

Sediment stress across location and morphology

1 **Title: Characterizing transcriptomic responses to sediment stress across location and**
2 **morphology in reef-building corals**

3

4 **Running head: Sediment stress across location and morphology**

5

6 Jill Ashley*¹, Hailey McKelvie², John D. Freeman², Polina Shpilker², Lauren H. Zane¹, Danielle

7 M. Becker-Polinski¹, Lenore Cowen², Robert H. Richmond³, Valérie Paul⁴, Francois Seneca⁵,

8 Hollie M. Putnam¹

9

10 **Affiliations:**

11 ¹Department of Biological Sciences, University of Rhode Island, 120 Flagg Rd, 02881, RI, USA.

12 ²Department of Computer Science, Tufts University, 177 College Ave, 02155, MA, USA.

13 ³Kewalo Marine Lab, University of Hawai'i at Mānoa, 41 Ahui St, 96813, HI, USA.

14 ⁴Smithsonian Marine Station at Fort Pierce, 701 Seaway Dr, 34949, FL, USA.

15 ⁵Centre Scientifique de Monaco, 8 Quai Antoine 1er, 98000 Monaco, Monaco.

16

17 *Corresponding author

18 jillashey@uri.edu

19

20

21

22

23

Sediment stress across location and morphology

1 **Abstract**

2 Anthropogenic activities increase sediment suspended in the water column and deposition on
3 reefs can be largely dependent on colony morphology. Massive and plating corals have a high
4 capacity to trap sediments, and active removal mechanisms can be energetically costly.
5 Branching corals trap less sediment, but are more susceptible to light limitation caused by
6 suspended sediment. Despite deleterious effects of sediments on corals, few studies have
7 examined the molecular response of corals with different morphological characteristics to
8 sediment stress. To address this knowledge gap, this study assessed the transcriptomic responses
9 of branching and massive corals in Florida and Hawai*ī* to varying levels of sediment exposure.
10 Gene expression analysis revealed a molecular responsiveness to sediments across species and
11 sites. Differentially Gene Expression (DEG) followed by Gene Ontology (GO) enrichment
12 analysis identified that branching corals had the largest transcriptomic response to sediments, in
13 developmental processes and metabolism, while significantly enriched GO terms were highly
14 variable between massive corals, despite similar morphologies. Comparison of DEGs within
15 orthogroups revealed that while all corals had DEGs in response to sediment, there was not a
16 concerted gene set response by morphology or location. These findings illuminate the species
17 specificity and genetic basis underlying coral susceptibility to sediments.

18 **Keywords:** transcriptomics, sediment, corals, RNASeq, morphology

19

20

21

22

23

Sediment stress across location and morphology

1 **Introduction**

2 Coral reefs are incredibly diverse marine ecosystems, providing numerous ecological and
3 economic services such as biodiversity, cultural value, coastal protection, fisheries, and tourism
4 (Reaka-Kudla 1997; Sumaila and Cisneros-Montemayor 2010; Costanza et al. 2014). Reef-
5 building corals form a critical nutritional symbiotic relationship with unicellular photosynthetic
6 algal endosymbionts in the family Symbiodiniaceae (LaJeunesse et al. 2018). The carbohydrates
7 produced by the algal photosynthesis are translocated to the coral to be used as its primary
8 energy source, supporting the daily respiratory carbon demand of tropical corals (Muscatine and
9 Porter 1977; Muscatine et al. 1984). This coral-algal symbiosis fuels reef productivity and
10 accretion (Roth 2014), but is sensitive to changing environmental conditions that can impact the
11 symbiosis such as light, nutrients, temperature, pH, and sediment (Hoegh-Guldberg et al. 2007;
12 Davy, Allemand, and Weis 2012). For example, under exposure to sedimentation, or the
13 downward fall of sediment from the water column toward the benthos (Schlaefer, Tebbett, and
14 Bellwood 2021), corals display reduced photosynthetic efficiency (Weber, Lott, and Fabricius
15 2006; Rushmore, Ross, and Fogarty 2021), increased respiration rates (Riegl and Branch 1995;
16 Browne et al. 2014), decreased calcification (Bak 1978), and rapid consumption of energy
17 reserves (Sheridan et al. 2014).

