth (dO₂, μ ml l⁻¹), nitrate concentration (NO₃⁻, μ M), phosphate concentration (PO₄³⁻, 297 μ M), silicate concentration (SiO₂, μ M). The products N^{*} and Si^{*} were derived from the 298 monthly climatologies of the above-mentioned nutrient fields: N^* refers to the excess of NO_3^- 299 to PO_4^{3-} relative to the Redfield ratio (N^{*} = [NO₃] - 16[PO₄^{3-}]) and can be used as a tracer of 300 denitrification and N₂ fixation. Si^{*} is the excess of SiO₂ to NO₃ (Si^{*} = [SiO₂] - [NO₃⁻]) and is 301 302 relevant to diatom growth, as healthy diatoms take up silicate and nitrate in a one-to-one ratio 303 (Sarmiento & Gruber, 2006). Several additional variables were considered as they represent complementary niche axes: surface wind speed (Wind, m s⁻¹); eddy kinetic energy (EKE, m² 304 s^{-2}) computed according to the algorithm of Qiu and Chen (2004) as a proxy for mesoscale 305 activity; chlorophyll-a concentration (Chl, mg m⁻³) as a proxy of surface phytoplankton 306 307 biomass; surface carbon dioxide partial pressure (pCO₂, µatm); photosynthetically active radiation (PAR, μ mol m⁻² s⁻¹); and mixed-layer-depth (MLD, m). Also, photosynthetically 308 available radiation (PAR) over the MLD (MLPAR, μ mol m⁻² s⁻¹) was added as an indicator of 309 available light within the mixed layer. For NO_3^{-7} , PO_4^{-3-7} , SiO_2 , Chl and EKE, we added their 310 311 log-transformed fields to achieve distributions that are closer to a normal distribution. The 312 monthly climatologies of all predictor variables were all projected onto the $1^{\circ}x1^{\circ}$ cell grid of 313 the WOA and their values were matched with the monthly species occurrence data.

314 We used a two-stage procedure to select the environmental predictors of the SDMs (section 315 2.4.3 for the models set up). First, we removed collinear predictors to avoid increasing the 316 uncertainty in regression models projections through coefficients inflation (Dormann et al., 317 2013). The collinearity of predictors at the global scale and at the FGs-level was calculated 318 for each species through pairwise Spearman's rank correlation coefficients. When a pair of 319 predictors displayed an average correlation coefficient above the widely used |0.7| threshold 320 (Dormann et al., 2013), the predictor closest to a normal distribution was kept. This first step narrowed down our initial 14 predictors to the following ten: SST, PAR, logNO₃, MLD, 321 logChl, logEKE, Si^{*}, N^{*}, Wind, and pCO₂. Second, to avoid model overfitting the number of 322 323 environmental predictors included in the SDMs was restricted to five to achieve a 10:1 ratio 324 of occurrences to predictors, following Guisan et al. (2017). To choose the five most 325 influential predictors, univariate random permutation tests were performed (n = 30 repetitions; 326 Appendix S6). For each species and SDMs, one of the ten potential predictors was randomly 327 reshuffled while the other nine were kept as is and SDMs were trained based on this 328 reshuffled dataset. Then, a correlation coefficient was calculated between the original vector

of unshuffled model prediction and the vector of predictions resulting from the reshuffled dataset. The higher the correlation coefficient, the lower the importance of the reshuffled predictor on the final SDM prediction. We chose the top five predictors for each FG and SDM separately (Appendix S6). These were then used to train the SDMs used for global habitat projections. For those species that could not be assigned to a FG due to missing functional traits data, the five predictors displaying the highest ranks of relative importance across all species combined were used in the SDMs (Appendix S6).

336

337 2.4.2. Background data

338 Correlative SDMs such as the ones used here require both presence and absence data. Yet, the 339 occurrence data used here typically lack absence data, therefore pseudo-absence data were 340 generated (Benedetti et al., 2021). We followed the target-group approach of Phillips et al. 341 (2009) which is appropriate to model plankton species distributions based on their relatively 342 sparse data (Righetti et al., 2019; Benedetti et al., 2021). First, we investigated the spatial 343 distribution of the occurrence data of each FG separately. All FGs displayed similar patterns 344 of sampling effort except for FG2, FG4 and FG11 (Appendix S4), which had very few 345 occurrences in the Southern Ocean and the North Atlantic Ocean (Appendix S1). Therefore, 346 we chose to draw pseudo-absences from the total pool of sites (i.e., monthly 1° x 1° grid cells) 347 that displayed at least ten occurrences of all species included, except for those species 348 belonging to the three groups above. For species of FG2, FG4 and FG11, pseudo-absences 349 were only drawn from those sites where at least 10 different species of their FG were found 350 (i.e., their "target group"). The threshold of ten occurrences per site was chosen to avoid 351 relying on those sites where only a small fraction (e.g., only one or two species) of the overall 352 copepod community was detected. Once the sites were defined, the pseudo-absences of each 353 species were randomly drawn based on their occurrences in both their corresponding target 354 groups. This way, a species' background is located at the sites where its lack of presence is 355 most likely to reflect an actual absence. For each species, ten times more background data 356 than presences were generated following the guidelines of Barbet-Massin et al. (2012), and 357 presences were weighted ten times more than pseudo-absences.

358

359 2.4.3. Model configuration and habitat suitability projections

Following Benedetti et al. (2021), we developed an ensemble modelling approach that combined three SDMs types which were configured to avoid model overfitting (Merow et al., 2014): Generalized Additive Models (GAMs), Generalized Linear Models (GLMs), and

363 Artificial Neural Networks (ANNs). The GLM were run using a quadratic formula and a logit 364 link function. The GAMs were also set up to use a logit link function and a maximum of five 365 smoothing terms. ANNs were run using five cross validations with 200 iterations to optimize 366 number of unities in the hidden layer and the parameter for weight decay (Thuiller et al., 367 2016). The SDMs were trained on 80% of the presence/pseudo-absence data chosen at 368 random and tested on the remaining 20%. For each SDM, ten random cross evaluations runs 369 were carried out. For each cross-validation run, the true skill statistic (TSS) was calculated to 370 evaluate model performance. Only species displaying a mean TSS score >0.30 were used for 371 the final species and functional trait projections.

372 The SDMs were then projected in the conditions of the global ocean as a function of all 373 monthly climatologies of the environmental predictors included in the model to obtain 374 species-level maps of habitat suitability indices (HSI). Since we used FGs-specific and SDM-375 specific sets of predictors, the SDMs vary in number of successfully modelled species (254 376 for GAM, 244 for ANN, and 238 species for GLM). For each FG, annual mean HSI values 377 were obtained by averaging the monthly HSI maps across each species constituting the FG 378 and SDMs. These mean annual HSI estimates indicate which regions are suitable for each 379 species/FG.

380

381

2.5. Positioning FGs in environmental niche space

382 Following the approach of Benedetti, Vogt, et al. (2018), the mean univariate response curves 383 emerging from the GAMs were used to estimate the relative optimal conditions (i.e., niche 384 center) and tolerance (i.e., niche width) for each species and predictor (see Appendix S7). The 385 niche center was calculated as the weighted median and the niche width as the difference between the weighted 10th and 90th percentiles. The HSI values were used as weights. For this 386 387 analysis, a set of eight predictors common to all species (i.e., unlike the top five FG-specific 388 predictors used for mapping, as described in section 2.4.1) had to be defined based on the ranking of predictors (Appendix S6): SST, logChl, logNO₃, MLD, Si^{*}, PAR, N^{*} and logEKE. 389 390 Eight predictors were chosen to cover all top ranking predictors based on the overall rankings. 391 To investigate the inter-species and inter-FGs similarities in the above-mentioned niche 392 characteristics and test if the FGs present distinct environmental niches, a principal 393 component analysis (PCA; Legendre & Legendre, 2012) was performed on the niche centers 394 and niche widths for all 8 predictors to ordinate the species according to their niche center and 395 niche width. This space composed by the principal components (PCs) of the PCA is hereafter 396 referred to as the niche space. To explore whether FGs differ significantly in their position in

397 niche space, we averaged the PCs scores of each FG based on the scores of their constituting 398 species. Non-parametric variance analysis (Kruskal-Wallis tests) was carried out to test if FGs 399 differ in their position in niche space). Post-hoc variance analyses (Dunn's test with a p-400 values adjustment following Bonferroni's method) were then performed to identify the pairs 401 of FGs that displayed significant variations in niche space positions.

402 403

2.6. Community weighted median (CWM) proportions of traits

404 On top of projecting the FGs-specific mean annual patterns of HSI, the CWM trait values for 405 body size, trophic group, feeding mode, myelination, and spawning mode were computed to 406 explore the biogeography of copepod traits. All qualitative traits were mapped by calculating 407 a CWM trait value based on the monthly species-specific HSI values as weights. More 408 precisely, for each grid cell and month, the CWM proportion was calculated as the sum of all 409 HSI values belonging to species exhibiting that trait over the sum of HSI values across all 410 species in that community (i.e. all the species in that grid cell). For body size (i.e., the only 411 continuous trait), we calculated the weighted median body size for each community, again 412 using monthly HSI values as weights. The CWM values were calculated according to Cormen 413 et al. (2009) using the matrixStats R package. Like for FGs, monthly CWM projections were 414 performed for every SDM type and mean annual CWM trait values were derived based on the 415 ensemble of projections (Appendices S8 and S9).

- 416
- 417

2.7. Ocean regionalization based on functional trait biogeography

418 To explore sub-global patterns of traits and FGs expression rather than reporting results for 419 broad latitudinal bands (e.g., going beyond the tropics *versus* the poles), clustering was used 420 to divide the global ocean into regions that displayed similar trait expressions (e.g., mean 421 annual projections of CWM body size, and CWM values of each trophic groups, feeding 422 modes, myelination, and spawning mode; Appendix S10). First, the spatial patterns of the 423 twelve CWM traits values were summarized through a PCA again (Appendix S10) and the 424 scores of each grid cell along the first four PCs (97.9% of total variance explained) were used 425 to derive a global Euclidean distance matrix. Second, following the same approach as in 426 section 2.3 and in Benedetti et al. (2021), various clustering approaches were explored based 427 on this distance matrix (three hierarchical clustering approaches and two widely used 428 partitioning approaches) and three internal cluster stability metrics (Dunn's index, 429 Connectivity, and the average silhouette width) were used to help us find an optimal number 430 (k) of regions (see Appendix S11). The three hierarchical approaches consisted in using three

431 alternative agglomeration linkages (average, complete and Ward's; Legendre & Legendre, 432 2012). The stability metrics enabled us to discard the two partitioning methods (kmeans and 433 partitioning around medoids; Fig. S11.1) and to narrow k down to 3 to 8 regions (Fig. S11.2). 434 Ultimately, we chose to present the CWM trait patterns obtained for k = 4 under Ward's 435 linkage based on the profiles of cluster stability metrics and because this linkage methods 436 minimizes intra-cluster variance. Values of k > 4 simply lead to smaller regions nested in 437 those obtained for k = 4.

