

1 **Title**

2 Intraspecific diversification and mitonuclear discordance in native versus introduced areas: co-introduction of
3 *Dolicirroplectanum lacustre*, a monogenean gill parasite of the invasive Nile perch *Lates niloticus*

4

5 **Author information**

6 Name: Kelly J. M. Thys

7 Affiliation: Research Group Zoology: Biodiversity and Toxicology, Centre for Environmental Sciences, Hasselt
8 University, Agoralaan Gebouw D, B-3590 Diepenbeek, Belgium

9 e-mail: kelly.thys@uhasselt.be

10 ORCID: 0000-0002-9350-3752

11

12 Name: Maarten P.M. Vanhove

13 Affiliation 1: Research Group Zoology: Biodiversity and Toxicology, Centre for Environmental Sciences, Hasselt
14 University, Agoralaan Gebouw D, B-3590 Diepenbeek, Belgium

15 Affiliation 2: Department of Botany and Zoology, Faculty of Science, Masaryk University, Kotlářská 2, 611 37
16 Brno, Czech Republic

17 e-mail: maarten.vanhove@uhasselt.be

18 ORCID: 0000-0003-3100-7566

19

20 Name: Jonas W. J. Custers

21 Affiliation: Utrecht University, Department of Biology, Padualaan 8, 3584 CH Utrecht, The Netherlands

22 e-mail: j.w.j.custers@students.uu.nl

23

24 Name: Nathan Vranken

25 Affiliation 1: KU Leuven, Laboratory of Biodiversity and Evolutionary Genomics, Department of Biology,
26 Charles Deberiotstraat 32, 3000 Leuven, Belgium

27 Affiliation 2: Royal Museum for Central Africa, Biology department, Section Vertebrates, Leuvensesteenweg 13,
28 3080 Tervuren, Belgium

29 e-mail: nathan.vranken@kuleuven.be

30 ORCID: 0000-0003-0728-4715

31 Name: Maarten Van Steenberge

32 Affiliation 1: Operational Directorate Taxonomy and Phylogeny, Royal Belgian Institute for Natural Sciences,

33 Vautierstraat 29, B-1000 Brussels, Belgium

34 Affiliation 2: Research Group Zoology: Biodiversity and Toxicology, Centre for Environmental Sciences, Hasselt

35 University, Agoralaan Gebouw D, B-3590 Diepenbeek, Belgium

36 e-mail: maarten.vansteenberge@naturalsciences.be

37 ORCID: 0000-0002-6964-9014

38

39 Name: Nikol Kmentová

40 Affiliation 1: Research Group Zoology: Biodiversity and Toxicology, Centre for Environmental Sciences, Hasselt

41 University, Agoralaan Gebouw D, B-3590 Diepenbeek, Belgium

42 Affiliation 2: Department of Botany and Zoology, Faculty of Science, Masaryk University, Kotlářská 2, 611 37

43 Brno, Czech Republic

44 e-mail: nikol.kmentova@uhasselt.be

45 ORCID: 0000-0001-6554-9545

46 **Abstract**

47 The Nile perch (*Lates niloticus*) is a notorious invasive species. The introductions of Nile perch into several lakes
48 and rivers in the Lake Victoria region led to the impoverishment of the trophic food webs, particularly well
49 documented in Lake Victoria. Along with the introductions of the Nile perch, its parasites were co-introduced.
50 *Dolicirroplectanum lacustre* (Monogenea, Diplectanidae) is a gill parasite of latid fishes (*Lates* spp.) inhabiting
51 several major African freshwater systems. We examined the intra-specific diversification of *D. lacustre* from *L.*
52 *niloticus* in Lake Albert (native range) and Lake Victoria (introduced range) by assessing morphological and
53 genetic differentiation, and microhabitat preference. We expected reduced morphological and genetic diversity
54 for *D. lacustre* in Lake Victoria compared to Lake Albert, as a result of the historical introductions.
55 *Dolicirroplectanum lacustre* displays high morphological variability within and between African freshwaters.
56 Mitonuclear discordance within the morphotypes of *D. lacustre* indicates an incomplete reproductive barrier
57 between the morphotypes. The diversification in the mitochondrial gene portion is directly linked with the
58 morphotypes, while the nuclear gene portions indicate conspecificity. Based on our results, we reported reduced
59 genetic and morphological diversity, potentially being a result of a founder effect in Lake Victoria.

60

61

62 **Keywords**

63 Parasite co-introduction; Lake Albert; Lake Victoria; Host-parasite interaction; Mito-nuclear discordance; COI

64 Introduction

65 The Nile perch *Lates niloticus* (Linnaeus, 1758) is considered one of the world's most invasive species by the
66 Invasive Species Specialist Group of the International Union for Conservation of Nature (IUCN) (Lowe et al.
67 2000). The introductions of the non-native Nile perch into several of the lakes and rivers in the Lake Victoria
68 region, led to the reduction of species and functional diversity, and as a result dramatically restructured the ecology
69 of the area (Pringle 2011). These introductions were particularly well documented in Lake Victoria (Ogutu-
70 Ohwayo and Hecky 1991; Reynolds et al. 1995; Pringle 2005). In the 1950s and 1960s, the Nile perch was
71 repeatedly introduced into Lake Victoria by the release of captured individuals from Lake Albert and Lake
72 Turkana (Pringle 2005). In addition to the introductions of non-native tilapias, eutrophication, and overfishing,
73 the introductions of predatory Nile perch lead to the depletion of most, and the extinction of some endemic fish
74 species, causing impoverishment of the trophic food webs (Ogutu-Ohwayo and Hecky 1991; Seehausen et al.
75 1997; Goudswaard et al. 2006; van Zwieten et al. 2016). The introductions transformed the local artisanal fishery
76 in Lake Victoria into a global industrial fishery, with dramatic consequences for the local livelihoods (Pringle
77 2011).

78 Following Pringle (2005), the source population of Nile perch is located in Lake Albert. In Lake Albert, *L.*
79 *niloticus* occurs in sympatry with *Lates macrophthalmus* Worthington, 1929. While *L. niloticus* is larger and
80 mainly occurs inshore, *L. macrophthalmus* is a smaller, offshore species (Holden 1967). The limited genetic
81 evidence based on allozymes suggests that these are indeed distinct species, and that Nile perch in Lake Victoria
82 are largely represented by individuals of *L. niloticus* from Lake Albert, but the presence of *L. macrophthalmus* in
83 Lake Victoria cannot be excluded (Hauser et al. 2022).

84 The threats of alien species have been highlighted many times, but the co-introduction of their possibly co-invasive
85 parasites deserves more attention, as its consequences are usually underestimated (Peeler et al. 2004; Lymbery et
86 al. 2014; Kmentová et al. 2018; Šimková et al. 2018). *Co-introduced* parasites are parasites that have been
87 transported with an alien host to a new locality outside their natural range. There, they become established by
88 survival, reproduction and dispersal within the alien hosts. In case parasites were transported but have not
89 established in the introduced range, it is unlikely they are ever recorded (Lymbery et al. 2022). *Co-invasive*
90 parasites are those that have been co-introduced, and then have spread to new, native hosts (Lymbery et al. 2014),
91 so-called spill-over event(s) (Prenter et al. 2004; Goedknecht et al. 2016, 2017). The invasive potential, and success
92 of parasite establishment is affected by several ecological factors, such as the size of the founder population of
93 the host and consequently that of the parasite (Anderson and May 1991; Sakai et al. 2001; Dlugosch and Parker

94 2008), the parasite's life cycle, and environmental biotic and abiotic conditions (Taraschewski 2006; Lymbery et
95 al. 2014). Hence, co-invasion is not a straightforward consequence of co-introduction. Along with the
96 introductions of the Nile perch, some of its parasites were co-introduced, more specifically, the monogenean
97 (Platyhelminthes, Monogenea) parasite *Dolicirroplectanum lacustre* (Thurston & Paperna, 1969), and possibly,
98 the copepods *Ergasilus kandti* van Douwe, 1912 and *Dolops ranarum* (Stuhlmann, 1892) (Thurston and Paperna
99 1969), the myxosporean *H. ghaffari* Ali, 1999, and the nematodes *Contraecaecum multipapillatum* (Drasche, 1882)
100 and *Cucullanus* spp. Müller, 1777 (Outa et al. 2021).

101 Monogeneans are parasitic flatworms that can be identified morphologically by their hard parts: the attachment
102 organ (haptor) and the male copulatory organ (MCO) (Woo 2006). These ectoparasites have a direct life cycle in
103 which short-living, ciliated larvae (oncomiracidia) hatch from eggs and attach to their single host, presumably
104 synchronised with host behaviour (Kearn 1973). More specifically, members of Diplectanidae metamorphose on
105 their single host into an adult with a functional male reproductive system, and later a fully mature hermaphrodite
106 that delivers eggs after cross-fertilisation (Whittington et al. 1999). Monogeneans usually exhibit high host
107 specificity (i.e., restricted to one particular host group or host species) (Odening 1976). Exceptions have been
108 recorded, such as *Neobenedenia melleni* (MacCallum, 1927), a pathogenic species which has the broadest host
109 specificity of all monogenean species, recorded from over 100 host species (Bullard et al. 2000; Whittington and
110 Horton 2022).

111 Various speciation processes (e.g., co-speciation and intrahost speciation) are proposed to affect monogenean
112 diversity (Šimková et al. 2004; Vanhove et al. 2015). Parasites tend to diversify faster than their hosts due to their
113 short generation times and faster mutation rates (Nieberding et al. 2004; Nieberding and Olivieri 2007). This
114 should be especially apparent in comparison with long-lived and large hosts, as would be expected for *D. lacustre*
115 from its latid hosts (Kmentová et al. 2020a). The study of parasite population structure may increase the resolution
116 for understanding host population structure, phylogeography (Nieberding et al. 2004), and introduction pathways
117 (Ondračková et al. 2012). Following (Buchmann and Lindenstrøm 2002), host species selection is a consequence
118 of dynamic interactions between the parasite and its host over time. Selection acts on the scale of microhabitats
119 on the gills of a single host species, through differential microhabitat preferences (Koskivaara et al. 1992;
120 Buchmann and Uldal 1997; Raymond et al. 2006). Purportedly, the morphology of the attachment organs is an
121 important adaptation to the monogenean's host, and to specific microhabitats within the host, since these structures
122 could influence the ability of the parasite to infect, feed, and reproduce on a specific host species (Šimková et al.
123 2002).

124 *Dolicirroplectanum lacustre* (Monogenea, Diplectanidae) is the single known monogenean gill parasite of four
125 species of lates perches (Perciformes, Latidae) in African freshwaters (*L. niloticus*, *Lates microlepis* Boulenger,
126 1898, *Lates angustifrons* Boulenger, 1906, and *Lates mariae* Steindachner, 1909) that have been examined for
127 parasites (Kmentová et al. 2020a). *Dolicirroplectanum lacustre* displays a high morphological variation: a “wide
128 range of shapes and sizes” (Thurston and Paperna 1969). In the original description of *D. lacustre*, Thurston &
129 Paperna (1969) noted these differences and identified a so-called ‘slender form’, with a well delineated haptor at
130 the posterior end of the body, and a usually longer ‘gravid form’ which is proportionally wider and has an
131 embedded haptor. In accordance with Thurston & Paperna (1969), Kmentová et al. (2020) reported distinct
132 morphotypes of *D. lacustre* in Lake Albert and continuous variation across several African freshwater systems.
133 Although *D. lacustre* showed high morphological variation, the genetic differentiation between parasites of
134 allopatric host species was not as high as typically associated with distinct diplectanid species (Kmentová et al.
135 2020a).

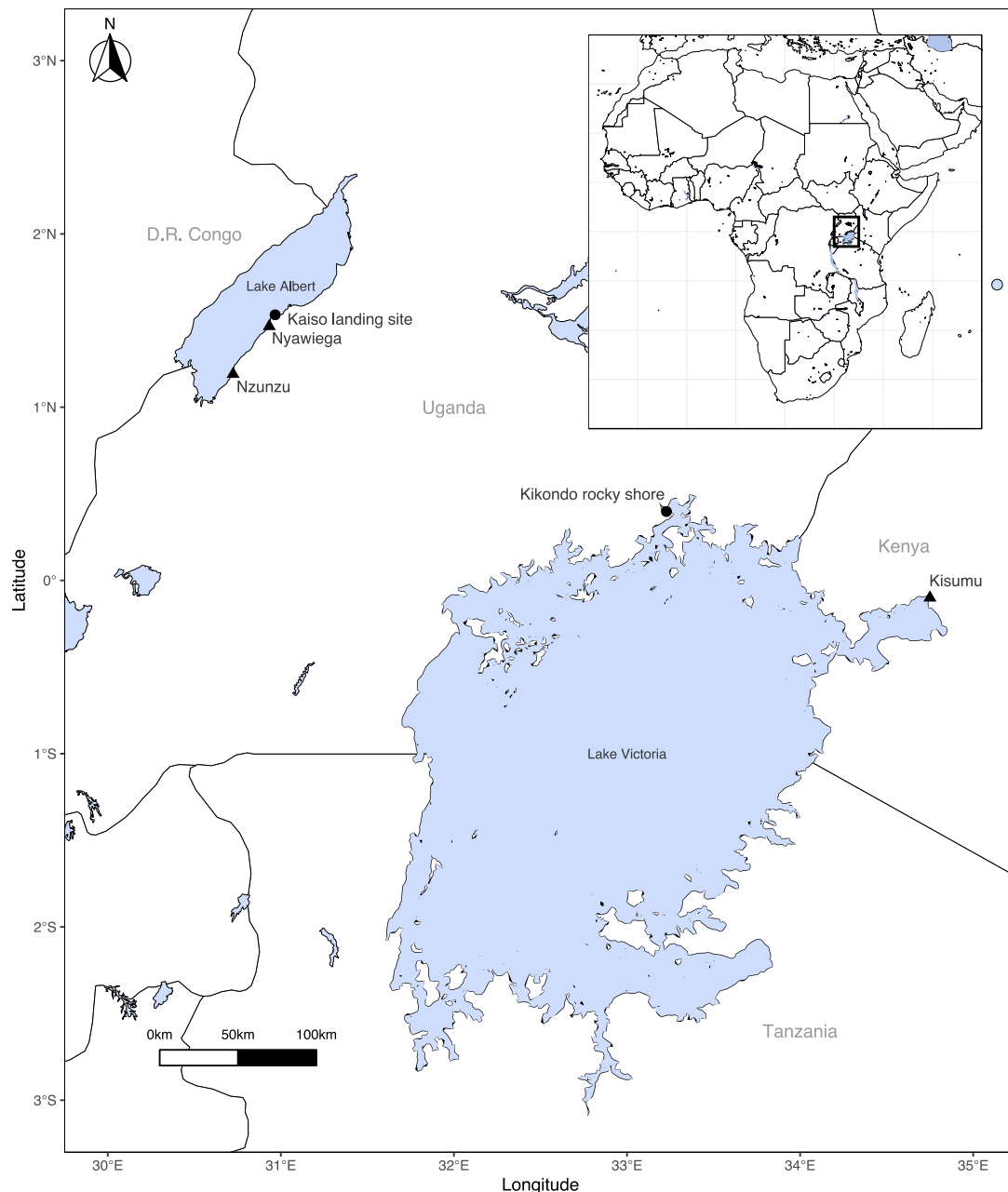
136 In this study, we investigate the fine-scale morphological and genetic differentiation of *D. lacustre* from Nile
137 perch in its native (Lake Albert) and introduced range (Lake Victoria). We hypothesise that: (1) a founder effect
138 had taken place in Lake Victoria. Hence, we expect reduced genetic and morphological diversity within *D.*
139 *lacustre* in its introduced range in comparison with its native population in Lake Albert; (2) we expect high
140 phenotypic variation and low genetic differentiation in Lake Albert, as supported by the presence of distinct
141 morphotypes in earlier studies (Thurston and Paperna 1969; Kmentová et al. 2020a); (3) there has been a niche
142 shift among the populations in Lake Albert that resulted in differential gill microhabitat preferences which led to
143 morphological changes of *D. lacustre*, producing an (imperfect) reproductive barrier between the morphotypes.