18
19 While sediment transport naturally occurs on reefs, suspended sediment caused by anthropogenic
20 activities such as dredging, runoff, and coastal development have increased (Rogers 1990;
21 Fabricius 2005; Erftemeijer et al. 2012). Although suspended sediment stress is becoming
22 commonplace in coastal coral reefs, corals in different locations may be exposed to differing
23 sediment stress. For example, corals residing in the Maunaloa Harbor in Maui, Hawaii are

Sediment stress across location and morphology

1 exposed to sediment types that range from fine mud (particles less than 0.62mm) to coarse
2 carbonate sediment (particles greater than 0.62mm), depending on their location in the harbor
3 (Piniak and Brown 2008). Corals in certain parts of the Great Barrier Reef regularly experience
4 turbidity, or the amount of suspended particulate matter in the water column (Flores et al. 2012),
5 that can be influenced by wave height, rainfall and season, among other factors (Brodie et al.
6 2013); however, increasing river discharges of fine sediments and nutrients (nitrogen and
7 phosphorus) have been found to significantly affect turbidity (Fabricius et al. 2013). Coastal
8 cities, like Miami, Florida, have had large-scale dredging projects over the past two decades,
9 frequently exposing reefs to high levels of suspended sand and silt (PBS&J 2008; Miller et al.
10 2016; Cuning et al. 2019).

11
12 Deposited sediment and suspended sediment are the two primary ways that sediment interacts
13 with corals (Rogers 1990; Fabricius 2005; Erftemeijer et al. 2012). Deposited sediment occurs
14 when sediment particles settle directly on the coral surface, making physical contact with the
15 tissue. Passive removal of sediment includes gravity or flow removing it from the coral surface
16 (Lasker 1980; Jones, Fisher, and Bessell-Browne 2019). In response to deposited sediment,
17 corals can also initiate an acute response to attempt to move the sediment using active
18 mechanisms. Active sediment removal mechanisms include ciliary and tentacle movement,
19 increased mucus production, and hydrostatic inflation (Rogers 1990; Stafford-Smith and Ormond
20 1992; Stafford-Smith 1993; Bessell-Browne et al. 2017). However, these active mechanisms are
21 often very energetically costly, and thus cannot be sustained for long periods of time (Riegl and
22 Branch 1995; Erftemeijer et al. 2012). If the sediment deposition rate exceeds the coral's
23 sediment clearance rate, sediment will accumulate on the coral, reducing heterotrophic feeding

Sediment stress across location and morphology

1 and light transmission to algal endosymbionts and creating hypoxic conditions near the coral
2 tissue, which often leads to tissue necrosis and coral mortality (Philipp and Fabricius 2003;
3 Weber, Lott, and Fabricius 2006; Weber et al. 2012).

4
5 Corals can also, or alternatively, interact with suspended sediment, which occurs when particles
6 such as clay, silt, and sand are moved into the water column by some natural or anthropogenic
7 disturbance and remain in the water column for a period of time (Rogers 1990; Fabricius 2005;
8 Erfteimeijer et al. 2012). Suspended sediment reduces the amount of light that reaches the coral,
9 impeding the ability of the algal endosymbionts to photosynthesize and provide the coral host
10 with sufficient energy for metabolism and growth (Rogers 1990; Fabricius 2005; Erfteimeijer et
11 al. 2012; Bessell-Browne et al. 2017). Reduced photosynthetic efficiency can induce corals to
12 switch to heterotrophic feeding, a much less efficient way to obtain carbon than through its
13 endosymbionts (Muscatine and Porter 1977; Anthony and Fabricius 2000; Houlbrèque and
14 Ferrier-Pagès 2009). Additionally, heterotrophic feeding in the presence of sediments may lead
15 the coral to ingest sediment particles, disrupting its nutritional intake and potentially acting as a
16 vector for harmful bacteria and toxins (Erfteimeijer et al. 2012; Studivan et al. 2022). Suspended
17 sediment has also been observed to induce immune responses and increase disease prevalence in
18 corals (Pollock et al. 2014; Sheridan et al. 2014).