438

439 **3. Results**

440 3.1. Copepod FGs

441 The 343 copepod species retained were clustered based on their combinations of body size, 442 feeding mode, trophic group, myelination, and spawning mode. Eleven FGs could be derived 443 from the functional dendrogram, which reflects inter-species functional similarity (Fig. 1; 444 Appendix S4). The first dichotomy in the dendrogram occured between the small to medium-445 sized cruise feeders (FG1, 2, 3) and other feeding modes. FG1 was composed of 14 species 446 that are small (median \pm IQR = 1.10 \pm 0.52 mm) cruise-feeding herbivores. Almost all species 447 (13 out of 14) were sac-spawning and myelinated species. All but one species belonged to the 448 Clausocalanus genera. FG2 was defined by small (0.80 \pm 0.30 mm) cruise-feeding 449 detritivores. The 22 species of this group were all amyelinated and sac-spawning. All species 450 in group 2 belonged to the Oncaeidae family. FG3 was composed of 46 medium-sized (2.61 \pm 451 1.70 mm) detritivores. Most species in FG3 were myelinated (74%) and free-spawning (80%). 452 Cruise- and current-feeding were the two predominant feeding modes, with Spinocalanus, 453 Metridia and Scaphocalanus being the dominant genera in this FG.

The second dichotomy in the dendrogram separated the small and large sac-spawning carnivores (FG4 and 5) from the rest. FG4 consisted of 29 small (1.22 ± 0.70 mm) carnivorous ambush-feeders. They were all sac-spawning and amyelinated. The 29 species were part of the Corycaeidae family and thus belonged to either the *Corycaeus*, *Farranula* or *Vettoria* genera. FG5 was made up of 29 species of large (4.58 ± 2.75 mm) sac-spawning current- and cruise-feeding carnivores from the Sapphirinidae and Euchaetidae families.

The third dichotomy of the dendrogram separated the large and medium-sized current feeders (FG6 and 7) from the rest. FG6 was a group of 13 very large (6.80 ± 2.15 mm) species that were either current-feeding herbivores (85%) or fully omnivorous (15%). They were all myelinated free-spawners. The two most represented genera were *Calanus* and *Eucalanus*. FG7 was the largest FG with 55 medium-sized (1.80 ± 2.26 mm) and current-feeding herbivores. Most were myelinated (95%) and free-spawning (89%). The main genera
contributing to the composition of FG7 were *Calanus*, *Calocalanus* and *Paracalanus*.

467 The fourth dichotomy occured between the small and medium-sized omnivores and the

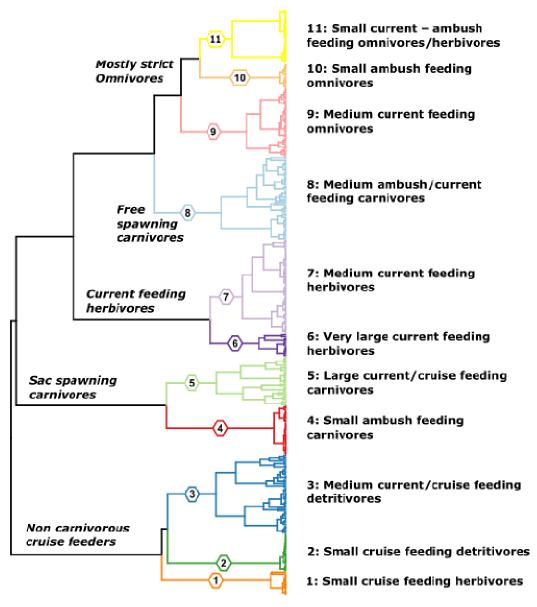
468 medium free spawning carnivores (FG8). FG8 consisted of 49 medium-sized $(3.40 \pm 1.7 \text{ mm})$

469 omnivorous-carnivorous species which were predominantly amyelinated (86%) and free-

470 spawning (80%). This diverse group mixed current- (60%) and ambush-feeders (40%) and

- 471 gathered 12 genera with *Candacia*, *Haloptilus* and *Heterorhabdus* being the most dominant.
- 472 FG9 was composed of 40 medium-sized (2.76 ± 1.98 mm), current-feeding omnivorous
- 473 species. All were free-spawning and mostly amyelinated (72%). Pleuromamma, Gaetanus,
- 474 and *Labidocera* were the main genera.

475 Finally, the fifth and last main dichotomy separated the small and medium-sized mixed-476 feeding omnivores (FG11) from the small ambush feeders (FG10). FG10 was also rather 477 homogeneous and contained 14 small (1.10 ± 0.52 mm), ambush-feeding, amyelinated and 478 sac-spawning omnivores. All species belonged to either the Oithona or the Dioithona genera 479 from the Oithonidae family. FG11 was a group of 32 small (1.50 \pm 0.67 mm) species 480 predominantly belonging to the Acartia and Centropages genera. The species belonging to 481 Acartia were omnivorous-herbivorous, whereas the others were omnivorous. All species in 482 FG11 were amyelinated free-spawning current-ambush feeders.



484

485 Figure 1: Functional dendrogram representing the inter-species traits dissimilarity for 343 486 copepod species based on their functional trait combinations. The hierarchical clustering was 487 performed on a Euclidean distance matrix issued from the species coordinates ensuing from a 488 Factor Analysis of Mixed Data (FAMD). Each leaf of the functional dendrogram represents a 489 copepod species and the eleven functional groups (FGs) identified through the hierarchical 490 clustering approach are numbered and highlighted in color. The main taxonomic groups 491 representing the FGs are as follows: FG1: *Clausocalanus* spp.; FG2: Oncaeidae; FG3: 492 Spinocalanus, spp., Scaphocalanus spp. and Metridia spp.; FG4: Corycaeidae; FG5: 493 Sapphirinidae and Euchaetidae; FG6: Calanidae; FG7: Paracalanus spp., Calocalanus spp., 494 and Calanus spp.; FG8: Haloptilus spp., Heterorhabdus spp. and Candacia spp.; FG9: 495 Pleuromamma spp., Gaetanus spp., and Labidocera spp.; FG10: Oithonidae; FG11: Acartia spp. and *Centropages* spp. 496

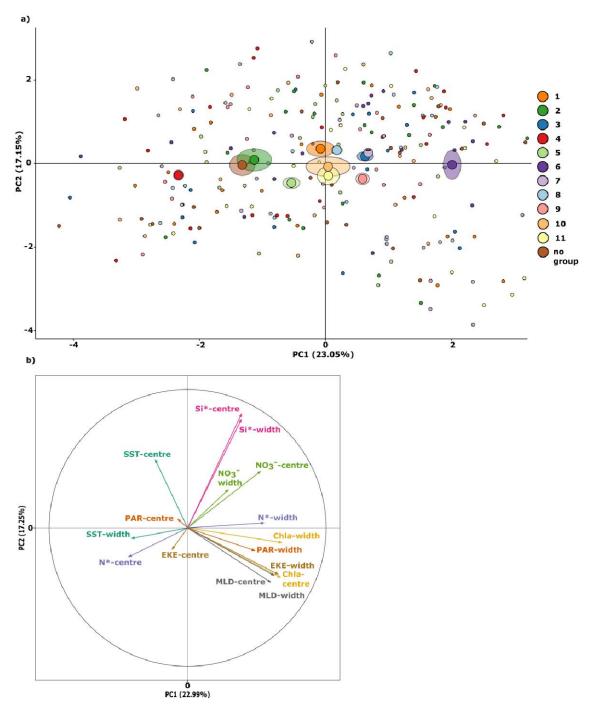
498 The functional space defined by the PCs of the FAMD (Appendix S12) served as an 499 alternative description of the reduced functional space. The first four PC of the FAMD 500 explained 79.21% of the total variance in functional traits. The largest functional distance 501 along the first component was found between FG4 and FG6. The Euclidean distance matrix 502 represented on the functional dendrogram (Fig. 1) was compared to a Gower distance matrix 503 computed from the functional traits (i.e., no dimension reduction; Appendix S4) based on the 504 AUC criterion. We found an AUC of 0.808, which indicated that the reduced FAMD space 505 results in a fairly good representation of the inter-species functional dissimilarity according to 506 Mouillot et al. (2021). Furthermore, the importance of a functional trait in scoring the 507 functional space was quantified by examining the drop in AUC score if said trait was omitted. 508 Body size was found to be the most important (-0.322 in AUC), followed by trophic group (-509 0.265), feeding mode (-0.163), spawning mode (-0.091), and myelination (-0.063).

510

511 3.2. FGs in environmental niche space

512 The environmental niches of the copepod species were described by their univariate niche 513 centers and widths for the top eight predictors we defined as part of our core set or predictors 514 (section 2.5, Appendix S6), and these were used to examine FGs position in niche space based 515 on a PCA (Fig. 2; see Appendix S13 for PCs 3 and 4). The niche characteristics that 516 contributed the most positively to PC1 (relative contribution to PC1 given in brackets when 517 >5%) were: logChl center (12.16%), logChl width (12.51%), EKE width (11.78%), MLD center (9.96%), MLD width (10.78%), N^* width (7.48%), Si^* width (5.64%) and Si^* center 518 519 (5.19%). The species with negative scores on PC1 were those that are characterized by higher SST width (5.65%) and N^{*} center (4.45%). The three most important niche characteristics 520 scoring PC2 were: Si^{*} center (30.88%), Si^{*} width (27.90%) and SST center (11.17%). Species 521 522 with positive PC1 scores were those affiliated with wider niches and conditions of higher 523 concentrations of nutrients and chlorophyll-a, stronger seasonal variations, and overall higher 524 water column turbulence and mixing.

bioRxiv preprint doi: https://doi.org/10.1101/2022.02.24.481747; this version posted February 25, 2022. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under aCC-BY-NC-ND 4.0 International license.



527 Figure 2: Position of a) functional groups (FGs, larger circles) and species (smaller circles) in 528 environmental niche space according to the first two principal components (PCs) of a 529 principal component analysis (PCA) performed on b) the species-level niche characteristics 530 derived from GAM-based univariate response curves of the chosen environmental predictors. 531 Niche centers were calculated as the weighted median value of the corresponding predictor as an estimate of the species' optimum for each predictor. Niche widths were calculated as the 532 range between the weighted 10^{th} and 90^{th} quantiles and estimate the relative tolerance range of 533 534 the species to the corresponding predictor. The semi-transparent ellipses indicate two times 535 the value of the standard errors associated with the mean PC scores of the FGs.