144

145 **Materials and methods**

146 **Sampling of the Nile perch and its gill parasites**

147 Fish samples of *L. niloticus* from Lake Victoria and Lake Albert were examined (Table 1). Fresh specimens were
148 obtained from local fishermen in Kaiso, Uganda (Lake Albert) and Kikondo, Uganda (Lake Victoria) during a
149 field expedition in 2019 (Figure 1). In total, the gills of 15 fish specimens (8 samples from Lake Albert, 7 samples
150 from Lake Victoria) were dissected and preserved in ethanol (99% EtOH). The pairs of gills from 12 randomly
151 chosen specimens (5 specimens from Lake Albert, 7 specimens from Lake Victoria) were examined following the
152 standard protocol of Ergens and Lom (1970) using a Leica EZ4 stereomicroscope. Monogenean gill parasites were
153 extracted, some of the individuals were cut in three parts with the posterior and anterior parts mounted on slides



154

155 **Fig. 1** Sampling localities of *Lates niloticus* at Lake Albert and Lake Victoria: Kaiso landing site, Lake Albert,
156 Uganda (N 00°23'52.0", E 33°13'39.6") (n = 8) and Kikondo rocky shore, Lake Victoria, Uganda (N 01°31'59.3",
157 E 30°58'00.9") (n = 7). Sampled localities from Kmentová et al. (2020a) depicted by triangles: Kisumu, Lake
158 Victoria, Kenya (00°06'S, 34°45'E; n = 3) and Nzunzu, Lake Albert, Uganda (N 01°11'27.6", E 30°43'26.4"; n =
159 12) and Nyawiega, Lake Albert (01°28'N, 30°56'E; n = 3). We used the geographical databases 'Global Lakes
160 and Wetlands Database' (WWF, Lehner and Dooll) and the 'High Resolution World Vector Map' from package
161 rnatuarearth (South 2021) to map the lakes, and countries respectively. The map was created using the packages
162 sf (Pebesma 2018), ggspatial (Dunnington 2021) and ggplot 2 (Wickham 2016) in R Studio v 1.4.1106 (RStudio
163 Team 2020).

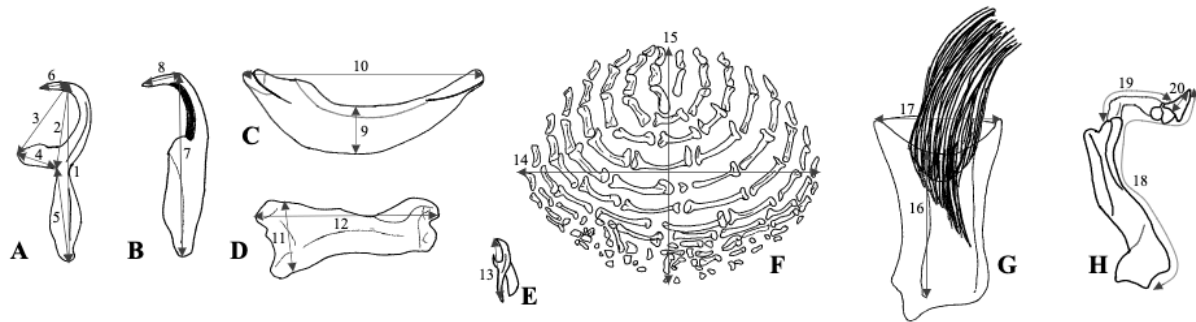
164 (using Hoyer's medium as fixative) for morphological characterisation, and the central part preserved in ethanol
165 (99% EtOH) for genetic analyses. In addition, several individuals were mounted on slides as entire specimens and
166 fixed (Hoyer's medium). Whole mounted parasite specimens were drawn by group according to locality and
167 morphotype (Thurston & Paperna 1969): Lake Albert (gravid), Lake Albert (slender), Lake Victoria using a Leica
168 DM2500 optical microscope at a magnification of 1000x (objective x100 1.32 oil XHC PL FLUOTAR, ocular
169 x10) and a reMarkable® graphical tablet. Drawings were edited in Adobe Photoshop® 2021 (v 22.3.1).
170 Infection parameters prevalence (P) and mean infection intensity (MI) were calculated following Bush et al.
171 (1997). Prevalence was determined as the proportion of host specimens infected with *D. lacustre*. The mean
172 infection intensity was calculated as the ratio of the total number of parasite specimens of *D. lacustre* to the total
173 number of host specimens infected by *D. lacustre*. During extraction of the parasites, the microhabitat of each
174 parasite specimen on the gill was recorded by subdivision of each gill arch into nine microhabitats. Following
175 Gobbin et al. (2021) each gill arch was subdivided along the longitudinal axis (dorsal, median, ventral) and
176 transversal axis (proximal, central, distal; from gill bar to tips of gill filaments).

177

178 *Morphometrics of the parasites*

179 Measurements of the hard parts, total body size and distances between the two pairs of eyespots were obtained
180 from 71 specimens at a magnification of 1000x (objective x100/1.32 oil XHC PL FLUOTAR, ocular x10) with
181 differential interference contrast on a Leica DM2500 optical microscope using Las X software v3.6.0.20104. A
182 total of 20 parameters of the hard parts of the haptor and MCO were measured (Figure 2). The terminology of
183 (Justine and Henry 2010) was followed. To investigate morphological differentiation in haptor morphology
184 between localities and morphotypes, raw haptor measurements were analysed by multivariate statistical
185 techniques in RStudio® (R Core Team 2020). A Principal Component Analysis (PCA) was performed with 23
186 measurements from 56 individuals (combined with measurements of 22 individuals from Kmentová et al. (2020))
187 using the package FactoMineR (Lê et al. 2008). Individuals were selected for the PCA when more than 50% of
188 the measured characteristics were obtained. Missing values for measurements were imputed by the variable mean
189 as default in the package factoextra (Kassambara and Mundt 2017). Results were visualised using the packages
190 ggplot2 (Wickham, 2016), reshape2 (Wickham, 2007), factoextra (Kassambara and Mundt 2017), tidyr (Wickham
191 et al. 2021), and MASS (Venables & Ripley 2002).

192



193

194 **Fig. 2** Measurements for sclerotized structures of haptor and reproductive organs of *Dolicirroplectanum lacustre*:

195 A, Ventral anchor: 1, total length; 2, length to notch; 3, length to inner root, 4, inner root length, 5, outer root
196 length, 6, point length; B - Dorsal anchor: 7, total length, 8, point length; C - Ventral bar: 9, maximum width, 10,
197 straight length; D - Dorsal bar: 11, maximum width, 12, straight length; E - Hook: 13, hook length; F -
198 Squamodisc: 14, squamodisc width; 15, squamodisc length; G - Male copulatory organ: 16, copulatory tube
199 length; 17, copulatory tube width; H - Vagina: 18, total length; 19, tube length; 20, point length. Structures drawn
200 of *D. lacustre* from host specimen HP4318 from Lake Victoria.

201

202

203 To investigate the relationship between groups per lake and morphotype, and level of differentiation in the
204 measurements of the MCO, parametric tests were applied. Analysis of Variance (ANOVA) was applied for overall
205 effects; for pairwise comparisons, unpaired t-tests were applied, implemented in the package stats (R Core Team
206 2020). Assumptions of normality and homogeneity of variance were tested by Shapiro-Wilk's W tests from the
207 package stats (R Core Team 2020) and Levene's tests from the car package (Fox and Weisberg 2019),
208 respectively, as well as by graphical examination of the residuals.

209

210 *Molecular characterisation of the parasites*

211 Whole genomic DNA was extracted from the central part of 64 monogenean parasite individuals (20 from Lake
212 Victoria, 44 from Lake Albert) that were preserved in 99% ethanol using the following protocol. The sample was
213 spinned down and ethanol was removed. 195 μ L of TNES buffer (400 mM NaCl, 20 mM EDTA, 50 mM Tris (pH
214 8, 0.5% SDS)), and 5 μ L of proteinase K (20 mg/mL) were added to the sample. After 45 min of incubation at 55
215 $^{\circ}$ C, 2 μ L of InvitrogenTM yeast RNA (10 mg/mL) was added as a carrier. Next, 65 μ L of NaCl (5 M) and 290 μ L
216 of 96% ethanol were added, and the sample was cooled for 60 min at -20 $^{\circ}$ C. The sample was spinned down for
217 15 min at 18,000 rcf to a small pellet. Supernatant was removed and substituted by 1 mL of chilled 70% ethanol.

218 Next, the samples were centrifuged for 5 min at 18,000 rcf. The ethanol rinse step (removal of supernatant,
219 addition of ethanol and centrifugation) was repeated once again. The ethanol was removed, and the DNA was
220 eluted in 100 μ L of 0.1X TE (0.02 % tween-20). The DNA extract was placed at 4 $^{\circ}$ C for resuspension overnight,
221 and stored at a temperature of -20 $^{\circ}$ C.

222 Sequences of a portion of the mitochondrial cytochrome *c* oxidase subunit 1 (COI) gene, and nuclear gene portions
223 from the small and large ribosomal subunit gene (18S rDNA and 28S rDNA) and internal transcriber spacer 1
224 (ITS-1) were obtained following PCR. The three nuclear markers evolve at different rates and are suitable to
225 assess genetic divergence at the interspecific level (Vanhove et al. 2013). COI is the current standard marker for
226 intraspecific genetic differentiation in many monogean taxa (Kmentová et al. 2020b) due to its relatively high
227 rate of molecular evolution in comparison to rDNA. Part of the mitochondrial COI gene was amplified using
228 ASmit1 (5'-TTTTTTGGGCATCCTGAGGTTTAT-3') (Littlewood et al. 1997) combined with Schisto3 (5'-
229 TAATGCATMGGAAAAACA-3') (Lockyer et al. 2003), and with Asmit2 (5'-
230 TAAAGAAAGAACATAATGAAAATG-3') (Littlewood et al. 1997) in a nested PCR following Vanhove et al.
231 (2015). For both primer combinations, the amplification reaction contained 24 μ L of PCR mix (0.2 μ L of 1 unit
232 Invitrogen Taq polymerase, 2.5 μ L Invitrogen PCR buffer, 2 μ L of 2 mM MgCl₂, 0.5 μ L of 0.2 mM dNTPs, 2 μ L
233 of 0.8 μ M of each primer, 14.80 μ L of dd H₂O) with 1 μ L of isolated DNA (concentration was not measured) in
234 a total reaction volume of 25 μ L and was performed under the following conditions for the first reaction: initial
235 denaturation at 95 $^{\circ}$ C for 5 min, 40 cycles of 1 min at 94 $^{\circ}$ C, 1 min at 44 $^{\circ}$ C and 1 min at 72 $^{\circ}$ C, and a final 7 min
236 at 72 $^{\circ}$ C. The second (nested) reaction was performed under the following conditions: 5 min at 95 $^{\circ}$ C, 40 cycles
237 of 1 min at 94 $^{\circ}$ C, 1 min at 50 $^{\circ}$ C and 1 min at 72 $^{\circ}$ C, and a final 7 min at 72 $^{\circ}$ C. Primers C1 (5'-
238 ACCCGCTGAATTTAAGCAT-3') and D2 (5'-TGGTCCGTGTTTCAAGAC-3') (Hassouna et al. 1984) were
239 used for amplification of the partial 28S rDNA gene. Each PCR reaction contained 1 unit of Q5[®] Hot Start High-
240 Fidelity DNA Polymerase, 1X PCR buffer containing 0.1 mg/mL bovine serum albumin (BSA), 0.2 mM dNTPs,
241 0.5 μ M of each primer and 2 μ L of isolated DNA (concentration not measured) in a total reaction volume of 25
242 μ L. The reaction proceeded under the following conditions: 2 min at 95 $^{\circ}$ C, 39 cycles of 20 s at 94 $^{\circ}$ C, 30 s at 65
243 $^{\circ}$ C, 1 min 30 s at 72 $^{\circ}$ C, and 10 min. at 72 $^{\circ}$ C. Partial 18S rDNA and ITS-1 were amplified using the S1 (5'-
244 ATTCCGATAACGAACGAGACT-3') (Sinnappah et al. 2001) and IR8 (5'-GCTAGCTGCGTTCTTCATCGA-
245 3') (Šimková et al. 2003) primers. Each reaction contained 1 unit of Q5[®] Hot Start High-Fidelity DNA
246 Polymerase, 1X PCR buffer containing 0.1 mg/mL BSA, 0.2 mM dNTPs, 0.5 μ M of each primer and 2 μ L of
247 isolated DNA (concentration not measured) in a total reaction volume of 25 μ L under the following conditions: 2

248 min at 95 °C, 39 cycles of 1 min at 94 °C, 1 min at 64 °C and 1 min 30 s at 72 °C, and a final elongation of 10
249 min at 72 °C. PCR amplification success was verified by agarose gel electrophoresis. A volume of 10 µl of the
250 PCR product was enzymatically purified for positive samples, using 4 µL of ExoSAP-IT reagent under the
251 following conditions: 15 min at a temperature of 37 °C, and 15 min at 80 °C. Purified PCR products were sent out
252 to Macrogen Europe B.V. for Sanger sequencing; amplification primers were used for sequencing reactions. The
253 acquired sequences for each marker were visually inspected and trimmed using the software Geneious v2021.1.1.
254 Sequences were aligned with previously published sequences of *D. lacustre* from Kmentová et al. (2020a) using
255 MUSCLE (Edgar 2004) under default settings as implemented in Geneious v2021.1.1.