19
20 In addition to sediment type, the morphology of the coral can modulate its interaction with
21 sediments. For example, massive, plating, and encrusting corals have a higher planar surface area
22 and thus higher capacity to trap sediments in comparison to branching corals with high three-
23 dimensional and more vertical structure. Sediment removal from massive corals often requires

Sediment stress across location and morphology

1 active removal mechanisms (Dallmeyer, Porter, and Smith 1982; Rogers 1990; Stafford-Smith
2 1993). However, massive corals may also be more resilient to high suspended sediment
3 concentrations because their greater surface area allows for increased opportunities to capture
4 light, maximizing the photosynthetic efficiency of their algal endosymbionts (Fabricius 2005;
5 Erftemeijer et al. 2012). In contrast, the relatively small surface area and vertical branches of
6 branching coral species means sediment is minimally trapped and can be more easily removed by
7 gravity or currents (Lasker 1980). Branching coral species typically have faster clearance rates
8 than non-branching species, and active removal mechanisms are required less frequently,
9 allowing branching corals to devote that energy towards other functions such as reproduction and
10 growth (Stafford-Smith 1993). While branching morphologies may be resistant to sediment
11 accumulation, they appear to be more sensitive to light limitation caused by suspended sediment
12 due to their relatively smaller surface area (Jones et al. 2020). Following a major dredging
13 project in Miami, FL, massive and encrusting species (*Montipora* and *Porites* spp.) had lower
14 clearance rates and accumulation was more common, as sediments landed on the surface of the
15 coral tissue. However, accumulation was not observed on branching species (*Acropora* and
16 *Pocillopora* spp.), despite extreme sediment levels (Jones, Fisher, and Bessell-Browne 2019).
17 Similar results were found in the evaluation of 22 Australian coral species, where branching
18 corals demonstrated significantly higher clearance rates than non-branching species, as sediment-
19 rejection efficiency (e.g., the percentage of coral tissue area cleared of sediment) may be a
20 function of spacing and diameter of branches (Stafford-Smith 1993). Additionally, studies on the
21 San Cristobal reef in Puerto Rico found that sediment accumulated on the flat parts of *Acropora*
22 *palmata* branches, while no sediment accumulation was observed on the cylindrical branches of

Sediment stress across location and morphology

1 *Acropora cervicornis* (Rogers 1983). Collectively, these studies support that morphology plays a
2 role in response to sediment stress.

3

4 While most research has focused primarily on physiological responses to sediment stress, there is
5 a small but growing body of work on gene expression of corals exposed to sediment stress. Early
6 microarray studies found that the upregulation of heat shock proteins (HSPs) occurred in
7 response to sediment stress (Wiens et al. 2000; Hashimoto et al. 2004). As a generalized stress
8 response protein, HSPs have also been implicated in coral response to other stressors, such as
9 temperature and ocean acidification (DeSalvo et al. 2008; DeSalvo et al. 2010; Kaniewska et al.
10 2012). Another general response to sediment stress is upregulation of biomarkers of oxidative
11 stress (Morgan, Edge, and Snell 2005), more specifically, thioredoxin, a protein that modulates
12 redox and cell-to-cell signaling (Tomanek 2015). A potentially more specific gene responding to
13 sediment stress is indicated by the differential expression of urokinase plasminogen activator
14 surface receptor (uPAR) transcripts in the coral *Diploria strigosa* along a
15 sedimentation/pollution gradient in Castle Harbor, Bermuda (Morgan, Edge, and Snell 2005).
16 uPAR is associated with proteolysis, wound healing and inflammation, and is hypothesized to
17 contribute to coral tissue remodeling in response to elevated levels of sedimentation (Morgan,
18 Edge, and Snell 2005). Genes related to immunity, as well as energy metabolism, were also
19 implicated in the transcriptomic response to sediment stress in corals from two locations,
20 Singapore (*Goniastrea pectinata* and *Mycedium elephantotus*) and Eilat, Israel (*G. pectinata*
21 only) using RNASeq (Bollati et al. 2021). While there were some methodological differences
22 between their experiments, shared mechanisms were identified across different species and
23 populations, demonstrating a conserved response to sediment stress across species and sediment

Sediment stress across location and morphology

1 types. However, the corals evaluated in that study, *G. pectinata* and *M. elephantotus*, both have
2 similar morphological characteristics (massive and encrusting), indicating that this response may
3 only be relevant to massive and encrusting corals (Bollati et al. 2021).