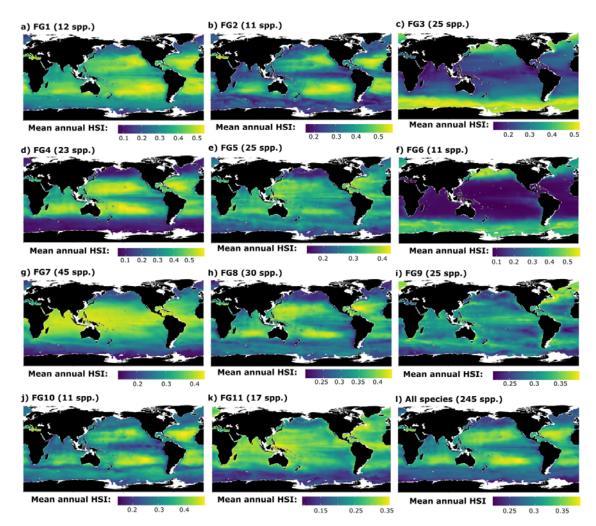
536

537 Post-hoc variance analyses showed significant (Dunn's tests; p < 0.05) inter-FGs variations in 538 niche characteristics and PC scores (see Appendix S14). The largest distance in niche space 539 (PC1) was found between FG4 and FG6 (Fig. 2a; p = 6.5e-10). Along PC1, FG4 also differed 540 significantly from FG1 (p = 0.042), FG3 (p = 1.6e-06), FG7 (p = 1.3e-08), FG8 (p = 1.4e-4), 541 FG9 (p = 2.3e-06), FG10 (p = 0.011) and FG11 (p = 0.005). Conversely, FG6 differed 542 significantly from FG2 (p = 4.5e-4), FG5 (p = 0.001) and FG8 (p = 0.049) along PC1. FG6 543 also differed significantly from FG2 (p = 0.001) and FG5 (p = 0.002). None of the FGs did 544 showed significant variations along PC2. Only FG3 and FG7 showed significant variations 545 along PC3 (p = 4.2e-3; Appendices S13 and S14). To summarize, when FGs 4 and 6 were not 546 accounted for, none of the remaining nine FGs showed significant variations in PC1 and PC2 547 scores (all p > 0.05). This was due to the high level of inter-FG environmental niche 548 characteristics overlap, which implies that distinct FGs display broad environmental niches 549 and can share similar abiotic habitats.

550

551 3.3. Mean annual habitat suitability indices (HSI) patterns

552 For all three SDMs and 11 FGs combined, we found SST to be the most important predictor for constraining the models, followed by PAR, logNO₃, MLD, logChl, logEKE, Si^{*}, N^{*}, 553 554 Wind, and pCO₂ (Appendix S6). SST remained the most important predictor for every FG 555 when looking at predictor importance per FG. Therefore, most FGs displayed mean annual 556 HSI patterns that were driven by the latitudinal temperature gradients at the first-order. The 557 impact of second-order predictors became clearer on a sub-global scale. Aggregating the 558 average monthly HSI projected across SDMs per FG allowed us to estimate their mean annual 559 habitat suitability patterns (Fig. 3).



561

562 Figure 3: Maps of mean annual habitat suitability index (HSI) the eleven copepod functional groups 563 (FG) identified in the present study: a) FG1 (small myelinated cruise-feeding omnivores-herbivores), 564 b) FG2 (small non myelinated cruise-feeding detritivores), c) FG3 (medium-sized myelinated mixed-565 feeding or cuise-feeding detritivores), d) FG4 (small amyelinated ambush-feeding carnivores), e) FG5 566 (large current- or cruise-feeding carnivores), f) FG6 (very large myelinated current-feeding 567 omnivores-herbivores), g) FG7 (medium-sized myelinated current-feeding omnivores-herbivores), h) 568 FG8 (medium-sized amyelinated ambush- or current-feeding carnivores), i) FG9 (medium-sized 569 amyelinated current-feeding omnivores), j) FG10 (small amyelinated ambush-feeding omnivores), k) 570 FG11 (small amyelinated mixed-feeding omnivores) and for l) all species together. Mean annual 571 estimates were derived from the 12 monthly estimates of mean HSI obtained for each of the three 572 species distribution models (SDMs) used (generalized linear models, generalized additive models and 573 artificial neural networks). The FG were defined based on the functional traits combinations of the 245 574 copepod species described by a factorial analysis on mixed data (FAMD) whose principal components 575 were used to perform hierarchical clustering on a Euclidean distance matrix with Ward's aggregation 576 link.

577

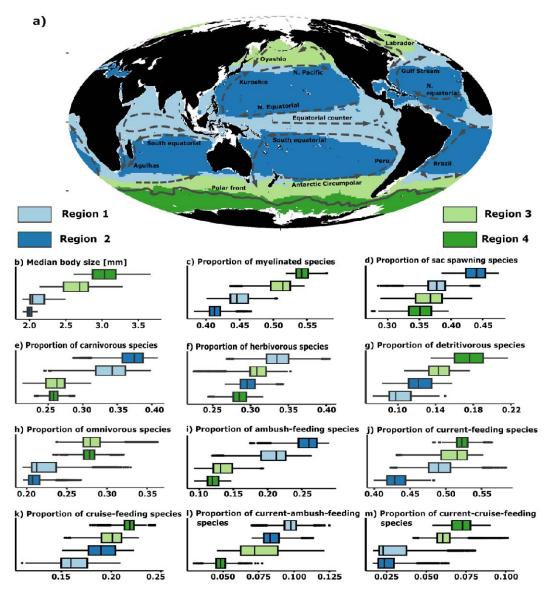
578 The mean annual HSI of FG3 was found to be maximal towards the poles and decrease 579 progressively towards the equator. In contrast, the mean annual HSI of FG1, FG2, FG4, FG8, 580 and FG10 were found to be highest in the tropics and decreased towards higher latitudes. For 581 those five FGs, slighter decreases in HSI were modelled towards the tropical upwelling

582 systems (e.g., Peru, Benguela), the northern part of the Indian Ocean and the Pacific 583 Equatorial counter current. Therefore, these five FGs reached maximal HSI in the 584 oligotrophic conditions of the tropical gyres. Meanwhile, FG5, FG9, and FG11 show less 585 marked latitudinal gradients in mean annual HSI values. No clear hotspot in HSI could be 586 found for these three FGs, but they all displayed lower HSI in higher latitudes than in the 587 tropics. Yet, the region where their mean annual HIS was lowest varied between groups: the 588 North Pacific and parts of the Southern Ocean for FG5 and FG9, the whole Southern Ocean 589 for FG11. The mean annual HSI pattern of FG7 followed the annual SST gradient very 590 closely, with maximal HSI values near the equator and progressive decrease towards the 591 poles.

592

593 3.4. Functional trait biogeography from CMW trait values and regionalization

594 We projected the CWM body size and the CWM values of trophic group, feeding mode, 595 spawning mode, and myelination for the global open ocean on a mean annual scale to 596 illustrate the biogeography of these key traits (Appendix S9). To summarize the main spatial 597 gradients of trait biogeography and assess how these could overlap with known large scale 598 oceanographic currents (i.e., currents and fronts), the global ocean was clustered based on 599 similarities in CWM values after reducing their dimensionality (see Appendices S10 and 600 S11). We identified four ocean regions that display significant contrasts in CWM trait 601 proportions (Fig. 4; Appendix S15).



603

604 Figure 4: The global ocean divided into a) four regions according to the principal components of a 605 Principal Component Analysis (PCA) based on the CWM values of the functional traits. Major 606 oceanographic circulation features (i.e., surface currents and fronts) are illustrated through dashed 607 arrows to highlight their overlap between the regions' boundaries. The distribution of CWM trait 608 values between the four regions are shown through the boxplots: b) median body size, CWM values of 609 c) myelinated species, d) sac spawning species, e) carnivores, f) herbivores, g) detritivores, h) 610 omnivores, i) ambush-feeders, j) current-feeders, k) cruise-feeders, l) current-ambush-feeders, and m) 611 current-cruise-feeders. The position of the Antarctic Polar Front was taken from Orsi and Harris 612 (2019). The positions of the main large-scale ocean currents were drawn according to Pidwirny 613 (2006). The lower, middle, and upper boundaries of the boxplots correspond to the 25th, 50th, and 614 75th percentiles respectively. The lower and upper whiskers extend no further than 1.5*IQR 615 (interquartile range) from the lower and upper hinges.

616

Region 1 primarily fell within the equatorial band and comprised coastal upwelling regionsand the main oxygen minimum zones. The copepod communities of Region 1 displayed lower

- 619 CWM body size (median \pm IQR = 2.043 mm \pm 0.194), higher proportions of carnivores
- 620 (0.342 \pm 0.040) and herbivores (0.335 \pm 0.029) but lower proportions of omnivores (0.213 \pm
- 621 0.030) and detritivores (0.098 \pm 0.023). The CWM values of myelinated (0.446 \pm 0.026), sac-
- spawning (0.377 \pm 0.026), and ambush-feeding (0.212 \pm 0.036) species were lower than the
- 623 CWM traits values found for the communities of Region 2 and higher than those found for the
- high latitude communities of Regions 3 and 4 (Appendix S13).
- Region 2 comprised the tropical gyres. We found that communities in Region 2 showed the
- lowest CW median body size $(2.002 \pm 0.057 \text{ mm})$ but the highest CWM values of ambush-
- 627 feeding (0.259 \pm 0.027) carnivorous (0.372 \pm 0.028), and sac-spawning (0.440 \pm 0.030)
- 628 copepods. Conversely, the CWM values of myelinated (0.413 \pm 0.016) species in Region 2
- 629 were the lowest across all four regions.
- 630 Region 3 was located poleward to the previous two regions and covered the North Atlantic
- 631 Ocean, the North Pacific Ocean and the waters located between the Polar Front and the
- Antarctic Circumpolar Current. The CWM body size $(2.691 \pm 0.342 \text{ mm})$, the CWM values
- of myelinated (0.514 \pm 0.028) and omnivorous (0.279 \pm 0.021) species were higher there than
- 634 in Regions 1 and 2. In contrast, the CWM values of sac-spawners (0.368 ± 0.041), carnivores
- 635 (0.266 \pm 0.026), and ambush feeders (0.132 \pm 0.027) was lower.
- Region 4 mainly corresponded to the Southern Ocean (i.e., grid cells south of the Antarctic
- 637 Polar Front, Fig. 4a). Region 4 displayed the highest CWM body size $(3.035 \pm 0.321 \text{ mm})$,
- higher CWM values of myelinated (0.541 \pm 0.018), omnivorous (0.278 \pm 0.013), and
- 639 detritivorous (0.176 \pm 0.030) species, and the lowest CWM values of sac-spawners (0.351 \pm
- 640 0.035). The CWM values of current-feeding (0.522 \pm 0.015) and cruise-feeding (0.218 \pm
- 641 0.009) species was higher than those of ambush-feeding (0.120 ± 0.016).
- 642

643 **4. Discussion**

644 4.1. Towards meaningful global zooplankton FGs in marine ecology

Here, we identified eleven copepod FGs based on combinations of species-level functional traits. Nine of these eleven FGs had also been found, or were nested within larger groups, in previous studies based on regional species pools (Pomerleau et al., 2015; Benedetti, Gasparini, & Ayata, 2016; Benedetti, Vogt, et al., 2018; Becker et al., 2021; summarized in Table 1), which suggests that most of the functions performed by copepods on a global scale should also be expressed at regional scales.