256

257 *Genetic differentiation and phylogeography*

258 Per marker, uncorrected pairwise genetic distances (p-distances) among sequences were computed in Geneious
259 Prime v2021.1.1. Already available sequences of the four genetic markers of *D. lacustre* from Lake Albert
260 (GenBank accession numbers: MK908145.1-MK908196.1, MK937576.1, MK937579.1-MK937581.1,
261 MK937574.1, MK937575.1) were included in the analyses. Haplotype networks were constructed using Median
262 Joining networks following Bandelt et al. (1999) in PopART v1.7104 (Leigh and Bryant 2015) for each marker
263 separately. The analyses of population structure and demographic history within *D. lacustre* were based on a 325
264 bp fragment of COI. This allows for the detection of recent evolutionary events, such as possible incipient
265 speciation as a result of host preference (Kmentová et al. 2016). Genetic diversity of COI was assessed by the
266 number of haplotypes and polymorphic sites, haplotype diversity (h) and nucleotide diversity (π), calculated in
267 Arlequin v3.5.2.2 (Excoffier and Lischer 2010). Differentiation among pre-defined populations (according to
268 locality and morphotype) was estimated by F_{ST} . Analysis of Molecular Variance (AMOVA) based on F-statistics
269 was applied to test for a significant population structure of *D. lacustre* at the level of locality (non-native and
270 native). To test for signals of past population expansion, two different neutrality tests, Tajima's D and Fu's F_s
271 were calculated in Arlequin v3.5.2.2 (Excoffier and Lischer 2010) for the two hypothetical populations. First, the
272 neutrality tests were calculated for the specimens from Lake Albert. Second, the Lake Victoria individuals were
273 added.

274

275 *Microhabitat description*

276 Parasite microhabitats were visualised by the combination of the spatial distribution of *D. lacustre* on each pair
277 of gills (left and right gills per host individual) into a single spatial distribution. Per lake/morphotype, the spatial

278 distribution of all individuals was summarised in a heat map using the `geom_tile` function as part of the `ggplot2`
279 package (Wickham, 2016). As replicates per population were limited, no statistical effects could be verified.

280

281 **Results**

282 A total of 124 monogenean gill parasites were retrieved from the 8 host specimens examined in this study (Table
283 1). The parasite specimens were morphologically identified as *Dolicirroplectanum lacustre*. Diplectanid-specific
284 haptor characteristics included the presence of ventral and dorsal squamodiscs, and three transversal bars
285 connected to pairs of ventral hooks and dorsal hooks. *Dolicirroplectanum lacustre* can be distinguished from other
286 diplectanids by the observed characteristics: a longer outer root in the ventral anchors (Thurston and Paperna
287 1969); a wide, robust and barrel-shaped MCO formed by two anteriorly oriented nested tubes, one encasing the
288 other (Kmentová et al. 2020a); the presence of an accessory piece; squamodiscs composed of variable concentric
289 rows of bone shaped rodlets forming open rings; reduced inner roots of the ventral anchor. Additional
290 characteristics include: a simple prostatic reservoir; seminal vesicle as an expansion of vas deferens; intercaecal,
291 pre-testicular ovary encircles right caecum (Thurston and Paperna 1969; Kmentová et al. 2020a). The dorsal bars
292 are large, bone-shaped structures with a broad base pointing towards the centre of the haptor. The ventral bar has
293 an elongated, oval shape with tapering ends. A sclerotized vagina was observed in 29 out of the 71 measured
294 specimens. Additional observed characteristics include two pairs of eyespots, of which the posterior eyespots are
295 larger and closer together, a thin tegument covering the body, and 14 rudimentary hooklets of equal sizes.
296 Infection parameters varied per location, with greater prevalence (P) and mean infection intensity (MI) in Lake
297 Albert compared to Lake Victoria (Table 1). We observed co-infections (infection by both morphotypes of *D.*
298 *lacustre* on a single host individual of *L. niloticus*) of the morphotypes on two host individuals in Lake Albert
299 (HP4094, HPA1b.X).

300

301 **Table 1** Overview of host species examined for monogenean parasites and infection parameters per locality.

Locality (Geographic coordinates, year)	Number of fish specimens examined	Number of infected fish specimens	Number of monogenean individuals	P/MI
Kaiso landing site, Lake Albert, Uganda (N 0°23'52.0", E 33°13'39.6")	5	5	90	1/18
Kikondo rocky shore, Lake Victoria, Uganda (N 1°31'59.3", E 30°58'00.9")	7	3	34	0.43/11.3

302 P, prevalence; MI, Mean infection intensity

303

304 *Morphotypes and morphometric variation*

305 During the screening procedure, clear size differences in total body length and haptoral sclerotised structures were
306 apparent between individuals of *D. lacustre* from Lake Albert. Therefore, we classified the specimens from Lake
307 Albert under the gravid morphotype (n=9) and slender morphotype (n=42) following Thurston and Paperna
308 (1969). Specimens of the gravid morphotype (Fig. 3a) were identified by the following haptoral features:
309 proportionally larger body, longer ventral anchors with a longer outer root, longer dorsal anchors with a shorter
310 tip, long and narrow ventral and dorsal bars, larger ventral and dorsal squamodiscs (Table 2). The MCO for gravid
311 individuals was longer, as was the vagina (Table 2). Also, the distance between the eyespots of both pairs was
312 larger in the gravid morphotype. Conversely, specimens of the slender morphotype (Fig. 3b) were identified by a
313 proportionally smaller body, shorter ventral anchors with a shorter outer root, shorter dorsal anchors with a longer
314 tip, wider but shorter ventral and dorsal bars, and smaller squamodiscs. The copulatory tube for the slender
315 morphotype was usually shorter, as was the vagina total length and tube length (Table 2; Table S1). The
316 individuals retrieved from Lake Victoria had sclerotized structures of the slender morphotype (Fig. 3c). However,
317 these individuals were smaller in total body size, they had shorter ventral and dorsal anchors, narrower ventral
318 and dorsal bars, and squamodiscs were less wide in comparison with both morphotypes in Lake Albert. The
319 specimens from Lake Victoria had a shorter copulatory tube and vagina in comparison with the specimens in Lake
320 Albert (Table 2).

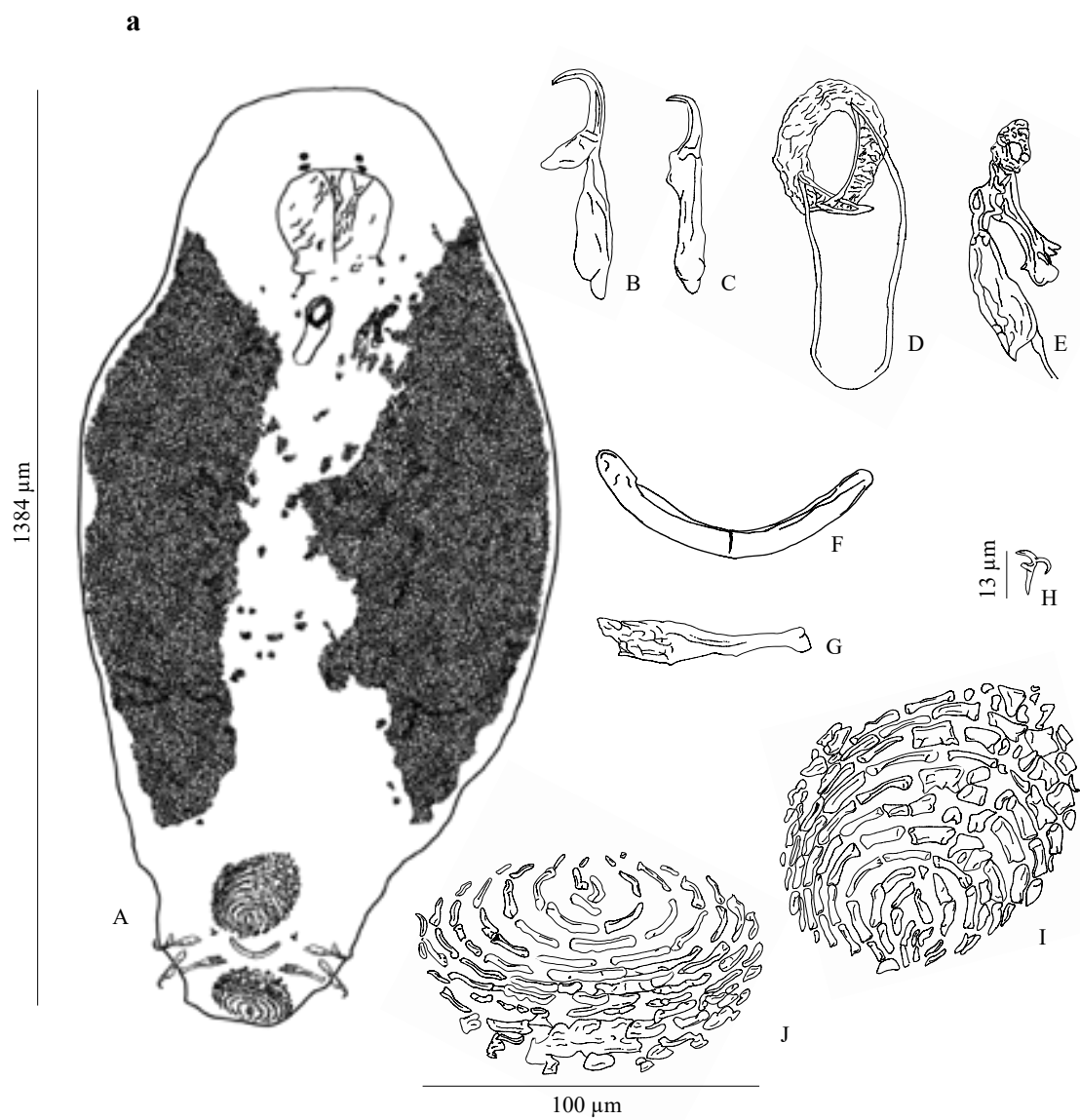
321 Variation in morphology of the measured characteristics was summarised in a biplot through PCA of 17
322 parameters (Table 2; Fig. 4). The PCA biplot (Fig. 4a) depicts the variation in morphometric measurements where
323 PC1 explains 46.67%, and PC2 explains 14.79% of the variation with main contributors: distance between the
324 eyespots of the larger pair, total body width, total length of the dorsal anchor, maximum width of the ventral bar,
325 and straight length of the ventral bar. The density plot of PC1 scores (Fig. 4b), in accordance with the biplot,
326 indicates a clearly distinctive morphometry in haptoral structures between two morphotypes present in Lake
327 Albert. The morphometry of the slender morphotype coincides with the morphometry of the specimens from Lake
328 Victoria (Fig. 4b). We observed similar morphometric variation for the slender morphotype from Lake Albert and
329 the specimens from Lake Victoria. The variation was larger for the gravid morphotype, although this could be an
330 artefact of sample size in Fig. 4 (since the 95 % C.I. is larger when less observations are included for the ellipse).
331 The separation of the gravid morphotype along the PC1 axis from the two other groups can be explained by larger
332 distance between the eyespots of the larger pair, a wider body, and a narrower ventral bar (Fig. 4a, Table 2).

333
334
335

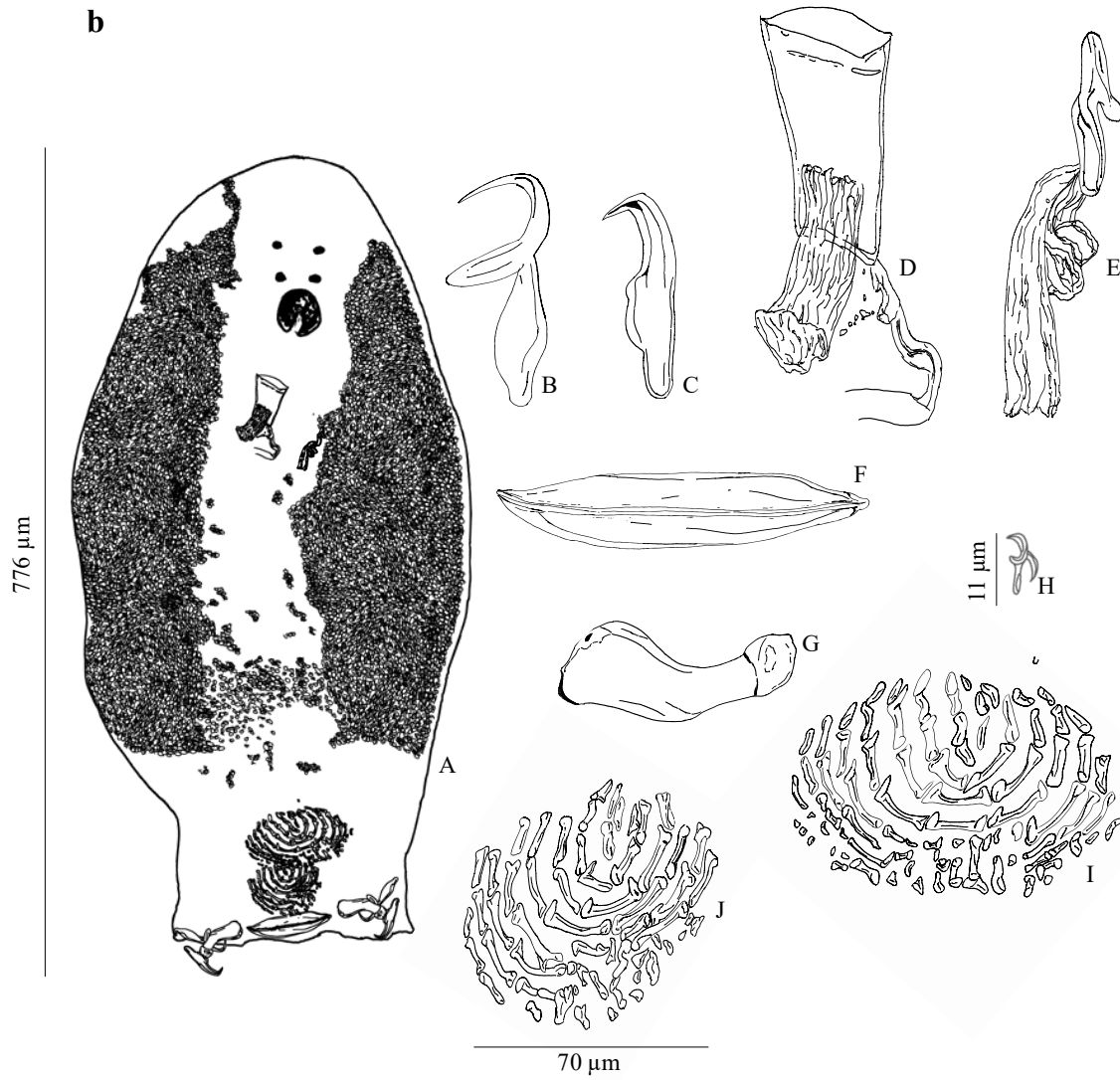
Table 2. Overview of morphometric measurements performed on haptoral and genital sclerotised structures of *Dolicirroplectanum lacustre*. Values depict the average measurements \pm the standard deviations; min. - max. values, and *n*, the number of individuals.

Parameters (μm)	<i>L. niloticus</i> , Lake Victoria	<i>L. niloticus</i> , Lake Albert (slender morphotype)	<i>L. niloticus</i> , Lake Albert (gravid morphotype)
Total length	533.4 \pm 47.0 (475.9–603.7); n=6	776.6 \pm 97.1 (601.3–899.3); n=8	1384.0; n=1
Total width	298.6 \pm 61.1 (222.2–417.4); n=12	361.6 \pm 91.7 (214.6–555.4); n=24	947.8 \pm 129.5 (843.7–1245.0); n=8
<i>Ventral anchor</i>			
Total length	47.8 \pm 2.4 (44.2–51.6); n=26	51.1 \pm 3.8 (45.4–59.3); n=25	63.2 \pm 4.0 (58.9–69.1); n=8
Length to notch	22.2 \pm 0.74 (20.6–23.5); n=26	21.1 \pm 1.8 (17.9–24.5); n=25	24.5 \pm 0.8 (23.0–25.7); n=8
Length to inner root	25.4 \pm 1.2 (22.4–27.8); n=26	25.5 \pm 1.3 (23.1–28.8); n=25	28.8 \pm 1.3 (26.5–30.4); n=8
Inner root length	12.4 \pm 0.9 (11.2–14.8); n=26	12.8 \pm 0.7 (11.1–14.0); n=25	14.9 \pm 0.9 (13.2–16.0); n=8
Outer root length	28.5 \pm 2.2 (22.5–32.9); n=26	31.7 \pm 3.3 (26.03–37.6); n=25	40.9 \pm 4.1 (36.5–46.5); n=8
Point length	9.8 \pm 0.9 (8.1–11.9); n=26	9.5 \pm 0.7 (7.9–11.0); n=25	10.6 \pm 1.3 (8.5–12.0); n=8
<i>Dorsal anchor</i>			
Total length	42.3 \pm 2.1 (37.2–44.9); n=26	45.0 \pm 2.6 (39.5–52.6); n=25	56.5 \pm 3.1 (50.4–59.7); n=8
Point length	8.6 \pm 1.3 (4.0–10.5); n=26	9.3 \pm 0.9 (7.8–10.5); n=25	6.5 \pm 0.8 (5.2–8.0); n=8
<i>Ventral bar</i>			
Straight length	65.7 \pm 5.2 (55.7–74.1); n=24	68.9 \pm 11.7 (37.4–94.6); n=25	80.4 \pm 10.0 (66.9–95.2); n=8
Maximum width	16.2 \pm 1.5 (11.7–18.1); n=24	16.8 \pm 2.3 (11.4–19.5); n=25	6.1 \pm 0.6 (5.5–6.8); n=8
<i>Dorsal bar</i>			
Straight length	44.8 \pm 3.9 (33.1–52.5); n=26	51.6 \pm 6.1 (34.1–61.5); n=25	56.4 \pm 6.6 (45.3–61.7); n=8
Maximum width	16.9 \pm 2.1 (13.2–23.0); n=26	17.6 \pm 2.7 (11.0–22.3); n=25	11.4 \pm 2.6 (6.7–14.2); n=8
<i>Ventral squamodisc</i>			
Width	69.3 \pm 13.5 (51.0–102.2); n=23	80.4 \pm 18.9 (50.1–125.4); n=19	115.9 \pm 29.8 (83.9–156.1); n=8
Length	63.5 \pm 11.9 (43.6–78.9); n=23	57.1 \pm 14.1 (28.4–86.9); n=19	108.7 \pm 39.5 (69.5–191.0); n=8
<i>Dorsal squamodisc</i>			
Width	59.2 \pm 11.8 (39.5–89.1); n=23	69.1 \pm 17.7 (34.2–102.5); n=22	99.3 \pm 27.7 (55.3–135.4); n=8
Length	55.5 \pm 12.8 (35.9–76.9); n=23	57.0 \pm 14.6 (26.7–84.2); n=18	71.6 \pm 14.1 (58.7–103.3); n=8
<i>Hook</i>			
Length	11.9 \pm 1.2 (10.5–16.6); n=24	11.0 \pm 1.1 (8.4–13.5); n=24	13.3 \pm 0.6 (12.1–14.0); n=8
<i>Male copulatory organ (MCO)</i>			
Straight length	52.2 \pm 13.3 (32.0–70.7); n=14	64.0 \pm 10.2 (46.9–88.2); n=27	91.5 \pm 7.8 (77.2–100.6); n=7
Straight width	27.4 \pm 7.8 (12.4–40.8); n=16	36.1 \pm 5.6 (19.2–44.9); n=27	40.4 \pm 3.8 (34.9–45.1); n=7
<i>Vagina</i>			
Total length	36.3 \pm 3.0 (32.9–41.6); n=6	43.2 \pm 8.2 (32.3–57.8); n=18	75.1 \pm 18.5 (54.3–97.8); n=4
Tube length	5.8 \pm 1.1 (3.7–7.1); n=7	8.5 \pm 1.1 (7.2–11.4); n=17	17.3 \pm 8.3 (7.9–26.8); n=4
Point length	7.6 \pm 2.0 (3.9–10.4); n=8	8.3 \pm 1.6 (6.4–11.6); n=19	11.6 \pm 2.0 (10.1–14.5); n=4
<i>Eyespots</i>			
Smaller pair distance	38.2 \pm 9.3 (16.9–49.5); n=28	44.8 \pm 10.6 (18.6–66.0); n=26	95.4 \pm 15.6 (67.1–111.6); n=7
Larger pair distance	20.2 \pm 6.7 (6.1–33.1); n=31	27.4 \pm 10.9 (13.3–65.9); n=28	93.3 \pm 15.9 (64.8–113.7); n=8

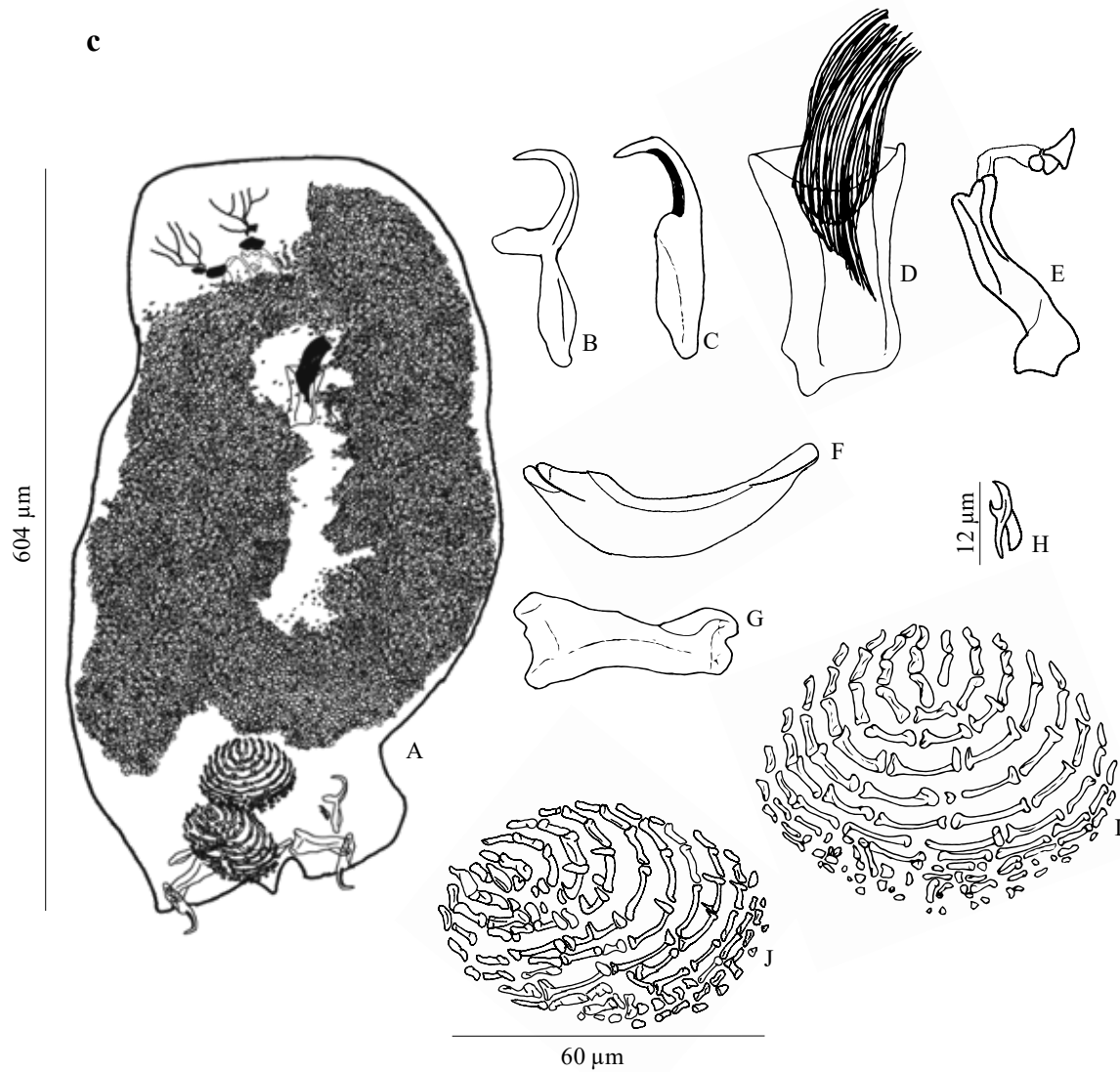
336



337



338



339

340 **Fig. 3** Morphotypes of *Dolocirroplectanum lacustre* collected from *Lates niloticus*. Specimens drawn from a

341 dorsal view with Hoyer's medium as fixative. a *Dolocirroplectanum lacustre* gravid morphotype from Lake Albert