- 652 **Table 1**: Table summarizing the overlap between the copepod functional groups (FGs) defined in the
- present study and those found in previous studies based on species composition and functional trait
- 654 composition.

This study	Main clade	Main traits	Pomerleau et al. (2015) - North Pacific Ocean	Benedetti et al. (2016) - Mediterranean Sea	Benedetti et al. (2018) - Mediterranean Sea	Becker et al. (2021) - South Atlantic Ocean
FG1	Clausocalanus	Small myelinated cruise-feeding omnivores- herbivores	Small current- and cruise- feeding omnivore- herbivores (Group 6)	Small cruise- feeding omnivores- herbivores (subset of Group 6)	Small cruise- feeding omnivores- herbivores (Group 7)	Small and large myelinated cruise- and current-feeding omnivores-herbivores (subset of Group B)
FG2	Oncaeidae	Small non myelinated cruise-feeding detritivores	No equivalent	Small cruise- feeding omnivores- detritivores (subset of Group 6)	Small sac- spawning detritivores (subset of Group 5)	Small amyelinated ambush- or cruise- feeding carnivores or detritivores (small subset of Group D)
FG3	Spinocalanus, Scaphocalanus, Metridia	Medium-sized myelinated mixed-feeding or cuise- feeding detritivores	No equivalent	Small cruise- feeding omnivores- detritivores (subset of Group 6)	Small sac- spawning detritivores (small subset of Group 5)	No equivalent
FG4	Corycaeidae	Small amyelinated ambush- feeding carnivores	No equivalent	Small ambush- feeding carnivores (Group 2)	Small ambush- feeding carnivores (Group 2)	Small amyelinated ambush- or cruise- feeding carnivores or detritivores (small subset of Group D)
FG5	Sapphirinidae, Euchaetidae	Large current- or cruise- feeding carnivores	No equivalent	Large cruise- feeding carnivores (Group 1)	Large cruise- or current-feeding feeding carnivores (Group 1)	Small and large myelinated cruise- and current-feeding omnivores-herbivores (subset of Group B)
FG6	Calanus, Eucalanus	Very large myelinated current-feeding omnivores- herbivores	No equivalent	No equivalent	No equivalent	No equivalent
FG7	Calanus, Paracalanus, Calocalanus	Medium-sized myelinated current-feeding omnivores- herbivores	Small and large current- feeding omnivores- herbivores (Group 5c)	Small and large current-feeding omnivores- herbivores (Group 4)	Small and large current-feeding omnivores- herbivores (Groups 3 and 4)	Large myelinated current-feeding omnivores-herbivores (subset of Group C)
FG8	Candacia, Haloptilus, Heterorhabdus	Medium-sized amyelinated ambush- or current-feeding carnivores	No equivalent	Large cruise- feeding carnivores (Group 1)	Large cruise- or current-feeding feeding carnivores (Group 1)	No equivalent
FG9	Pleuromamma, Gaetanus,	Medium-sized amyelinated	No equivalent	Small and large current-feeding	Small and large current-feeding	No equivalent

	Labidocera	current-feeding omnivores		omnivores- herbivores (subset of Group 4)	omnivores- herbivores (small subset of Group 4)	
FG10	Oithonidae	Small amyelinated ambush- feeding omnivores	No equivalent	Small ambush- feeding omnivores (Group 5)	Small ambush- feeding omnivores (Group 6)	Small amyelinated mixed- or ambush- feeding omnivores (Group A)
FG11	Acartia, Centropages	Small amyelinated mixed-feeding omnivores	Small ambush- feeding omnivores (subset of Group 4)	Small mixed- feeding omnivores (Group 3)	Small mixed- feeding omnivores (small subset of Group 4)	Small amyelinated mixed- or ambush- feeding omnivores (Group A)

655

656

The four FGs defined by Becker et al., (2021) based on a series of cruises in the South Atlantic Ocean were quite broad and often functionally heterogeneous (i.e., large carnivorous copepods mixed with smaller current-feeding omnivorous-herbivorous ones, or small cruisefeeding particles feeders mixed with small ambush-feeding carnivores). As a result, our FGs are often nested within, or scattered across, the groups defined by Becker et al. (2021).

662 Only the present FG3 and FG6 have no true counterparts in previous studies. These two 663 groups find their most suitable habitats towards the poles (Fig. 3) whose zooplankton 664 communities were not covered by the regional studies mentioned. The copepods of FG3 665 (Spinocalanus spp., Metridia spp., or Scaphocalanus spp.) are known to mainly inhabit 666 deeper ocean layers where they feed on falling particulate organic matter and zooplankton 667 carcasses (Yamaguchi et al., 2002; Sano et al., 2013). Therefore, this group contributes to the 668 remineralization of organic matter and marine snow at higher latitudes and/or in colder 669 conditions. FG6 comprises the largest current-feeding omnivorous-herbivorous copepods 670 from the Calanidae family (Eucalanus spp. and Calanus spp.). Such large copepods are not 671 present in the warm and oligotrophic conditions of the Mediterranean Sea hence their absence 672 in Benedetti, Gasparini, & Ayata (2016) and Benedetti, Vogt, et al. (2018). FG6 is a key 673 group for the biological carbon pump and lipid pump (Jónasdóttir et al., 2015; Visser et al., 674 2017; Steinberg & Landry, 2017; Brun et al., 2019) as it represents large-bodied grazers that 675 can actively feed on microphytoplankton, perform relatively strong vertical migrations and 676 generate large and fast sinking pellets (Stamieszkin et al., 2015; Ohman & Romagnan 2016; 677 Brun et al., 2019). This is confirmed by the position of FG6 in niche space (Fig. 2) as it is 678 affiliated to turbulent and seasonally varying conditions with higher nutrient and chlorophyll-

679 a concentrations. The ecological roles and functions ensured by the other nine FGs have 680 already been detailed and discussed in those previous studies. We support the statements of 681 these authors and how they described the ecological roles of the copepod FGs. Here, we 682 found a larger range of FGs compared to previous studies due to three main reasons: (i) we 683 investigated a global and larger (hundreds of species instead of tens) pool of species, (ii) we 684 accounted for myelination as an additional trait, and (iii) we retained body size as a 685 continuous trait, contrary to Benedetti, Gasparini & Ayata (2016) and Benedetti, Vogt et al. 686 (2018) whom relied on size classes which likely smoothed out important trait variations.

687 Overall, the largest differences in niche space (Fig. 2a, Appendix S14) occurred between two 688 main sets of FGs: (i) myelinated free-spawning large-bodied, or medium-bodied, current-689 feeding herbivores (FG3, 6, 7 and 9), and (ii) amyelinated sac-spawning small-bodied, or large-bodied, cruise- and ambush-feeding detritivores and carnivores (FG2, 4 and 5). The first 690 691 set is associated with conditions of stronger mixing and higher nutrients and chlorophyll-a 692 concentrations (Fig. 2b). The species constituting these FGs also show larger niche widths, 693 suggesting they display broader tolerances to monthly variations in environmental conditions. 694 Therefore, these FGs are found more frequently in high latitude environments, or boundary 695 current systems, where either seasonality or horizontal and vertical mixing lead to higher 696 mean annual productivity (Sarmiento & Gruber, 2006; Roy 2018). In contrast, the FGs of the 697 second set display narrower niches and are associated with conditions typical of the warmer tropical oligotrophic gyres (weaker water mixing and lower nutrients and chlorophyll-a 698 699 concentrations; Fig. 2b). Meanwhile, the groups that are in the center of the niche space (FG1, 700 8, 10 and 11) are those whose species are very scattered across said space, meaning these FGs 701 could not be associated to any particular environment at the scale of our study.

702 This continuum of FGs in niche space was also found in functional trait space. Indeed, the 703 FGs of sets (i) and (ii) were often found on opposite sides of the FAMD dimensions (see 704 Appendix S12). The global spatial distribution of the FGs and CWM trait values (Figs. 3 and 705 4) further support the statements above: larger myelinated free-spawning and active feeding 706 copepods occur more frequently near the poles, whereas smaller amyelinated sac-spawning 707 and passive feeding copepods tend to occur in tropical oligotrophic gyres. Similar continuums 708 of copepod functional traits and environmental niches were found at smaller scales 709 (references in Table 1) or globally (Brun et al., 2016).

710

711 Towards meaningful zooplankton FGs in marine ecosystem models

712 Our combination of analyses supports the view that planktonic copepods display a continuum 713 of functional traits with a strong latitudinal gradient (Fig. 3 and 4; Appendix S9), driven by 714 global gradients in abiotic conditions. Larger and myelinated active feeders present higher 715 metabolic rates as well as higher feeding, excretion and mortality rates (Kiørboe 2011a; Lenz 716 2012; Litchman et al., 2013; Kiørboe & Hirst, 2014; Stamieszkin et al., 2015; van Someren 717 Gréve, Almeda, & Kiørboe, 2017; Brun et al., 2019). A similar continuum can occur within a 718 trophic group: larger current-feeding phytoplankton grazers (e.g., FG6 and 7) can feed on 719 larger cells than their smaller congeners (FG1 and 11; Hansen et al., 1994), or larger active-720 feeding predators (FG5 and 8) can exert a top-down control on smaller ambush-feeding 721 predators (FG4) or smaller zooplankton in general. Consequently, such a continuum should be 722 explicitly represented in regional and global marine ecosystem models (e.g. similar to Serra-723 Pompei et al., 2020), which too often rely on a few size classes only (Le Quéré et al., 2005; 724 Sailley et al., 2013). However, with a few exceptions (Henson et al., 2021), global marine 725 ecosystem models do not yet routinely include >10 zooplankton FGs due to computational 726 constraints. Since these models aim to represent biomass dynamics (in units of carbon) 727 between food web components and their interactions with climate, their improvement should 728 first focus on including copepod FGs that contribute the most to community biomass, and 729 biogeochemical function (Le Quéré et al. 2005).