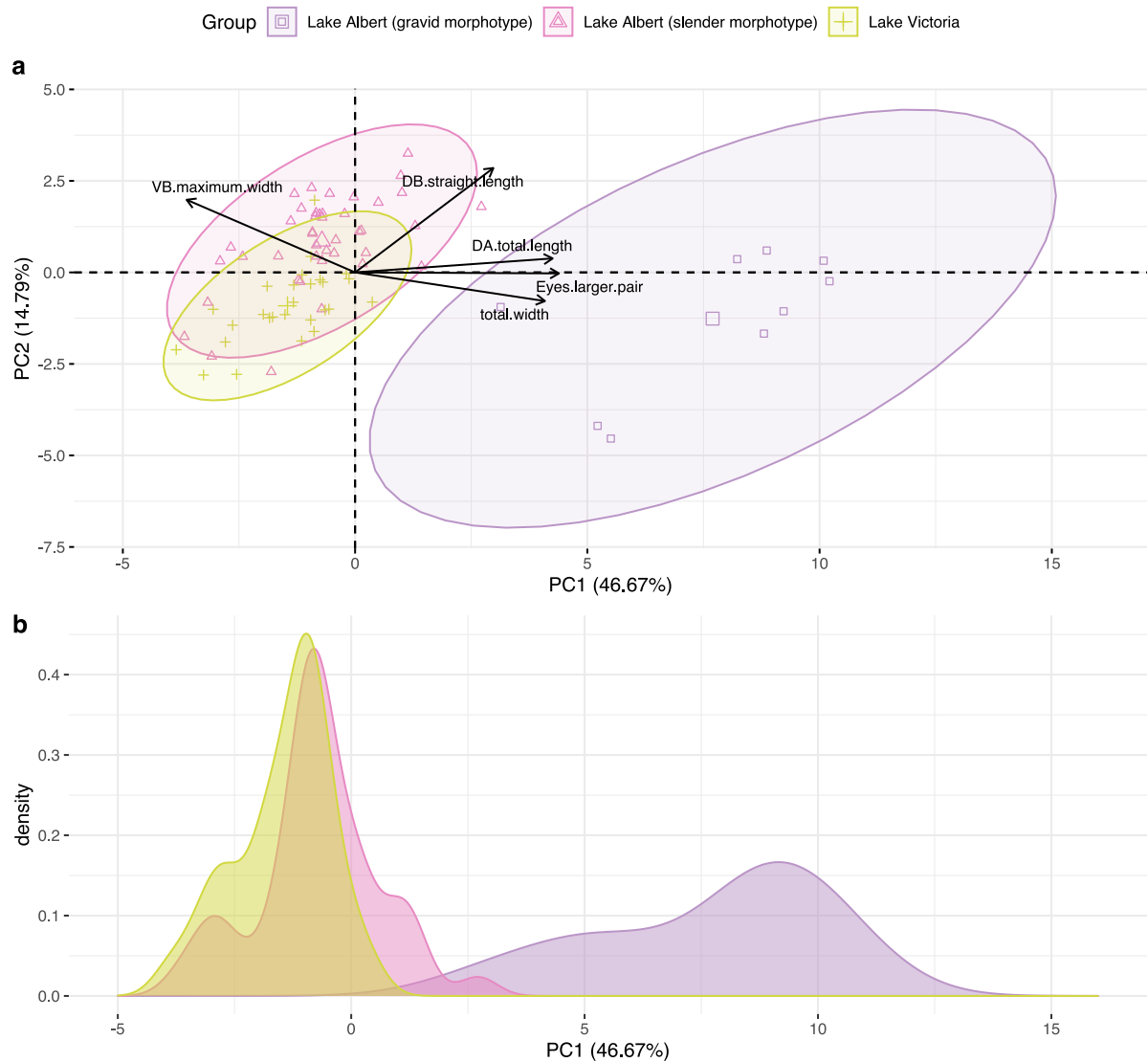
342 (host specimen HP4094); (b) *Dolocirroplectanum lacustre* slender morphotype from Lake Albert (host specimen

343 HP4099); (c) *Dolocirroplectanum lacustre* from Lake Victoria (host specimen HP4318). Drawings from the dorsal

344 view with A, whole mount; B, ventral anchor; C, dorsal anchor; D, male copulatory organ; E, vagina; F, ventral

345 bar; G, dorsal bar; H, hook; I, ventral squamodisc; J, dorsal squamodisc.

346

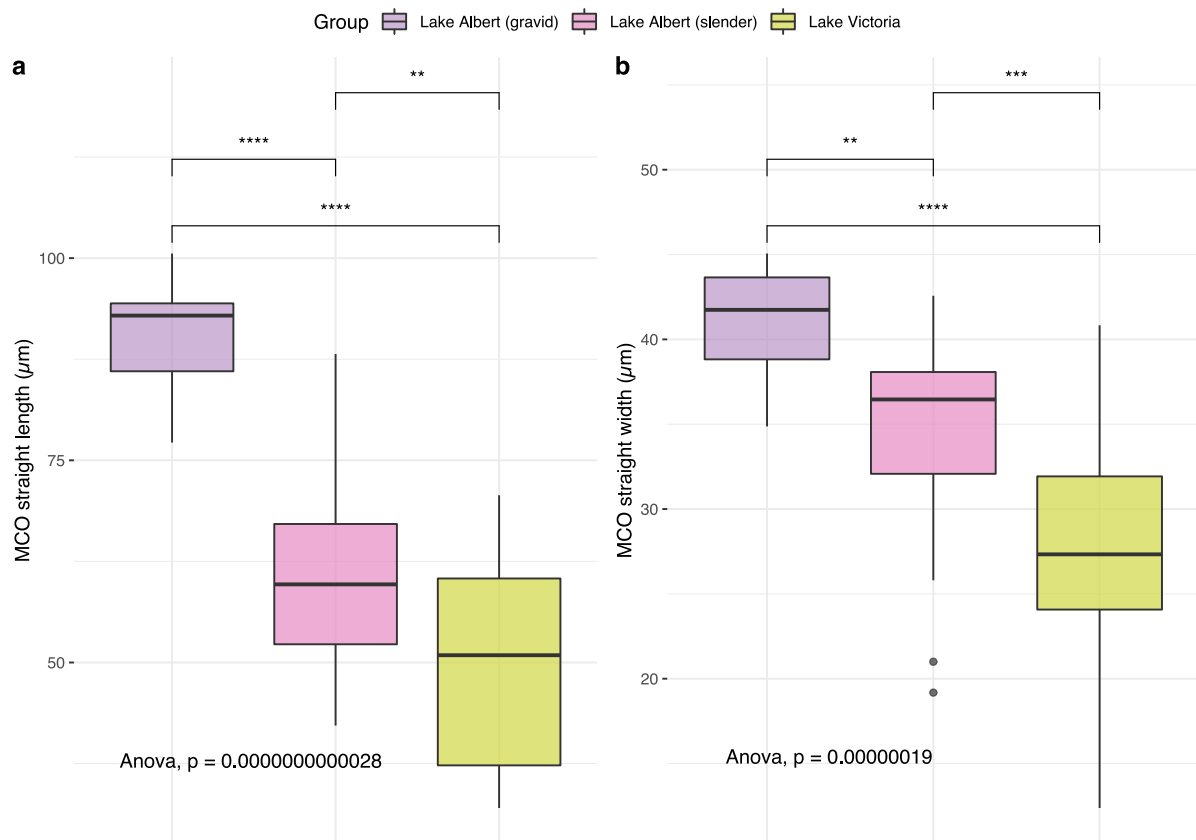


347

348 **Fig. 4** Morphometric variation of haptor structures and body size in *Dolicirroplectanum lacustre*. (a) Biplot
349 based on PCA from measured characteristics with gravid specimens (n= 9) and slender specimens (n = 42) from
350 Lake Albert, and specimens from Lake Victoria (n= 27) with 95% C.I. ellipses; (b) Density plots depicting the
351 morphometric variation of haptor structures across populations, summarised in PC1 that explains 46.67% of the
352 haptor variation. Colours and symbols depict different populations. Vectors indicate the 5 most influential
353 haptor and body size measurements (DB, dorsal bar; VB, ventral bar; DA, dorsal anchor).

354

355 MCO straight length (ANOVA, df = 2, $F = 40.36$, $p = 0.000000000028$) and straight width (ANOVA, df = 2, F
356 = 19.638, $p = 0.00000019$) were found to differ significantly among the three groups (Fig 5). Overall, MCO length
357 and width were largest for the gravid morphotype (n = 8) from Lake Albert, and smallest in the specimens from
358 Lake Victoria (n = 19). All pairwise comparisons revealed significant differences between the three studied groups
359 (unpaired t-test with Bonferroni correction) (Fig. 5).



360

361 **Figure 5.** Morphometric variation of the male copulatory organ (MCO) of *Dolocirroplectanum lacustre*. (a)
362 variation in MCO straight length (μm) between morphotypes; (b) variation in MCO straight width (μm) between
363 morphotypes. Colours depict the morphotype from which the MCO measurements were obtained. Following One-
364 way Anova, the p-values are depicted per MCO measurement, with pairwise comparisons following t-tests with
365 Bonferroni correction for multiple testing (ns, $p > 0.05$; * $p \leq 0.05$; ** $p \leq 0.01$; *** $p \leq 0.001$; **** $p \leq 0.0001$).
366 Box plots with statistical entries: minimum, first quartile, median, third quartile, maximum, outliers. Data from
367 this study combined with measurements of Kmentová et al. (2020a).

368

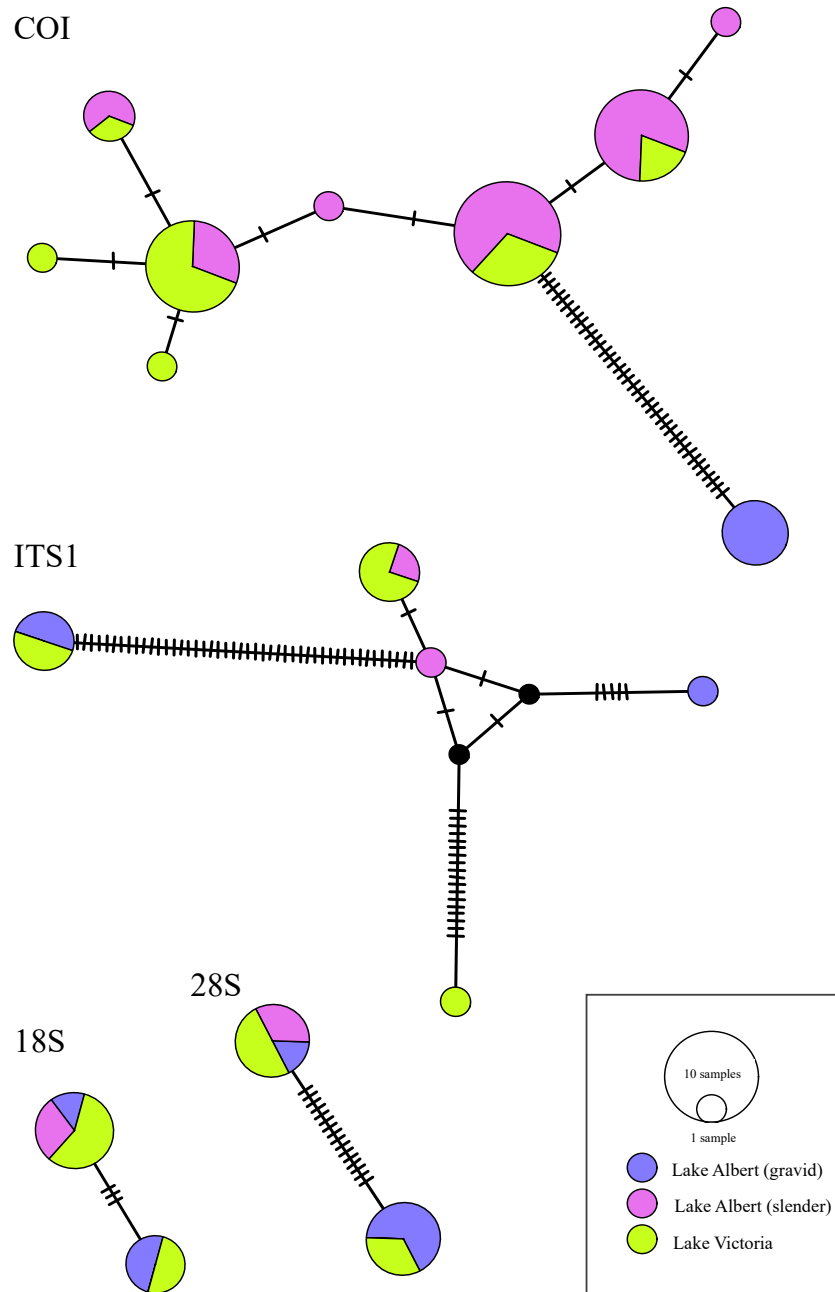
369 *Genetic diversity and phylogeography*

370 We found two haplotypes for the 18S rDNA marker (443 bp), two haplotypes for the 28S rDNA marker (778 bp),
371 five haplotypes for the ITS-1 marker (493 bp), and nine haplotypes in the COI marker (325 bp). Uncorrected p-
372 distances varied up to 0.7% for the 18S rDNA marker, and the two haplotypes were shared between lakes and
373 morphotypes. Accordingly, two haplotypes were shared between lakes and morphotypes for 28S rDNA, with
374 uncorrected p-distances up to 2.1%. For the ITS-1 marker, one haplotype was shared between lakes, and
375 uncorrected p-distances varied between 0.2 - 20.6% with a total of 17 indels between haplotypes. Between the
376 morphotypes, no haplotypes were shared for the ITS-1 marker, and uncorrected p-distances varied between 1.5 -

377 17.2%. For the mitochondrial COI marker, four haplotypes were shared between the lakes, and uncorrected p-
378 distances varied between 0.3 - 13.5%. No haplotypes were shared between the morphotypes, and uncorrected p-
379 distances varied between 12.5 - 13.5% in the COI haplotypes. As all ITS-1 haplotypes could be aligned, our
380 hypothesis that a single diplectanid species was examined is supported (Wu et al. 2007; Poisot et al. 2011). The
381 uncorrected p-distance over the COI fragment does not reach the “best-compromise threshold” (Meier et al. 2006)
382 for barcoding of 14.5% proposed by Vanhove et al. (2013), which indicates that the specimens are conspecific,
383 and all belong to *D. lacustre*.

384 We observed no genetic separation between specimens from different lakes. Genetic separation was observed
385 between the morphotypes of *D. lacustre* from Lake Albert in the COI marker, where the gravid morphotype was
386 represented by a single, exclusive haplotype (Fig. 6). There was haplotype sharing between the slender
387 morphotype and the Lake Victoria specimens in the COI marker, and three haplotypes were dominant, surrounded
388 by several satellite haplotypes, separated by single mutations. For the ITS-1 marker, a single haplotype was shared
389 between the gravid population and the Lake Victoria population. The ITS-1 marker did not show the same pattern
390 as the other nuclear markers. Marker incongruence was observed in both lakes. In Lake Albert, one specimen with
391 a gravid morphotype had a COI haplotype dominant in other individuals showing a gravid morphotype, but the
392 nuclear haplotypes were the same as the dominant haplotypes in the slender morphotype. Conversely, in Lake
393 Victoria, two specimens of the slender morphotype and COI haplotype, but with haplotypes of nuclear gene
394 portions corresponding with the gravid morphotype were characterised.