730 To try to rank the present FGs based on their contribution to total copepod community 731 abundance in the upper layers of the global ocean, we implemented a preliminary synthesis of 732 copepod abundance observations from various large-scale data sources (see Appendix S16). 733 When examining the relative contribution of the eleven FGs to mean annual community 734 abundance (Fig. S16b), three FGs emerge as the most abundant regardless of latitude and 735 sampling gears: FG10 (Oithonids; small ambush-feeding omnivores), FG2 (Oncaeids; small 736 cruise-feeding detritivores) and FG1 (Clausocalanus spp.; small cruise-feeding omnivores-737 herbivores). FG11 (Acartia spp. and Centropages spp.; small omnivorous mixed-feeders) also 738 contributes substantially to total copepod abundance in the northern Atlantic and Pacific 739 Oceans. Consequently, we encourage modelers to represent those three FGs more explicitly in 740 future global ecosystem models. Some additional groups show a higher relative abundance on 741 a regional scale, like FG4 in the tropics, or FG7 and FG9 in the Southern Ocean, which may 742 make these groups important for regional ecosystem modeling efforts. The remaining FGs 743 (FG3, 5, 6 and 8) display very low contributions to mean annual abundance, regardless of the 744 latitude and the sampling gear. Such groups might be of a lower priority for improving the 745 representation of zooplankton in marine ecosystem models. Nonetheless, these four FGs are

characterized by larger body sizes than FG1, 2 or 10 (Table S4.2) so their actual contribution
to biomass production could be substantially higher, especially FG6 which gathers the largest
filter-feeding Calanidae that are several millimeters longer than the species in FG1, FG2 or
FG10. Future work will help model the contribution of the present copepod FGs to
zooplankton biomass, and to clarify their relative priorities for inclusion in marine ecosystem
models.

752

753 4.2. Why do copepod functional traits show contrasting biogeographic patterns in the ocean? 754 We defined four main ocean regions according to their similarity in community-level 755 functional trait expression (Fig. 4a, Appendices S9 and S15). The first-order pattern 756 corresponded to the separation between those regions poleward to $\sim 45^{\circ}$ latitude (regions 3 757 and 4) and those equatorward to $\sim 45^{\circ}$ latitude (regions 1 and 2; Appendix S10). Then, two 758 key second-order patterns emerged: the regions south of the Antarctic Polar Front (region 4) 759 showed a community trait composition distinct than those of the other high latitudes (e.g., the 760 Arctic Ocean or the North Pacific Ocean), and the oligotrophic tropical gyres (region 2) were 761 separated from the boundary upwelling and the equatorial current systems (region 1). The 762 largest dissimilarity in community trait expression was found between regions 2 and 4 (Fig. 763 4b-m) and was driven by differences in CWM body size, myelination, carnivory and ambush-764 feeding vs. current-feeding (Appendices S9, S10 and S15). These trait patterns emerge from environment-based distribution models, so they reflect combinations of environmental 765 766 conditions, but they also reflect biological processes that lead to the selection of trade-offs 767 between traits through abiotic and/or biotic filtering. In other words, certain trait combinations 768 are more competitive than others under varying conditions of temperature or food availability 769 because of physiological constraints or because they lead to lower mortality rates (Litchman 770 et al., 2013; Barton et al., 2013; van Someren Gréve et al., 2017; McGinty et al., 2021).

771 Median copepod body size decreased from the poles to the equator with a slight increase in 772 upwelling systems. Such a pattern has been documented by other studies and is primarily 773 driven by the strong negative relationship between the body size of marine ectotherms and 774 temperature, according to Bergmann's rule (Brun et al., 2016; McGinty et al., 2018; Evans et 775 al., 2020; Brandão et al., 2021; Campbell et al., 2021). The precise processes underlying 776 Bergmann's rule remain debated, but it is likely that warmer temperatures (or a factor 777 confounded with temperature) decrease growth efficiency and/or promote the maturation of adults at smaller body sizes (Atkinson 1994; Isla et al., 2008). Dissolved oxygen 778 779 concentration could also limit maximal body size in marine ectotherms under the "oxygen

hypothesis" (Audzijonyte et al., 2019), but so far temperature outperformed oxygen in
explaining variations in body size structure (Campbell et al., 2021).

782 Bergmann's rule occurs at the intraspecific level and at the interspecific level, meaning that 783 warming-induced decreases in median body size can emerge if smaller clades replace larger 784 ones along a latitudinal gradient. Such a turnover in size classes has been observed as well 785 (Evans et al., 2020; Brandão et al., 2021) and is supported by our SDM projections. Indeed, 786 functional traits that represent shifts in clade composition (i.e., those that show strong 787 taxonomic clustering) such as myelination, spawning strategy, carnivory or feeding modes 788 displayed latitudinal gradients that are positively, or negatively, collinear with the body size 789 gradient (Appendices S9 and S10). The latitudinal patterns of these functional traits likely 790 result from changes in food availability, quality, and predation pressure that are known to 791 vary greatly from polar regions to tropical gyres (Brun et al., 2016; Horne et al., 2016; van 792 Someren Gréve et al., 2017; Roy, 2018).

793 Lipid-rich myelin sheaths enable a faster conduction of nerve responses. This promotes faster 794 reaction times and thus more efficient feeding or escape behaviors (Lenz, 2012). Myelin 795 sheaths are cholesterol-rich so they require larger metabolic investments of dietary lipids 796 (Lenz, 2012), which helps explain why copepod communities show larger proportion of small 797 amyelinated taxa in tropical gyres where smaller lipid-poor phytoplankton dominate (Roy, 798 2018). Therefore, it is not surprising that the proportions of myelinated species follow the 799 same spatial pattern as body size and current-feeding (Fig. 4b,c,j,m) and peak in productive 800 environments characterized by larger and lipid-rich plankton (Roy 2018).

801 Similarly, sac-spawning is a more energy-conservative spawning strategy as it reduces egg-802 mortality at the cost of fecundity and hatching speed. The increased proportions of sac-803 spawners in tropical gyres (Fig. 4d) could reflect an adaptation to limited food availability 804 (Kiørboe & Sabatini, 1994; Barton et al., 2013) and higher rates of carnivory (Fig. 4e; 805 Woodd-Walker et al., 2002) and egg cannibalism among copepods (Ohman & Hirche, 2001; 806 Segers & Taborsky, 2011). Furthermore, ambush-feeding is a passive feeding mode that 807 lowers predation risk and energy costs compared to current- and cruise-feeding, but at the 808 expense of feeding efficiency (Kiørboe, 2011a; van Someren Gréve et al., 2017). 809 Consequently, the increased proportion of ambush-feeders in region 2 (Fig. 4i) should also 810 result from trade-offs in functional traits expression driven by abiotic and biotic filtering 811 (Litchman et al., 2013), as oligotrophic gyres seem to promote food-webs with increased 812 carnivorous predation (Fig. 4e) and resource competition (Wood-Walker et al., 2002; Prowe, 813 Visser, Andersen, Chiba, & Kiørboe, 2019).

814 However, our approach is based on presence data and habitat suitability indices rather than 815 abundances, which likely underestimates the contribution of very abundant ambush-feeding 816 species like *Oithona similis* at high latitudes (Gallienne & Robins, 2001; Pinkerton et al., 817 2010; Prowe, Visser, Andersen, Chiba, & Kiørboe, 2019). As a result, we probably 818 underestimate the proportion of ambush-feeders in regions such as the Southern Ocean 819 compared to Prowe et al. (2018), although these authors discarded other ambush-feeding 820 copepods such as the Corycaeidae (Benedetti et al., 2016; Brun et al., 2017). Together, these 821 elements support the fact that our modelled zooplankton functional traits patterns emerge 822 from interactions between environmental conditions and the relative fitness (i.e., trade-offs) 823 resulting from different trait combinations.

824 The climatological conditions used to train the SDMs are influenced by large scale oceanic 825 currents as highlighted by our regionalization analysis (Fig. 4). We found the boundaries of 826 the four main regions to overlap with the trajectories of well-known boundary currents as well 827 as equatorial counter currents and fronts. Unfortunately, we cannot disentangle the effects of 828 dispersal by currents on the observed community trait expression (and their modelled spatial 829 patterns) from the effects of gradients in temperature and/or productivity, which select species 830 based on their physiological requirements. Although ocean basin connectivity is rather high at 831 the scale of our study (Jönsson & Watson, 2016), dispersal by marine currents and mesoscale 832 processes are known to impact plankton community structure (Richter et al., 2020; Sommeria-833 Klein et al., 2021). For instance, the Antarctic Polar Front seems to be the main delimiter 834 between regions 3 and 4, which coincides with the view that it imposes a strong physical 835 barrier on passively drifting zooplankton (Murphy et al., 2021). Yet, recent evidence showed 836 that many epipelagic plankton can cross Southern Ocean fronts thanks to the very dynamic 837 meandering eddies (Murphy et al., 2021), which could explain the deviations of the region 4 838 boundaries from the Antarctic Polar Front. Therefore, the gradient in community trait 839 expression separating regions 3 and 4 may also stem from the latitudinal gradient in 840 temperature and biogeochemical conditions that select species and their traits based on 841 physiological and metabolic constraints. Similarly, the gradients in CWM traits (i.e., the 842 poleward succession of regions 2, 1 and 3) modelled in the North Pacific and North Atlantic 843 Oceans could be driven by the temperature gradient or the presence of western boundary 844 currents (e.g., the Gulf Stream and the North pacific Current) which carry southern warm-845 water communities eastward while they are getting mixed with northern cold-water 846 communities. Such processes could explain why the CWM trait distribution of region 1 lies in 847 between those of regions 2 and 3 (Fig. 4). This uncertainty calls for more studies combining

distribution models with dispersal rates and limitations (D'Amen et al., 2018; Shipley et al.,

849 2021) to study the link between zooplankton functional traits and the "seascape" (Sommeria-

851

4.3. Caveats and future directions

853 Our results are sensitive to the three main steps of our framework: (i) the way the species 854 were positioned in a functional space and how the FGs were defined from the latter, (ii) the 855 quantity and quality of the functional trait data considered, and (iii) the SDMs chosen to 856 generate the CWM traits values. We carefully investigated the sensitivity of our FGs 857 definition to alternative choices made in our clustering approach (Appendix S4). Most of the 858 comparable studies working with functional trait spaces rely on a Gower distance matrix and 859 a principal coordinate analysis (PCoA; Legendre & Legendre, 2012) to ordinate taxa as a 860 function of their trait combinations (see the synthesis by Mouillot et al., 2021). Here, we 861 relied on an alternative dimensionality reduction analysis (i.e., the FAMD) to avoid negative 862 eigenvalues (i.e., imaginary dimensions) and the relatively low levels of explained variance 863 that are often inherent to the use of a PCoA (Legendre & Legendre, 2012; Mouillot et al., 864 2021). We assessed how using this alternative dimension reduction analysis affected the 865 quality of the functional trait space by computing the AUC criterion recommended by 866 Mouillot et al. (2021). We found an AUC value of 0.81, which is substantially higher than the recommended 0.7 threshold and indicates a "high quality trait space" according to these 867 868 authors. In addition, the first four PCs of our FAMD explained nearly 80% of the variance in 869 species traits, which reinforces our confidence in the quality of our functional trait space.

870 Beyond the dimensionality reduction step, the choice of the distance matrix and aggregation 871 link could have been determining factors in drawing the functional dendrogram and thus 872 defining the FGs (Fig. 1). We evaluated the similarity between the functional dendrograms 873 emerging from alternative clustering approaches and found that they were all highly 874 positively correlated (mean Baker's Gamma correlation coefficient was 0.75 ± 0.12 ; 875 Appendix S4). This indicates that all dendrograms displayed very similar structure and 876 explains how these dendrograms would have led to similar FGs composition. Ultimately, the 877 approach that combined the FAMD, an Euclidean distance matrix, and Ward's aggregation 878 linkage was chosen as our standard approach, since it provided ecologically meaningful FGs 879 (i.e., groups that are neither too large and functionally heterogeneous nor too numerous and 880 functionally redundant).