395



396

397 **Fig. 6** Haplotype networks of *Dolocirroplectanum lacustre* from Lake Albert and Lake Victoria based on
398 sequences of COI mtDNA (n=45), ITS1 rDNA (n=11), 28S rDNA (n=12), 18S rDNA (n=11) combining data
399 generated during this study and by Kmentová et al. (2020a). Circles represent different haplotypes with the size
400 proportional to the number of individuals sharing the haplotype. Colours correspond to the different populations.
401 Mutation steps as hatch marks.

402

403 The genetic diversity in the COI mtDNA gene portion was largest in Lake Albert (Table 3). The number of
404 polymorphic sites in COI was 46 (n=29) in Lake Albert, and six (n=16) in Lake Victoria. Five sites were

405 polymorphic (n=24) in the slender morphotype, and for the gravid morphotype only a single haplotype was
 406 represented in five individuals. We observed similar levels of nucleotide and haplotype diversity between *D.*
 407 *lacustre* in both lakes, however values for these diversity measures were larger in Lake Albert. F_{ST} values were
 408 significant between the Lake Albert morphotypes ($F_{ST} = 0.97049$, p-value = 0.00000). Likewise, the gravid
 409 morphotype and the specimens from Lake Victoria were genetically differentiated ($F_{ST} = 0.97024$, p-value =
 410 0.00000). Conversely, low F_{ST} values ($F_{ST} = 0.16794$, p-value = 0.01802) for the Lake Victoria specimens and the
 411 slender specimens from Lake Albert indicate higher genetic similarity between these populations. Following
 412 AMOVA between specimens from Lake Albert and Lake Victoria ($F_{ST} = 0.05983$, p-value = 0.03812), most of
 413 the variation was present within the lakes (94.02%) in comparison with 5.98% among-lake variation.

414

415 *Demographic history*

416 No signatures of population expansion could be detected for *D. lacustre* from the studied lakes. Neutrality test
 417 statistics were not significant for Lake Albert alone (Tajima's $D = 0.537$, $p = 0.774$; Fu's $F_S = 9.872$, $p = 0.995$),
 418 as for Lake Albert and Lake Victoria combined (Tajima's $D = -0.230$, $p = 0.453$; Fu's $F_S = 7.137$, $p = 0.981$).

419

420 *The gill microhabitats*

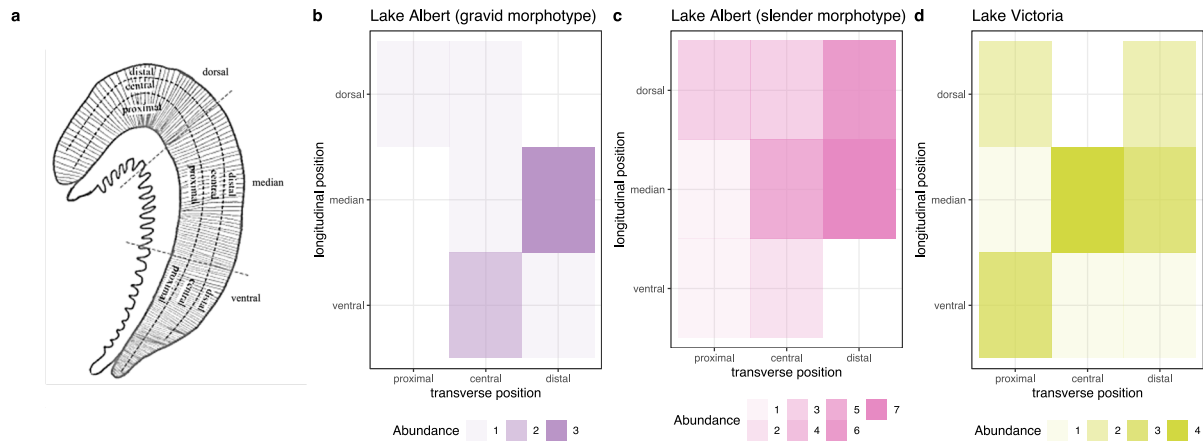
421 Individuals of *D. lacustre* were predominantly attached to the median-central and median-distal microhabitats on
 422 the gills (Fig. 7). In Lake Victoria, parasites were mostly retrieved from the right gill chamber (2 left, 18 right).
 423 In Lake Albert, most specimens were retrieved from the left gill chamber (slender: 50 left, 22 right; gravid: 8 left).
 424 In co-infections, both morphotypes occupied the same microhabitats, but several specimens of the gravid
 425 morphotype were mostly found on the median-ventral portion of the gill (Fig. 7). Whereas specimens from Lake
 426 Victoria were attached to the proximal-ventral portion of the gill, this was uncommon in Lake Albert.

427

428 **Table 3.** Genetic diversity indices of *Dolicirroplectanum lacustre* in Lake Albert (per population) and Lake
 429 Victoria inferred from a 325 bp portion of the mitochondrial cytochrome *c* oxidase subunit I (COI) region.

Population	<i>n</i>	H	S	Maximum uncorrected p-distance	Hd	π
Lake Albert	29	7	46	13.2 %	0.8079 ± 0.0405	0.041228 ± 0.021246
Lake Victoria	16	6	6	1.2 %	0.7667 ± 0.0839	0.004949 ± 0.003462
Lake Albert (slender)	24	6	5	1.5 %	0.7536 ± 0.0562	0.004515 ± 0.003159
Lake Albert (gravid)	5	1	0	/	/	/

430 *n*, sample size; H, number of haplotypes; S, number of polymorphic sites; Hd, haplotype diversity; π , nucleotide diversity



431

432 **Fig. 7.** Microhabitat distribution of *Dolocirroplectanum lacustre* visualised in heat plots. (a) division of the fish
 433 gill into nine microhabitats across a longitudinal and transversal axis; (b) Lake Victoria; (c) gravid morphotype
 434 from Lake Albert; (d) slender morphotype from Lake Albert.

435

436 Discussion

437 In Lake Albert we morphologically identified two distinct morphotypes of *Dolocirroplectanum lacustre*, a
 438 monogenean parasite infecting latid fishes across the African freshwater system. A single morphotype (slender)
 439 is suggested to be co-introduced in Lake Victoria with the historical introductions of the Nile perch. The variation
 440 in haptor structures of the slender morphotype coincides with the variation in specimens from Lake Victoria,
 441 with the latter showing slightly less phenotypic and genetic variation. We observed spatial similarity in the
 442 microhabitat distribution between the specimens of the slender morphotype from both lakes. Co-infections with
 443 both morphotypes were observed in Lake Albert. The two morphotypes were genetically separated in the COI
 444 mtDNA region, but genetic differentiation did not meet the level typically associated with species-delineation.
 445 Patterns in the ITS-1 region differed in the slender specimens between Lake Victoria and Lake Albert. Mitonuclear
 446 marker incongruence was observed within the morphotypes of *D. lacustre*.

447

448 *A host species harbours the greatest diversity of parasites in the area where it resided longest*

449 The haplotype sharing between studied locations supports the co-introduction of *D. lacustre* with *L. niloticus* to
 450 Lake Victoria from Lake Albert (Pringle 2005). This agrees with host data provided by Hauser et al. (1998), which
 451 confirmed the presence of *L. niloticus* from Lake Albert in Lake Victoria and excluded the successful
 452 establishment of introduced Nile perch from Lake Turkana. We observed large variation in haptor morphology
 453 in Lake Albert (native range) and identified the two morphotypes as in earlier studies (Thurstun and Paperna 1969;

454 Kmentová et al. 2020a). The presence of different species of Nile perch (Hauser et al. 1998), that occur in sympatry
455 in Lake Albert may have provided *D. lacustre* with different gill habitats and could account for the morphological
456 variation between morphotypes in Lake Albert. Under this scenario, the observations of co-infections would
457 require secondary contact and host switching or hybridisation between the two host taxa. The reduced phenotypic
458 variation (and presence of a single morphotype) in Lake Victoria may be an indication for a founder effect. In
459 accordance, we have observed reduced genetic diversity in the specimens from Lake Victoria in comparison with
460 Lake Albert. When relatively few parasite specimens were co-introduced, a ‘founder effect’ would be expected
461 (Mayr 1942). Accordingly, a host species is expected to harbour the greatest diversity of parasites in the area
462 where it resided longest (Manter 1966). This argument is supported by the low number of introduced hosts (n =
463 382) into Lake Victoria (Hauser et al. 1998).

464

465 *Phenotypic plasticity: discordant variation and mitonuclear discordance*

466 The observed variation in the morphology and resulting morphotypes, in combination with below species-level
467 genetic differentiation, suggests phenotypic plasticity of the species *D. lacustre* following the definition of DeWitt
468 et al. (1998): species with “the potential to produce a range of different, relatively fit phenotypes in multiple
469 environments”. Discordance between molecular and morphological differentiation has been observed for other
470 diplectanid monogenean species, namely *Lamellodiscus* spp. (Diplectanidae) (Desdevises et al. 2000; Poisot et al.
471 2011). Following Desdevises et al. (2000), these inconsistencies could be attributed to phenotypic variability of
472 some characters related to the environment and the host (Villar-Torres et al. 2019). Moreover, the observed
473 phenotypic variability and the genetic structuring of the morphotypes observed in the COI marker could be an
474 indication for ongoing sympatric speciation or diversification processes (Poulin 2002) of *D. lacustre* in Lake
475 Albert. Parasites with direct life cycles, like monogeneans, are expected to experience more favourable conditions
476 for sympatric speciation, especially when infecting long-living hosts (Brooks and McLennan 1993), such as the
477 Nile perch. Similarly, the observed morphological differentiation in the MCO could be an indication of ongoing
478 diversification or incipient speciation (Šimková et al. 2002), explained by the ‘reinforcement hypothesis’ (Rohde
479 1991): “reinforcement of reproductive barriers is one of the main factors resulting in niche segregation”.

480 Although the presented population structure based on the COI marker indeed agrees with the reproductive
481 isolation of the gravid morphotype, the mitonuclear discordance gives a different picture, and indicates an
482 incomplete reproductive barrier between the two morphotypes. Potentially, this discordance may be explained by
483 intraspecific gene flow by hybridisation between the morphotypes in Lake Albert, or incomplete lineage sorting

484 (Després 2019) the former has earlier been linked with the evolution of invasiveness (Culley and Hardiman 2008).
485 Therefore, the observed mitonuclear discordance in Lake Victoria deserves particular attention in future research.

486

487 *A 'failure to diverge' in Lake Tanganyika versus ongoing diversification in Lake Albert*

488 Although former studies identified two different morphotypes of *D. lacustre* in Lake Albert, no genetic
489 differentiation between the morphotypes had been investigated. In Kmentová et al. (2020a), a 'failure to diverge'
490 was observed for *D. lacustre*, as the species had not diversified over the variety of African freshwater systems,
491 even when infecting different host species in Lake Tanganyika. This was hypothesised to be attributed to a low
492 rate of molecular evolution. Our data supports the severely reduced genetic diversity in Lake Tanganyika in
493 comparison to other studied areas (Table 5). Moreover, the haplotype network for COI mtDNA indicates that the
494 gravid lineage from Lake Albert is genetically more distinctive from the slender lineage in Lake Albert and Lake
495 Victoria, compared to the specimens from Lake Tanganyika (Fig. 8). The reduced diversification in *D. lacustre*
496 from Lake Tanganyika may be explained by the recently estimated timing of colonisation by *L. niloticus* and
497 divergence (1.27-1.76 MYA) of the four extant *Lates* spp. in Lake Tanganyika (Koblmüller et al. 2021). The
498 observation of greater morphological and genetic diversity in Lake Albert agrees with this hypothesis.

499

500 *Future perspectives*

501 Although this study sheds light on the population-scale differentiation and the co-introduction of *D. lacustre*,
502 several questions on its co-introduction and diversification mechanisms remain. We suggest: (1) to investigate
503 whether the diversification patterns after co-introduction, observed in this study, occur across different freshwater
504 systems, we should increase the number of samples over various sampling sites within the entire native range
505 (including Lake Turkana) and introduced range (Lake Kyoga) of Nile perch. By sample acquisition from historical
506 collections, broader spatial and temporal patterns of diversification after co-introduction could be examined; (2)
507 with sample collection note should be taken of the host (species) identity, in particular in Lakes Albert and
508 Turkana, where two sympatric species could occur, since host identity may contribute to the observed
509 diversification patterns of *D. lacustre*; (3) the application of Next Generation Sequencing approaches to study
510 genome-wide intraspecific variation in *D. lacustre*, in order to identify the determinants of the observed diversity
511 and mitonuclear discordance; (4) to undertake a more detailed study on the gill environment of Nile perch and
512 microhabitat selection of *D. lacustre* in Lake Albert to determine whether microhabitat preference plays a role in
513 the ongoing morphological and genetic diversification and possible sympatric speciation of *D. lacustre*.

514 **Acknowledgments**

515 The authors would like to thank the Zoology Research group: Biodiversity & Toxicology at Hasselt University.
516 Natascha Steffanie is acknowledged for sample processing, and Armando Cruz Laufer is thanked for laboratory
517 assistance. We acknowledge the Royal Belgian Institute of Natural Sciences and the Royal Museum for Central
518 Africa for sample provision.