⁸⁵⁰ Klein et al., 2021; Richter et al., 2020).

881 The quantity and the quality of the species-level functional trait information considered for 882 our study largely determined the FGs and their ecological meaning. The traits chosen for this 883 study only cover a fraction of the traits mentioned in the literature (Litchman et al., 2013; 884 Brun et al., 2017; Appendix S3). Therefore, our study is likely underestimating the true 885 diversity of copepod functions and FGs present in the ocean. However, it could also be that 886 the present functional groups are quite representative of natural copepod communities and 887 that adding further traits would only subdivide those FGs into smaller and functionally 888 homogeneous ones. The trait compilation of Brun et al. (2017) included the following 889 additional functional traits: growth rate, clearance rate, ingestion rate, egg diameter and 890 production, or the production of resting eggs and diapausing stages throughout the life cycle. 891 Yet, the availability and the coverage of those traits remained too poor for the wide pool of 892 species studied here (Appendix S3), reflecting the historic measurement biases that largely 893 focused on larger Calanoida species such as *Calanus* spp. Since a broad range of metabolic 894 and physiological traits scale allometrically with body size (Kiørboe & Hirst, 2014), 895 additional traits could be inferred based on body size measurements and phylogenetic 896 distances (Molina-Velegas et al., 2018). It also implies that the present FGs characterized by 897 larger body sizes (e.g., FG5 and 6) should display larger metabolic rates. Further lab-based and observation-based are needed to uncover the functional trait expression for a larger range 898 899 of copepod species and families.

900 Differences between model projections are the main source of uncertainty in species 901 community composition and diversity under a SDMs ensemble approach (Benedetti, 902 Guilhaumon, Adloff, & Ayata, 2018; Diniz-Filho et al., 2009). We found some regional 903 differences between the mean annual HSI projections of the SDMs. When considering all 904 copepod species together, the GLM estimated higher mean HSI than the two other model 905 types for the Southern Ocean (Appendix S8). Such variability could arise from how the 906 various SDMs cope with limited predictors and species occurrence availability in winter 907 conditions at the very high latitudes. Indeed, the copepod occurrence data (Appendix S1) and 908 the climatological satellite observations (e.g., PAR and logChl) have poor coverage at 909 latitudes $>60^\circ$ in winter. The sampling effort in the Southern Ocean is slightly lower than the 910 global average between April and October and is extremely low between May and September 911 (Appendix S1), and satellite-based PAR and chlorophyll-a measurements are not available for 912 winter months due to the low light levels at high latitudes. Therefore, our mean annual 913 projections for the polar oceans are likely more representative of summer conditions. Most 914 GLMs and GAMs included one or both of these satellite-based climatologies as predictors

915 (Appendix S6). Yet only GLMs showed these relatively high HSI levels in the Southern 916 Ocean. Plus, SST was found to be the most influential predictor in both SDMs types 917 (Appendix S6). This implies that this discrepancy between GLMs and GAMs was not driven 918 by predictor selection but rather by how the responses of the species to SST and logChla/PAR 919 gradients were captured by those two types of models (Merow et al., 2014). Here, the less 920 complex response curves of the GLMs lead to higher average HSI at very cold temperatures 921 compared to GAMs. The HSI patterns obtained from the GAM and ANN were closer to the 922 latitudinal diversity gradients that were previously observed and modelled for marine 923 ectotherms, which seldom show increases in richness or habitat suitability towards the poles 924 (Tittensor et al., 2010; Benedetti et al., 2021). Consequently, we have more confidence in the 925 GAM-based and ANN-based annual HSI estimates that in the GLM-based ones.

926 Interestingly, inter-SDMs variability was much lower when looking at CWM traits 927 projections rather than HSI patterns (Appendix S8). All three SDMs lead to very similar 928 spatial patterns in CWM traits and the amplitude of these CWM values differed only slightly. 929 The regions displaying higher CWM traits variability depended on the trait considered, and 930 variability remained quite low overall (but see Appendix S8). Therefore, the present CWM 931 traits patterns display relatively low uncertainty compared to the HSI patterns. The fact that 932 CWM body size and CWM myelination show very similar spatial patterns and ranges to the 933 respective estimates of Brun et al. (2016) gives us further confidence in our approach and 934 spatial projections.

935 To conclude, we recommend the inclusion of multiple functional groups of copepods in 936 marine ecosystem models, based on our functional dendrogram (Fig. 1), with a level of 937 complexity of zooplankton representation that should be modulated by the scientific question, 938 the FGs dominating community biomass at the scale of the study region(s), the traits of 939 interest, and computing efficiency constraints. Ideally, future global marine ecosystems, that 940 cannot efficiently include 10 zooplankton groups, will include a few (3-4) FGs that cover the 941 main gradients observed in trait space (Fig. 1; Appendix S12), niche space (Fig. 2), 942 geographical space (Figs. 3, 4; Appendices S9, S10) and that represent important fractions of 943 biomass (Appendix S16). This would divide copepods into a few but ecologically meaningful 944 groups with clearly distinct ecological niches and also functional trait characteristics. For 945 other ecological applications or to represent food-web interactions across trophic levels (Ward 946 et al., 2012; Serra-Pompei et al., 2020), these groups could be split into further FGs with 947 specific characteristics (e.g., size classes, high lipid content, grazing and mortality rates etc.). 948 To achieve this, future studies are required to improve the coverage of functional traits data

949 towards more taxonomic groups and more quantitative traits (Appendix S3). Ongoing 950 compilations of zooplankton groups biomass data will also help rank the importance of the 951 present FGs worldwide and across regions. Ultimately, the present results and future data 952 compilations will help us study global patterns of zooplankton functional diversity (Mouillot 953 et al., 2013), which may link more closely to ecosystem function and service provision than 954 taxonomic diversity. Marine ecosystems will experience increasingly stressful conditions in 955 the next century due to anthropogenic climate change (IPCC, 2021). This will strongly 956 restructure zooplankton community richness and composition as species migrate poleward to 957 track suitable habitats (Benedetti et al., 2021). How these changes will impact functional trait 958 expression and functional diversity at the community-level remains too poorly understood. 959 The present study is thus a key step towards our understanding of how future climate change 960 will reshape the expression of marine functional diversity worldwide.

961

962 Data Availability Statement

The copepod species occurrences data used to train the species distribution models are publicly available on Zenodo (10.5281/zenodo.5101349). The newly implemented specieslevel functional trait table is available as Table S2 and will be made available through an open access repository upon publication of our study. Any computer code used to generate the results of the study are freely available upon request to the authors and all R codes are currently stored on the GitHub account of J.W. (https://github.com/jonas-wydler).

969

970 References

Aiello-Lammens, M. E., Boria, R. A., Radosavljevic, A., Vilela, B., & Anderson, R. P. (2015).
spthin: an R package for spatial thinning of species occurrence records for use in ecological niche
models. *Ecography*, 38(5), 541–545.

Atkinson, D. (1994). Temperature and Organism Size—A Biological Law for Ectotherms? In
M. Begon & A. H. Fitter (Eds.), *Advances in Ecological Research* (Vol. 25, pp. 1-58): Academic
Press.

Audigier, V., Husson, F., & Josse, J. (2016). A principal component method to impute missing
values for mixed data. *Advances in Data Analysis and Classification*, 10(1), 5–26.

Audzijonyte, A., Barneche, D. R., Baudron, A. R., Belmaker, J., Clark, T. D., Marshall, C. T.,
van Rijn, I. (2019). Is oxygen limitation in warming waters a valid mechanism to explain
decreased body sizes in aquatic ectotherms? *Global Ecology and Biogeography*, 28(2), 64-77.
doi:10.1111/geb.12847

983	Baker, F. B. (1974). Stability of two hierarchical grouping techniques case i: sensitivity to data
984	errors. Journal of the American Statistical Association, 69(346), 440-445.
985	Barnett, A. J., Finlay, K., & Beisner, B. E. (2007). Functional diversity of crustacean
986	zooplankton communities: towards a trait-based classification. Freshwater Biology, 52(5), 796-813.
987	Barton, A. D., Pershing, A. J., Litchman, E., Record, N. R., Edwards, K. F., Finkel, Z. V.,
988	Ward, B. A. (2013). The biogeography of marine plankton traits. Ecology Letters, 16(4), 522-534.
989	doi: 10.1111/ele.12063
990	Beaugrand, G., Edwards, M., & Legendre, L. (2010). Marine biodiversity, ecosystem
991	functioning, and carbon cycles. Proceedings of the National Academy of Sciences, 107(22), 10120-
992	10124. doi:10.1073/pnas.0913855107
993	Becker, É. C., Mazzocchi, M. G., de Macedo-Soares, L. C. P., Costa Brandão, M., &
994	Santarosa Freire, A. (2021). Latitudinal gradient of copepod functional diversity in the South Atlantic
995	Ocean. Progress in Oceanography, 199, 102710. doi:https://doi.org/10.1016/j.pocean.2021.102710
996	Benedetti, F., Gasparini, S., & Ayata, SD. (2016). Identifying copepod functional groups
997	from species functional traits. Journal of Plankton Research, 38(1), 159-166.
998	doi:10.1093/plankt/fbv096
999	Benedetti, F., Guilhaumon, F., Adloff, F., & Ayata, SD. (2018). Investigating uncertainties in
1000	zooplankton composition shifts under climate change scenarios in the Mediterranean Sea. Ecography,
1001	41(2), 345–360. doi: 10.1111/ecog.02434
1002	Benedetti, F., Vogt, M., Righetti, D., Guilhaumon, F., & Ayata, SD. (2018). Do functional
1003	groups of planktonic copepods differ in their ecological niches? Journal of Biogeography, 45(3), 604-
1004	616. doi: 10.1111/jbi.13166
1005	Benedetti, F., Vogt, M., Elizondo, U. H., Righetti, D., Zimmermann, N. E., & Gruber, N.
1006	(2021). Major restructuring of marine plankton assemblages under global warming. Nature
1007	Communications, 12(1), 5226. doi:10.1038/s41467-021-25385-x
1008	Boyer, T.P., J. I. Antonov, O. K. Baranova, C. Coleman, H. E. Garcia, A. Grodsky, D. R.
1009	Johnson, R. A. Locarnini, A. V. Mishonov, T.D. O'Brien, C.R. Paver, J.R. Reagan, D. Seidov, I. V.
1010	Smolyar, and M. M. Zweng, 2013: World Ocean Database 2013, NOAA Atlas NESDIS 72, S.
1011	Levitus, Ed., A. Mishonov, Technical Ed.; Silver Spring, MD, 209 pp.,
1012	http://doi.org/10.7289/V5NZ85MT
1013	Bron, J. E., Frisch, D., Goetze, E., Johnson, S. C., Lee, C. E., & Wyngaard, G. A. (2011).
1014	Observing copepods through a genomic lens. Frontiers in Zoology, 8(1), 1-15.
1015	https://doi.org/10.1186/1742-9994-8-22
1016	Brun, P., Payne, M. R., & Kiørboe, T. (2016). Trait biogeography of marine copepods-an
1017	analysis across scales. Ecology Letters, 19(12), 1403-1413. doi: 10.1111/ele.12688
1018	Brun, P., Payne, M. R., & Kiørboe, T. (2017). A trait database for marine copepods. Earth
1019	System Science Data, 9(1), 99-113. doi:10.5194/essd-2016-30

1020	Brun, P., Stamieszkin, K., Visser, A. W., Licandro, P., Payne, M. R., & Kiørboe, T. (2019).
1021	Climate change has altered zooplankton-fuelled carbon export in the North Atlantic. Nature Ecology
1022	& Evolution, 3(3), 416-423. https://doi.org/10.1038/s41559-018-0780-3
1023	Campbell, M. D., Schoeman, D. S., Venables, W., Abu-Alhaija, R., Batten, S. D., Chiba, S.,
1024	. Richardson, A. J. (2021). Testing Bergmann's rule in marine copepods. Ecography, 44(9), 1283-
1025	1295. doi:https://doi.org/10.1111/ecog.05545
1026	Cormen, T. H., Leiserson, C. E., Rivest, R. L., & Stein, C. (2009). Introduction to algorithms.
1027	MIT and McGraw.
1028	D'Amen, M., Mod, H. K., Gotelli, N. J., & Guisan, A. (2018). Disentangling biotic
1029	interactions, environmental filters, and dispersal limitation as drivers of species co-occurrence.
1030	Ecography, 41(8), 1233-1244. doi:https://doi.org/10.1111/ecog.03148
1031	Diniz-Filho, J. A. F., Mauricio Bini, L., Fernando Rangel, T., Loyola, R. D., Hof, C., Nogués-
1032	Bravo, D., & Araújo, M. B. (2009). Partitioning and mapping uncertainties in ensembles of forecasts
1033	of species turnover under climate change. Ecography, 32(6), 897-906. doi: 10.1111/j.1600-
1034	0587.2009.06196.x
1035	Dormann, C. F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., others (2013).
1036	Collinearity: a review of methods to deal with it and a simulation study evaluating their performance.
1037	<i>Ecography</i> , 36(1), 27–46. doi: 10.1111/j.1600-0587.2012.07348.x
1038	Flynn, K. J., St John, M., Raven, J. A., Skibinski, D. O. F., Allen, J. I., Mitra, A., & Hofmann,
1039	E. E. (2015). Acclimation, adaptation, traits and trade-offs in plankton functional type models:
1040	reconciling terminology for biology and modelling. Journal of Plankton Research, 37(4), 683-691.
1041	doi:10.1093/plankt/fbv036
1042	Gallienne, C., & Robins, D. (2001). Is Oithona the most important copepod in the world's
1043	oceans? Journal of Plankton Research, 23(12), 1421–1432. https://doi.org/10.1093/plankt/23.12.1421
1044	Guisan, A., Thuiller, W., & Zimmermann, N. E. (2017). Habitat suitability and distribution
1045	models: with applications in R. Cambridge University Press.
1046	Hansen, B., Bjornsen, P. K., & Hansen, P. J. (1994). The size ratio between planktonic
1047	predators and their prey. Limnology & Oceanography, 39(2), 395-403.
1048	Hébert, MP., Beisner, B. E., & Maranger, R. (2017). Linking zooplankton communities to
1049	ecosystem functioning: toward an effect-trait framework. Journal of Plankton Research, 39(1), 3-12.
1050	doi:10.1093/plankt/fbw068
1051	Henson, S. A., Cael, B. B., Allen, S. R., & Dutkiewicz, S. (2021). Future phytoplankton
1052	diversity in a changing climate. Nature Communications, 12(1), 5372. doi:10.1038/s41467-021-
1053	25699-w
1054	Hijmans, R. J. (2012). Cross-validation of species distribution models: removing spatial
1055	sorting bias and calibration with a null model. Ecology, 93(3), 679-688.
1056	doi:https://doi.org/10.1890/11-0826.1

1057	Horne, C. R., Hirst, A. G., Atkinson, D., Neves, A., & Kiørboe, T. (2016). A global synthesis
1058	of seasonal temperature- size responses in copepods. Global Ecology and Biogeography, 25(8), 988-
1059	999. doi: 10.1111/geb.12460
1060	Isla, J. A., Lengfellner, K., & Sommer, U. (2008). Physiological response of the copepod
1061	Pseudocalanus sp. in the Baltic Sea at different thermal scenarios. Global Change Biology, 14(4),
1062	895-906. doi:https://doi.org/10.1111/j.1365-2486.2008.01531.x
1063	IPCC (2021). Climate Change 2021: The Physical Science Basis. Contribution of Working
1064	Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change [Masson-
1065	Delmotte, V., P. Zhai, A. Pirani, S.L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb,
1066	M.I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R. Matthews, T.K. Maycock, T. Waterfield, O.
1067	Yelekçi, R. Yu, and B. Zhou (eds.)]. Cambridge University Press. In Press.
1068	Jónasdóttir, S. H., Visser, A. W., Richardson, K., & Heath, M. R. (2015). Seasonal copepod
1069	lipid pump promotes carbon sequestration in the deep North Atlantic. Proceedings of the National
1070	Academy of Sciences, 112(39), 12122-12126. doi:10.1073/pnas.1512110112
1071	Jönsson, B. F., & Watson, J. R. (2016). The timescales of global surface-ocean connectivity.
1072	Nature Communications, 7, 11239. doi:10.1038/ncomms11239
1073	Kenitz, K. M., Visser, A. W., Mariani, P., & Andersen, K. H. (2017). Seasonal succession in
1074	zooplankton feeding traits reveals trophic trait coupling. Limnology & Oceanography, 62(3), 1184-
1075	1197.
1076	Kiørboe, T. (2011a). What makes pelagic copepods so successful? Journal of Plankton
1077	Research, 33(5), 677-685. doi:10.1093/plankt/fbq159
1078	Kiørboe, T. (2011b). How zooplankton feed: mechanisms, traits and trade-offs. Biological
1079	<i>Reviews</i> , 86(2), 311–339. doi: 10.1111/j.1469-185X.2010.00148.x
1080	Kiørboe, T., & Hirst, A. G. (2014). Shifts in mass scaling of respiration, feeding, and growth
1081	rates across life-form transitions in marine pelagic organisms. The American Naturalist, 183(4), E118-
1082	E130.
1083	Kiørboe, T., & Sabatini, M. (1994). Reproductive and life cycle strategies in egg-carrying
1084	cyclopoid and free-spawning calanoid copepods. Journal of Plankton Research, 16(10), 1353–1366.
1085	Legendre, P., & Legendre, L. (2012). Numerical Ecology. Elsevier.
1086	Le Quéré, C., Harrison, S. P., Colin Prentice, I., Buitenhuis, E. T., Aumont, O., Bopp, L.,
1087	others (2005). Ecosystem dynamics based on plankton functional types for global ocean
1088	biogeochemistry models. Global Change Biology, 11(11), 2016-2040. doi: 10.1111/j.1365-
1089	2486.2005.01004.x
1090	Le Quéré, C., Buitenhuis, E. T., Moriarty, R., Alvain, S., Aumont, O., Bopp, L., Vallina,
1091	S. M. (2016). Role of zooplankton dynamics for Southern Ocean phytoplankton biomass and global
1092	biogeochemical cycles. Biogeosciences, 13(14), 4111-4133. doi:10.5194/bg-13-4111-2016

1093	Lenz, P. H. (2012). The biogeography and ecology of myelin in marine copepods. Journal of
1094	Plankton Research, 34(7), 575-589. doi:10.1093/plankt/fbs037
1095	Litchman, E., Ohman, M. D., & Kiørboe, T. (2013). Trait-based approaches to zooplankton
1096	communities. Journal of Plankton Research, 35(3), 473-484. doi:10.1093/plankt/fbt019
1097	Longhurst, A. R. (2010). Ecological geography of the sea. Elsevier.
1098	McGinty, N., Barton, A. D., Record, N. R., Finkel, Z. V., & Irwin, A. J. (2018). Traits
1099	structure copepod niches in the North Atlantic and Southern Ocean. Marine Ecology Progress Series,
1100	601, 109-126. https://doi.org/10.3354/meps12660
1101	McGinty, N., Barton, A. D., Record, N. R., Finkel, Z. V., Johns, D. G., Stock, C. A., & Irwin,
1102	A. J. (2021). Anthropogenic climate change impacts on copepod trait biogeography. Global Change
1103	Biology, 27(7), 1431-1442. doi:https://doi.org/10.1111/gcb.15499
1104	Merow, C., Smith, M. J., Edwards Jr, T. C., Guisan, A., McMahon, S. M., Normand, S.,
1105	Elith, J. (2014). What do we gain from simplicity versus complexity in species distribution models?
1106	Ecography, 37(12), 1267–1281. doi: 10.1111/ecog.00845
1107	Molina-Venegas, R., Moreno-Saiz, J. C., Castro Parga, I., Davies, T. J., Peres-Neto, P. R., &
1108	Rodríguez, M. Á. (2018). Assessing among-lineage variability in phylogenetic imputation of
1109	functional trait datasets. Ecography, 41(10), 1740-1749. doi:https://doi.org/10.1111/ecog.03480
1110	Mouchet, M., Guilhaumon, F., Villéger, S., Mason, N. W., Tomasini, JA., & Mouillot, D.
1111	(2008). Towards a consensus for calculating dendrogram-based functional diversity indices. Oikos,
1112	117(5), 794–800. doi: 10.1111/j.2008.0030-1299.16594.x
1113	Mouillot, D., Graham, N. A., Villéger, S., Mason, N. W., & Bellwood, D. R. (2013). A
1114	functional approach reveals community responses to disturbances. Trends in Ecology & Evolution,
1115	28(3), 167-177. doi: 10.1016/j.tree.2012.10.004
1116	Mouillot, D., Loiseau, N., Grenié, M., Algar, A. C., Allegra, M., Cadotte, M. W., Auber,
1117	A. (2021). The dimensionality and structure of species trait spaces. <i>Ecology Letters</i> , 24(9), 1988-2009.
1118	doi:https://doi.org/10.1111/ele.13778
1119	Murphy, E. J., Johnston, N. M., Hofmann, E. E., Phillips, R. A., Jackson, J. A., Constable, A.
1120	J., Xavier, J. C. (2021). Global Connectivity of Southern Ocean Ecosystems. Frontiers in Ecology
1121	and Evolution, 9(454). doi:10.3389/fevo.2021.624451
1122	Ohman, M. D., & Hirche, H. J. (2001). Density-dependent mortality in an oceanic copepod
1123	population. Nature, 412(6847), 638-641. doi:10.1038/35088068
1124	Ohman, M. D., & Romagnan, JB. (2016). Nonlinear effects of body size and optical
1125	attenuation on diel vertical migration by zooplankton. Limnology & Oceanography, 61(2), 765-770.
1126	doi: 10.1002/lno.10251
1127	Orsi, A., & Harris, U. (2019). Fronts of the Antarctic circumpolar current-GIS data, ver. 1.
1128	Pagès, J. (2004). Analyse factorielle de données mixtes: principe et exemple d'application.
1129	Montpellier SupAgro, http://www. agro-montpellier. fr/sfds/CD/textes/pages1.pdf