519

520 **References**

- 521 Anderson RM, May RM (1991) Infectious diseases of humans: Dynamics and control. Oxford University Press
- 522 Bandelt HJ, Forster P, Rohl A (1999) Median-joining networks for inferring intraspecific phylogenies.
523 *Molecular Biology and Evolution* 16:37–48. <https://doi.org/10.1093/oxfordjournals.molbev.a026036>
- 524 Brooks DR, McLennan DA (1993) Parascript: Parasites and the language of evolution. Smithsonian series in
525 comparative evolutionary biology. *Smithsonian Institution Press* Washington
- 526 Buchmann K, Lindenstrøm T (2002) Interactions between monogenean parasites and their fish hosts.
527 *International Journal for Parasitology* 32:309–319. [https://doi.org/10.1016/s0020-7519\(01\)00332-0](https://doi.org/10.1016/s0020-7519(01)00332-0)
- 528 Buchmann K, Uldal A (1997) *Gyrodactylus derjavini* infections in four salmonids: comparative host
529 susceptibility and site selection of parasites. *Diseases of Aquatic Organisms* 28:201–209.
530 <https://doi.org/10.3354/dao028201>
- 531 Bullard S, Benz G, Overstreet R, et al (2000) Six new host records and an updated list of wild hosts for
532 *Neobenedenia melleni* (MacCallum) (Monogenea: Capsalidae). *Journal of the Helminthological Society of*
533 *Washington* 67:190–196
- 534 Culley TM, Hardiman NA (2008) The role of intraspecific hybridization in the evolution of invasiveness: A case
535 study of the ornamental pear tree *Pyrus calleryana*. *Biological Invasions* 11:1107–1119.
536 <https://doi.org/10.1007/s10530-008-9386-z>
- 537 Desdevises Y, Jovelin R, Jousson O, Morand S (2000) Comparison of ribosomal DNA sequences of
538 *Lamellodiscus* spp. (Monogenea, Diplectanidae) parasitising *Pagellus* (Sparidae, Teleostei) in the North
539 Mediterranean Sea: Species divergence and coevolutionary interactions. *International Journal for*
540 *Parasitology* 30:741–746. [https://doi.org/10.1016/s0020-7519\(00\)00051-5](https://doi.org/10.1016/s0020-7519(00)00051-5)
- 541 Després L (2019) One, two or more species? Mitonuclear discordance and species delimitation. *Molecular*
542 *Ecology* 28:3845–3847. <https://doi.org/10.1111/mec.15211>

- 543 Dewey Dunnington (2021). ggspatial: Spatial Data Framework for ggplot2. R package version 1.1.5.
544 <https://CRAN.R-project.org/package=ggspatial>
- 545 DeWitt TJ, Sih A, Wilson DS (1998) Costs and limits of phenotypic plasticity. *Trends in Ecology & Evolution*
546 13:77–81. [https://doi.org/10.1016/s0169-5347\(97\)01274-3](https://doi.org/10.1016/s0169-5347(97)01274-3)
- 547 Dlugosch KM, Parker IM (2008) Founding events in species invasions: Genetic variation, adaptive evolution,
548 and the role of multiple introductions. *Molecular Ecology* 17:431–449. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-294x.2007.03538.x)
549 294x.2007.03538.x
- 550 Edgar RC (2004) MUSCLE: Multiple sequence alignment with high accuracy and high throughput. *Nucleic*
551 *Acids Research* 32:1792–1797. <https://doi.org/10.1093/nar/gkh340>
- 552 Ergens R, Lom J (1970) Causative Agents of Fish Diseases. Academia
- 553 Excoffier L, Lischer HEL (2010) Arlequin suite ver 3.5: a new series of programs to perform population
554 genetics analyses under Linux and Windows. *Molecular Ecology Resources* 10:564–567.
555 <https://doi.org/10.1111/j.1755-0998.2010.02847.x>
- 556 Fox, J., Weisberg, S. (2019). An {R} Companion to Applied Regression, Third Edition. Thousand Oaks CA:
557 Sage. URL:<https://socialsciences.mcmaster.ca/jfox/Books/Companion/>
- 558 Gobbin TP, Vanhove MPM, Seehausen O, Maan ME (2021) Microhabitat distributions and species interactions
559 of ectoparasites on the gills of cichlid fish in Lake Victoria, Tanzania. *International Journal for*
560 *Parasitology* 51:201–214. <https://doi.org/10.1016/j.ijpara.2020.09.001>
- 561 Goedknecht MA, Feis ME, Wegner KM, et al (2016a) Parasites and marine invasions: Ecological and
562 evolutionary perspectives. *Journal of Sea Research* 113:11–27.
563 <https://doi.org/10.1016/j.seares.2015.12.003>
- 564 Goedknecht MA, Schuster A-K, Buschbaum C, et al (2016b) Spillover but no spillback of two invasive parasitic
565 copepods from invasive Pacific oysters (*Crassostrea gigas*) to native bivalve hosts. *Biological Invasions*
566 19:365–379. <https://doi.org/10.1007/s10530-016-1285-0>
- 567 Goudswaard K (P. C), Witte F, Katunzi EFB (2006) The invasion of an introduced predator, Nile perch (*Lates*
568 *niloticus*, L.) in Lake Victoria (East Africa): Chronology and causes. *Environmental Biology of Fishes*
569 81:127–139. <https://doi.org/10.1007/s10641-006-9180-7>
- 570 Hassouna N, Mithot B, Bachellerie J-P (1984) The complete nucleotide sequence of mouse 28S rRNA gene.
571 Implications for the process of size increase of the large subunit rRNA in higher eukaryotes. *Nucleic Acids*
572 *Research* 12:3563–3583. <https://doi.org/10.1093/nar/12.8.3563>

- 573 Hauser L, Carvalho GR, Pitcher TJ, Ogutu-ohwayo R (1998) Genetic affinities of an introduced predator: Nile
574 perch in Lake Victoria, East Africa. *Molecular Ecology* 7:849–857. [https://doi.org/10.1046/j.1365-](https://doi.org/10.1046/j.1365-294x.1998.00399.x)
575 [294x.1998.00399.x](https://doi.org/10.1046/j.1365-294x.1998.00399.x)
- 576 Holden MJ (1967) The systematics of the genus *Lates* (Teleosti: Centropomidae) in Lake Albert, East Africa.
577 *Journal of Zoology* 151:329–342. <https://doi.org/10.1111/j.1469-7998.1967.tb02119.x>
- 578 Jombart, T., 2008. Adegnet: a R package for the multivariate analysis of genetic markers. *Bioinformatics* 24,
579 1403–1405. <https://doi.org/10.1093/bioinformatics/btn129>
- 580 Justine J-L, Henry É (2010) Monogeneans from *Epinephelus chlorostigma* (Val.) (Perciformes: Serranidae) off
581 New Caledonia, with the description of three new species of diplectanids. *Systematic Parasitology* 77:81–
582 105. <https://doi.org/10.1007/s11230-010-9263-x>
- 583 Kassambara, A., Mundt, F., 2017. factoextra: Extract and Visualize the Results of Multivariate Data Analyses.
- 584 Kearn GC (1973) An endogenous circadian hatching rhythm in the monogenean skin parasite *Entobdella soleae*,
585 and its relationship to the activity rhythm of the host (*Solea solea*). *Parasitology* 66:101–122.
586 <https://doi.org/10.1017/s0031182000044486>
- 587 Kmentová N, Gelnar M, Mendlová M, et al (2016) Reduced host-specificity in a parasite infecting non-littoral
588 Lake Tanganyika cichlids evidenced by intraspecific morphological and genetic diversity. *Scientific*
589 *Reports* 6. <https://doi.org/10.1038/srep39605>
- 590 Kmentová N, Koblmüller S, van Steenberge M, et al (2020a) Failure to diverge in African Great Lakes: The
591 case of *Dolicirroplectanum lacustre* gen. nov. comb. nov. (Monogenea, Diplectanidae) infecting latid
592 hosts. *Journal of Great Lakes Research* 46:1113–1130. <https://doi.org/10.1016/j.jglr.2019.09.022>
- 593 Kmentová N, Koblmüller S, van Steenberge M, et al (2020b) Weak population structure and recent
594 demographic expansion of the monogenean parasite *Kapentagyryus* spp. infecting clupeid fishes of Lake
595 Tanganyika, East Africa. *International Journal for Parasitology* 50:471–486.
596 <https://doi.org/10.1016/j.ijpara.2020.02.002>
- 597 Kmentová N, van Steenberge M, Thys van den Audenaerde DFET, et al (2018) Co-introduction success of
598 monogeneans infecting the fisheries target *Limnothrissa miodon* differs between two non-native areas:
599 The potential of parasites as a tag for introduction pathway. *Biological Invasions* 21:757–773.
600 <https://doi.org/10.1007/s10530-018-1856-3>

- 601 Koblmüller S, Schöggel CA, Lorber CJ, et al (2021) African lates perches (Teleostei, Latidae, Lates): Paraphyly
602 of Nile perch and recent colonization of Lake Tanganyika. *Molecular Phylogenetics and Evolution*
603 160:107141. <https://doi.org/10.1016/j.ympev.2021.107141>
- 604 Koskivaara M, Valtonen ET, Vuori K-M (1992) Microhabitat distribution and coexistence of *Dactylogyrus*
605 species (Monogenea) on the gills of roach. *Parasitology* 104:273–281.
606 <https://doi.org/10.1017/s0031182000061710>
- 607 Lê, S., Josse, J. & Husson, F. (2008). FactoMineR: An R Package for Multivariate Analysis. *Journal of*
608 *Statistical Software*. 25(1). pp. 1-18.
- 609 Leigh JW, Bryant D (2015) popart: Full-feature software for haplotype network construction. *Methods in*
610 *Ecology and Evolution* 6:1110–1116. <https://doi.org/10.1111/2041-210x.12410>
- 611 Littlewood DTJ, Rohde K, Clough KA (1997) Parasite speciation within or between host species? Phylogenetic
612 evidence from site-specific polystome monogeneans. *International Journal for Parasitology* 27:1289–
613 1297. [https://doi.org/10.1016/s0020-7519\(97\)00086-6](https://doi.org/10.1016/s0020-7519(97)00086-6)
- 614 Lockyer AE, Olson PD, Østergaard P, et al (2003) The phylogeny of the Schistosomatidae based on three genes
615 with emphasis on the interrelationships of *Schistosoma Weinland*, 1858. *Parasitology* 126:203–224.
616 <https://doi.org/10.1017/s0031182002002792>
- 617 Lowe, S.M., Browne, M., Boudjelas, S., De Poorter, Maj. (2000). 100 of the World's Worst Invasive Alien
618 Species: A Selection from the Global Invasive Species Database. Published by The Invasive Species
619 Specialist Group (ISSG) a specialist group of the Species Survival Commission (SSC) of the World
620 Conservation Union (IUCN), 12pp. First published as special lift-out in *Aliens*.
- 621 Lymbery AJ, Morine M, Kanani HG, et al (2014) Co-invaders: The effects of alien parasites on native hosts.
622 *International Journal for Parasitology: Parasites and Wildlife* 3:171–177.
623 <https://doi.org/10.1016/j.ijppaw.2014.04.002>
- 624 Manter HW 1966. Parasites of fishes as biological indicators of recent and ancient conditions. In *Host-parasite*
625 *relationships* JE McCauley ed. Oregon State University Press, Corvallis: 59-72
- 626 Mayr E (1942) *Systematics and the Origin of Species*. Columbia University Press
- 627 Meier R, Shiyang K, Vaidya G, Ng PKL (2006) DNA Barcoding and Taxonomy in Diptera: A Tale of High
628 Intraspecific Variability and Low Identification Success. *Systematic Biology* 55:715–728.
629 <https://doi.org/10.1080/10635150600969864>

- 630 Nieberding C, Morand S, Libois R, Michaux JR (2004) A parasite reveals cryptic phylogeographic history of its
631 host. *Proceedings of the Royal Society of London Series B: Biological Sciences* 271:2559–2568.
632 <https://doi.org/10.1098/rspb.2004.2930>
- 633 Nieberding CM, Olivieri I (2007) Parasites: Proxies for host genealogy and ecology? *Trends in Ecology &*
634 *Evolution* 22:156–165. <https://doi.org/10.1016/j.tree.2006.11.012>
- 635 Odening K (1976) Conception and Terminology of Hosts in Parasitology. pp 1–93
- 636 Ogutu-Ohwayo R, Hecky RE (1991) Fish Introductions in Africa and Some of Their implications. *Canadian*
637 *Journal of Fisheries and Aquatic Sciences* 48:8–12. <https://doi.org/10.1139/f91-299>
- 638 Ondračková M, Matějusková I, Grabowska J (2012) Introduction of *Gyrodactylus perccotti* (Monogenea) into
639 Europe on its invasive fish host, Amur sleeper (*Percottus glenii*, Dybowski 1877). *Helminthologia*
640 49:21–26. <https://doi.org/10.2478/s11687-012-0004-3>
- 641 Pebesma, E., 2018. Simple Features for R: Standardized Support for Spatial Vector Data. *The R Journal* 10 (1),
642 439-446, <https://doi.org/10.32614/RJ-2018-009>
- 643 Peeler EJ, Gardiner R, Thrush MA (2004) Qualitative risk assessment of routes of transmission of the exotic fish
644 parasite *Gyrodactylus salaris* between river catchments in England and Wales. *Preventive Veterinary*
645 *Medicine* 64:175–189. <https://doi.org/10.1016/j.prevetmed.2004.05.005>
- 646 Poisot T, Verneau O, Desdevises Y (2011) Morphological and molecular evolution are not linked in
647 *Lamellodiscus* (plathyhelminthes, monogenea). *PLoS ONE* 6:e26252.
648 <https://doi.org/10.1371/journal.pone.0026252>
- 649 Poulin R (2002) The evolution of monogenean diversity. *International Journal for Parasitology* 32:245–254.
650 [https://doi.org/10.1016/s0020-7519\(01\)00329-0](https://doi.org/10.1016/s0020-7519(01)00329-0)
- 651 Prenter J, MacNeil C, Dick JTA, Dunn AM (2004) Roles of parasites in animal invasions. *Trends in Ecology &*
652 *Evolution* 19:385–390. <https://doi.org/10.1016/j.tree.2004.05.002>
- 653 Pringle RM (2005) The origins of the Nile perch in Lake Victoria. *BioScience* 55:780.
654 [https://doi.org/10.1641/0006-3568\(2005\)055\[0780:tootnp\]2.0.co;2](https://doi.org/10.1641/0006-3568(2005)055[0780:tootnp]2.0.co;2)
- 655 Pringle RM (2011) Nile Perch. In: *Encyclopedia of Biological Invasions*. Univ of California Press
- 656 R Core Team (2020). R: A language and environment for statistical computing. R. Foundation for Statistical
657 Computing, Vienna, Austria.