1130 Pecuchet, L., Reygondeau, G., Cheung, W. W. L., Licandro, P., van Denderen, P. D., Payne, 1131 M. R., & Lindegren, M. (2018). Spatial distribution of life-history traits and their response to 1132 environmental gradients across multiple marine taxa. Ecosphere, 9(10), e02460. 1133 doi:https://doi.org/10.1002/ecs2.2460 1134 Phillips, S. J., Dudík, M., Elith, J., Graham, C. H., Lehmann, A., Leathwick, J., & Ferrier, S. 1135 (2009). Sample selection bias and presence-only distribution models: implications for background and 1136 pseudo-absence data. Ecological Applications, 19(1), 181–197. 1137 Pidwirny, M. (2006). Surface and subsurface ocean currents: Ocean current map. 1138 fundamentals of physical edition. Retrieved 2021-08.20, geography. 2nd from 1139 http://www.physicalgeography.net/ fundamentals/8q 1.html 1140 Pinkerton, M. H., Smith, A. N. H., Raymond, B., Hosie, G. W., Sharp, B., Leathwick, J. R., & 1141 Bradford-Grieve, J. M. (2010). Spatial and seasonal distribution of adult Oithona similis in the 1142 Southern Ocean: Predictions using boosted regression trees. Deep Sea Research Part I: 1143 Oceanographic Research Papers, 57(4), 469-485. doi:https://doi.org/10.1016/j.dsr.2009.12.010 1144 Pomerleau, C., Sastri, A. R., & Beisner, B. E. (2015). Evaluation of functional trait diversity 1145 for marine zooplankton communities in the northeast subarctic pacific ocean. Journal of Plankton 1146 Research, 37(4), 712–726. doi:10.1093/plankt/fbv045 1147 Prowe, A. E. F., Visser, A. W., Andersen, K. H., Chiba, S., & Kiørboe, T. (2019). 1148 Biogeography of zooplankton feeding strategy. Limnology & Oceanography, 64(2), 661-678. doi: 1149 10.1002/lno.11067 1150 Qiu, B., & Chen, S. (2004). Seasonal modulations in the eddy field of the south pacific ocean. Journal 1151 of Physical Oceanography, 34(7), 1515–1527. 1152 Revgondeau, G., Guieu, C., Benedetti, F., Irisson, J.-O., Avata, S.-D., Gasparini, S., & 1153 Koubbi, P. (2017). Biogeochemical regions of the Mediterranean Sea: An objective multidimensional 1154 multivariate environmental approach. Progress in Oceanography, and 151, 138-148. 1155 doi:https://doi.org/10.1016/j.pocean.2016.11.001 1156 Richter, D. J., Watteaux, R., Vannier, T., Leconte, J., Frémont, P., Reygondeau, G., . . . Da 1157 Silva, O. (2020). Genomic evidence for global ocean plankton biogeography shaped by large-scale 1158 current systems. BioRxiv, 867739. 1159 Ricotta, C. (2005). A note on functional diversity measures. Basic and Applied Ecology, 6(5), 1160 479-486. 1161 Sailley, S., Vogt, M., Doney, S., Aita, M., Bopp, L., Buitenhuis, E., ... Yamanaka, Y. (2013). Comparing food web structures and dynamics across a suite of global marine ecosystem models. 1162 1163 Ecological Modelling, 261, 43–57. 1164 Segers, F. H. I. D., & Taborsky, B. (2011). Egg size and food abundance interactively affect 1165 juvenile growth and behaviour. **Functional** Ecology, 25(1), 166-176.

1166 doi:https://doi.org/10.1111/j.1365-2435.2010.01790.x

1167	Serra-Pompei, C., Soudijn, F., Visser, A. W., Kiørboe, T., & Andersen, K. H. (2020). A
1168	general size- and trait-based model of plankton communities. Progress in Oceanography, 189,
1169	102473. doi:https://doi.org/10.1016/j.pocean.2020.102473
1170	Shipley, B. R., Bach, R., Do, Y., Strathearn, H., McGuire, J. L., & Dilkina, B. (2021).
1171	megaSDM: integrating dispersal and time-step analyses into species distribution models. Ecography,
1172	44, 1-12. doi:https://doi.org/10.1111/ecog.05450
1173	Sommeria-Klein, G., Watteaux, R., Ibarbalz Federico, M., Pierella Karlusich Juan, J.,
1174	Iudicone, D., Bowler, C., & Morlon, H. (2021). Global drivers of eukaryotic plankton biogeography in
1175	the sunlit ocean. Science, 374(6567), 594-599. doi:10.1126/science.abb3717
1176	Stamieszkin, K., Pershing, A. J., Record, N. R., Pilskaln, C. H., Dam, H. G., & Feinberg, L. R.
1177	(2015). Size as the master trait in modeled copepod fecal pellet carbon flux. Limnology &
1178	Oceanography, 60(6), 2090–2107. doi: 10.1002/lno.10156
1179	Steinberg, D. K., & Landry, M. R. (2017). Zooplankton and the ocean carbon cycle. Annual
1180	Review of Marine Science, 9, 413-444. doi: 10.1146/annurev-marine-010814-015924
1181	Stocker, T. (2014). Climate change 2013: the physical science basis: Working group i
1182	contribution to the fifth assessment report of the intergovernmental panel on climate change.
1183	Cambridge University Press.
1184	Thuiller, W., Georges, D., Engler, R., Georges, M. D., & Breiner F. (2020). biomod2:
1185	Ensemble Platform for Species Distribution Modeling. R package version 3.4.6. https://CRAN.R-
1186	project.org/package=biomod2
1187	Tittensor, D. P., Mora, C., Jetz, W., Lotze, H. K., Ricard, D., Berghe, E. V., & Worm, B.
1188	(2010). Global patterns and predictors of marine biodiversity across taxa. Nature, 466(7310), 1098-
1189	1101. doi:10.1038/nature09329
1190	Turner, J. T. (2004). The importance of small planktonic copepods and their roles in pelagic
1191	marine food webs. Zoological Studies, 43(2), 255–266.
1192	Turner, J. T. (2015). Zooplankton fecal pellets, marine snow, phytodetritus and the ocean's
1193	biological pump. Progress in Oceanography, 130, 205–248.
1194	http://dx.doi.org/10.1016/j.pocean.2014.08.005
1195	van Someren Gréve, H., Almeda, R., & Kiørboe, T. (2017). Motile behavior and predation risk
1196	in planktonic copepods. Limnology & Oceanography, 62(5), 1810-1824. doi: 10.1002/lno.10535
1197	Veloz, S. D. (2009). Spatially autocorrelated sampling falsely inflates measures of accuracy
1198	for presence-only niche models. Journal of Biogeography, 36(12), 2290–2299.
1199	Violle, C., Navas, ML., Vile, D., Kazakou, E., Fortunel, C., Hummel, I., & Garnier, E.
1200	(2007). Let the concept of trait be functional! Oikos, 116(5), 882-892. doi: 10.1111/j.2007.0030-
1201	1299.15559.x
1202	Visser, A. W., Grønning, J., & Jónasdóttir, S. H. (2017). Calanus hyperboreus and the lipid
1203	pump. Limnology & Oceanography, 62(3), 1155-1165. doi:10.1002/lno.10492

Woodd-Walker, R. S., Ward, P., & Clarke, A. (2002). Large-scale patterns in diversity and
community structure of surface water copepods from the Atlantic ocean. *Marine Ecology Progress Series*, 236, 189–203.

1207

1208 Biosketch

1209 Fabio Benedetti is a postdoctoral researcher and Meike Vogt a senior research scientist in 1210 the Environmental Physics (UP) group of ETH Zürich. Both share broad interests in plankton 1211 biogeography and functional diversity and theirs links with biodiversity, ecosystem function 1212 and biogeochemical cycles in the global ocean. F.B. is a macroecologist specialized in trait-1213 based approaches and plankton diversity modelling. M.V. is a marine ecosystem modeler 1214 specialized plankton functional types. Jonas Wydler has successfully completed his MSc 1215 degree in Environmental System Sciences at ETH Zürich, under the supervision of F.B. and 1216 M.V., which is the subject of this work. 1217 F.B. and M.V. co-designed the study and F.B. collated the data used in the analyses and

1218 provided expertise with regard to every methodology used. J.W. conducted the numerical

1219 analyses and wrote the first version of the manuscript under the supervision of F.B. and M.V.

1220 F.B. wrote the final version of the manuscript with input from both M.V. and J.W.

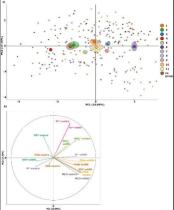




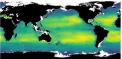








a) FG1 (12 spp.)

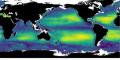


0.1 0.2 0.3 0.4 0.5

0.1 0.2 0.3 0.4 0.5

Mean annual HSI:

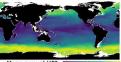
b) FG2 (11 spp.)



Mean annual HSI: 0.2 0.3 0.4

e) FG5 (25 spp.)

c) FG3 (25 spp.)



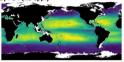
Mean annual HSI: f) FG6 (11 spp.)

0.5

0.4

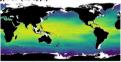
0.2 0.3 0.4

d) FG4 (23 spp.)



Mean annual HSI:

g) FG7 (45 spp.)



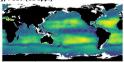
0.2

0.4

0.3 0.4

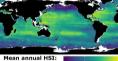
Mean annual HSI:

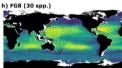
j) FG10 (11 spp.)



Mean annual HSI:

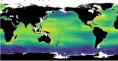
0.2





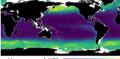
Mean annual HSI: 0.25 0.3 0.35 0.4

k) FG11 (17 spp.)



Mean annual HSI:

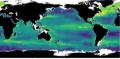
0.25 0.35



Mean annual HSI:

0.1 0.2 0.3 0.4 0.5

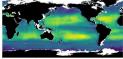




Mean annual HSI:

0.25 0.3

I) All species (245 spp.)



Mean annual HSI

0.25 0.3 0.35