- 658 Raymond KMN, Chapman LL, Lanciani CA (2006) Host, macrohabitat, and microhabitat specificity in the gill
659 parasite *Afrodiplozoon polycotyleus* (Monogenea). *Journal of Parasitology* 92:1211–1217.
660 <https://doi.org/10.1645/ge-621r.1>
- 661 Reynolds JE, Gréboval DF, Mannini P (1995) Thirty years on: The development of the Nile perch fishery in
662 Lake Victoria. In: *The Impact of Species Changes in African Lakes*. Springer Netherlands, pp 181–214
- 663 Rohde K (1991) Intra- and interspecific interactions in low density populations in resource-rich habitats. *Oikos*
664 60:91. <https://doi.org/10.2307/3544997>
- 665 RStudio Team (2020). RStudio: Integrated Development for R. RStudio, PBC, Boston, MA URL
666 <http://www.rstudio.com/>
- 667 Sakai AK, Allendorf FW, Holt JS, et al (2001) The population biology of invasive species. *Annual Review of*
668 *Ecology and Systematics* 32:305–332. <https://doi.org/10.1146/annurev.ecolsys.32.081501.114037>
- 669 Seehausen O, Alphen JJM van, Witte F (1997) Cichlid fish diversity threatened by eutrophication that curbs
670 sexual selection. *Science* 277:1808–1811. <https://doi.org/10.1126/science.277.5333.1808>
- 671 Šimková A, Morand S, Jobet E, et al (2004) Molecular phylogeny of congeneric monogenean parasites
672 (*Dactylogyrus*): a case of intrahost speciation. *Evolution* 58:1001. <https://doi.org/10.1554/03-606>
- 673 Šimková A, Ondračková M, Gelnar M, Morand S (2002) Morphology and coexistence of congeneric
674 ectoparasite species: Reinforcement of reproductive isolation? *Biological Journal of the Linnean Society*
675 76:125–135. <https://doi.org/10.1111/j.1095-8312.2002.tb01719.x>
- 676 Šimková A, Plaisance L, Matějusková I, et al (2003) Phylogenetic relationships of the Dactylogyridae
677 Bychowsky, 1933 (Monogenea: Dactylogyridea): The need for the systematic revision of the
678 Ancyrocephalinae Bychowsky, 1937. *Systematic Parasitology* 54:1–11.
679 <https://doi.org/10.1023/a:1022133608662>
- 680 Šimková A, Řehulková E, Rasoloariniaina JR, et al (2018) Transmission of parasites from introduced tilapias: A
681 new threat to endemic Malagasy ichthyofauna. *Biological Invasions* 21:803–819.
682 <https://doi.org/10.1007/s10530-018-1859-0>
- 683 Sinnappah ND, Lim L-HS, Rohde K, et al (2001) A paedomorphic parasite associated with a neotenic
684 amphibian host: Phylogenetic evidence suggests a revised systematic position for sphyranuridae within
685 anuran and turtle polystomatoineans. *Molecular Phylogenetics and Evolution* 18:189–201.
686 <https://doi.org/10.1006/mpev.2000.0877>
- 687 South, A. (2021). *rnaturalearth*: World Map Data from Natural Earth. <https://docs.ropensci.org/rnaturalearth>

- 688 Taraschewski H (2006) Hosts and parasites as aliens. *Journal of Helminthology* 80:99–128.
689 <https://doi.org/10.1079/joh2006364>
- 690 Thurston JP, Paperna I (1969) *Diplectanum lacustris* sp.nov. (Dactylogyroidea: Diplectanidae), a monogenetic
691 trematode from the gills of the Nile perch. *Proceedings of the Helminthological Society of Washington*
692 36:214–218
- 693 van Zwieten PAM, Kolding J, Plank MJ, et al (2016) The Nile perch invasion in Lake Victoria: Cause or
694 consequence of the haplochromine decline? *Canadian Journal of Fisheries and Aquatic Sciences* 73:622–
695 643. <https://doi.org/10.1139/cjfas-2015-0130>
- 696 Vanhove M, Tessens B, Schoelinck C, et al (2013) Problematic barcoding in flatworms: A case-study on
697 monogeneans and rhabdocoels (Platyhelminthes). *ZooKeys* 365:355–379.
698 <https://doi.org/10.3897/zookeys.365.5776>
- 699 Vanhove MPM, Pariselle A, van Steenberge M, et al (2015) Hidden biodiversity in an ancient lake:
700 Phylogenetic congruence between Lake Tanganyika trophic cichlids and their monogenean flatworm
701 parasites. *Scientific Reports* 5: <https://doi.org/10.1038/srep13669>
- 702 Venables, W. N. & Ripley, B. D. (2002) *Modern Applied Statistics with S*. Fourth Edition. Springer, New York.
703 ISBN 0-387-95457-0
- 704 Villar-Torres M, Repullés-Albelda A, Montero FE, et al (2019) Neither *Diplectanum* nor specific: A dramatic
705 twist to the taxonomic framework of *Diplectanum* (Monogenea: Diplectanidae). *International Journal for*
706 *Parasitology* 49:365–374. <https://doi.org/10.1016/j.ijpara.2018.11.003>
- 707 Whittington ID, Horton MA (2022) A revision of *Neobenedenia* Yamaguti, 1963 (Monogenea: Capsalidae)
708 including a redescription of *N. melleni* (MacCallum, 1927) Yamaguti, 1963. *Journal of Natural History*
709 30:1113–1156. <https://doi.org/10.1080/00222939600770611>
- 710 Wickham H (2007). “Reshaping Data with the reshape Package.” *Journal of Statistical Software*, 21(12), 1–20.
711 <http://www.jstatsoft.org/v21/i12/>.
- 712 Wickham H (2021). tidy: Tidy Messy Data. R package version 1.1.3. <https://CRAN.Rproject.org/package=tidy>
- 713 Wickham, H (2016). ggplot2: Elegant Graphics for Data Analysis. Springer-Verlag New York. ISBN 978-3-
714 319-24277-4, <https://ggplot2.tidyverse.org>.
- 715 Woo PTK (2006) *Fish diseases and disorders*. CABI

716 Wu X-Y, Zhu X-Q, Xie M-Q, Li A-X (2007) The evaluation for generic-level monophyly of Ancyrocephalinae
717 (Monogenea, Dactylogyridae) using ribosomal DNA sequence data. *Molecular Phylogenetics and*
718 *Evolution* 44:530–544. <https://doi.org/10.1016/j.ympev.2007.03.025>

719

720 **Statements & Declarations**

721 **Funding**

722 Study supported by Special Research Fund (BOF) UHasselt: BOF21DOC08 (KJMT), BOF20TT06 (MPMV), and
723 BOF21PD01 (NK); by Czech Science Foundation standard project GA19-13573S; and by Research Foundation
724 – Flanders (FWO-Vlaanderen) research grant 1513419N; infrastructure funded by EMBRC Belgium – FWO
725 project GOH3817N. The funders had no role in study design, data collection and analysis, decision to publish, or
726 preparation of the manuscript. We thank the Royal Belgian Institute of Natural Sciences and Royal Museum for
727 Central Africa for sample collection under the BELSPO Brain project, HIPE (BR/154/A1/HIPE).

728

729 **Author contributions**

730 All authors contributed to the study conception and design. Material preparation, and data collection were
731 performed by Kelly J.M. Thys, Jonas W.J. Custers, Nikol Kmentová, Nathan Vranken, and Maarten Van
732 Steenberge. Kelly J.M. Thys and Jonas W.J. Custers conducted data analysis. The original draft of the manuscript
733 was written by Kelly J.M. Thys, and revision of the previous versions of the manuscript was provided by Nikol
734 Kmentová, Maarten Van Steenberge, and Maarten P.M. Vanhove. Funding was acquired by Kelly J.M. Thys,
735 Nikol Kmentová, Maarten Van Steenberge, and Maarten P.M. Vanhove. All authors read and approved the final
736 manuscript.

737

738 **Compliance with Ethical Standards**

739 The authors declare that they have no competing interests.

740

741 **Data availability**

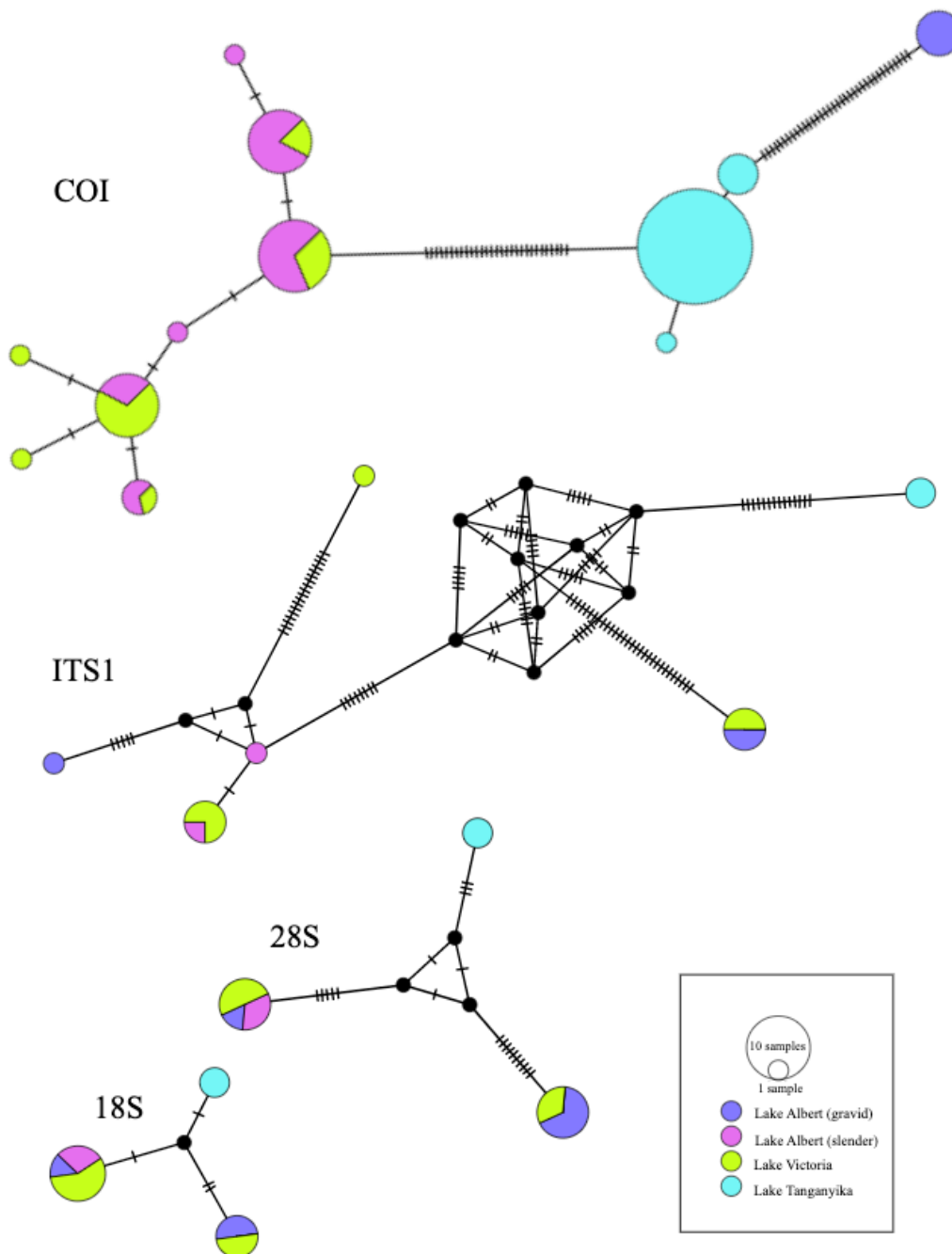
742 All data generated or analysed during this study are included in this published article [and its supplementary
743 information files].

744

745

746 **Supplementary data**

747



748

749

750 **Fig. 8** Haplotype networks of *Dolicirroplectanum lacustre* from Lake Albert, Lake Victoria and Lake Tanganyika
751 based on COI mtDNA sequences (n=83), ITS1 rDNA (n=13), 28S rDNA (n=14), 18S rDNA (n=13) combining
752 data generated during this study and Kmentová et al. (2020a). Circles represent different haplotypes with the size
753 proportional to the number of individuals sharing the haplotype. Colours correspond to the different populations.
754 Mutation steps as hatch marks.

755 **Table 5** Genetic diversity indices of *Dolicirroplectanum lacustre* in Lake Albert (per population) and Lake
756 Victoria inferred from a 325 bp portion of the mitochondrial cytochrome c oxidase subunit I (COI) region.

Population	<i>n</i>	H	S	Maximum uncorrected p-distance	Hd	π
Lake Albert	29	7	46	13.2 %	0.8079 ± 0.0405	0.041228 ± 0.021246
Lake Victoria	16	6	6	1.2 %	0.7667 ± 0.0839	0.004949 ± 0.003462
Lake Albert slender	24	6	5	1.5 %	0.7536 ± 0.0562	0.004515 ± 0.003159
Lake Albert gravid	5	1	0	/	/	/
Lake Tanganyika	38	3	2	0.3 %	0.2404 ± 0.0858	0.000757 ± 0.000976

757 *n*, sample size; H, number of haplotypes; S, number of polymorphic sites; Hd, haplotype diversity; π , nucleotide diversity