4
5 While these initial molecular analyses have provided important insights into the coral
6 transcriptomic response to sediment stress, less is known about shared molecular responses by
7 morphology and location/sediment type. To this end, our study aims to fill these knowledge gaps
8 by examining gene expression across different coral morphologies and the use of multiple
9 locations/types of sediment. Here we quantified the transcriptomic responses of corals with
10 different colony morphologies in response to different types of sediment stress. Floridian corals
11 (*Acropora cervicornis*, *Montastraea cavernosa* and *Orbicella faveolata*) were exposed to
12 sterilized white carbonate sediment for 18 days, whereas Hawaiian corals (*Montipora capitata*,
13 *Pocillopora acuta* (formally *Pocillopora damicornis*) and *Porites lobata*) were exposed to
14 unsterilized terrigenous red soil for up to 7 days. In this study, *A. cervicornis* and *P. acuta* were
15 categorized as branching corals, while *M. cavernosa*, *O. faveolata* and *P. lobata* were
16 categorized as massive corals. The morphology of *M. capitata* was considered as intermediate
17 between branching and plating, as *M. capitata* tends to form plates growing horizontally with
18 branches sprouting upward (Veron 2002).

19

20

21

22

23

Sediment stress across location and morphology

1 **Methods**

2 To characterize a broad range of responses to different kinds of sediment stress, two independent
3 experiments were performed in Florida and Hawaiꞏi (Fig. 1). In the first experiment, Hawaiian
4 corals were exposed to live terrigenous red soil for up to 7 days. In the other experiment,
5 Floridian corals were exposed to sterilized coral rubble sediment for 18 days. We acknowledge
6 that differences in experimental sediment regimes do not allow for direct statistical comparisons
7 between these two experiments at all levels. However, examining the patterns found in response
8 to sediment across two experiments and in orthogroups across taxa in response to sediment
9 provides the unique opportunity to describe these findings and test for shared transcriptomic
10 patterns in response to sediment stress.

11

12 Collections and Exposures

13 ***Hawaiꞏi***

14 All coral collections were obtained under Hawaiꞏi SAP permit (SAP 2015-48). *Montipora*
15 *capitata*, *Porites lobata* and *Pocillopora acuta* adult corals were collected in early June 2015
16 from Kāneohe Bay, Oāhu, Hawaiꞏi (21°25'59.1"N 157°47'11.1"W). One fragment was
17 collected per colony, and fragments were acclimated to tank conditions for one week post
18 collection. Following acclimation, 12 fragments per species were placed in 6-60 L outdoor tanks
19 (2 replicate tanks per treatment and 2 replicate fragments per tank for n=4 samples per treatment,
20 except *M. capitata* where n=3) at the Point Lab of the Hawaiꞏi Institute of Marine Biology
21 (HIMB). Fragments were exposed to sediment originating from a live terrigenous Hawaiian red
22 soil collected from the highest elevation point on Moku O Loē (Coconut Island). The total
23 suspended sediment (TSS) concentration was determined by the technique described in Cortes

Sediment stress across location and morphology

1 and Risk (1985). The unfiltered unsterilized terrigenous red dirt was mixed with fresh water to
2 mimic the sediment moved during a storm and extract the silt and clay (< 60 micron) used in the
3 experiment as in Bahr et al. (2020). Particles were filtered through mesh sieves to remove debris
4 and produce the final silt-clay mixture. Sediment was added from a concentrated stock at day 0
5 and again at day 4 to reach a low and a high total suspended sediment (TSS) concentrations ,
6 immediately following addition. To test the efficacy of our additions, triplicate samples were
7 taken from the experimental tanks following sediment dosing to measure the sediment load. The
8 sediment samples were filtered and oven dried on Whatman paper, and the sediment weight
9 measured with a precision scale at 150 ± 73 and 235 ± 63 mg/L, respectively. Powerheads
10 (Eheim Universal 300 L/h) were placed at the bottom of each tank to ensure water motion. The
11 temperature exposure was the natural profile of the lagoon water by HIMB during the time of the
12 experiment (Supplementary Figure 1A). Tanks received natural day/night cycles and intensity.
13 Coral fragments (n=4 replicate fragments per species), two from each of the two tanks, in each of
14 the three treatments were exposed (except *M. capitata* where n=3), and were collected and
15 preserved in liquid nitrogen at day 4 (24 samples on June 12, 2015), and day 7 (24 samples on
16 June 15, 2015) of the exposure. Samples were then stored at -80 °C until extraction.

17

18 *Florida*

19 Adult *Acropora cervicornis*, *Montastraea cavernosa*, and *Orbicella faveolata* colonies were
20 collected from Key West nursery of NOAA sanctuary, FL Keys on March 23, 2016 under
21 NOAA National Marine Sanctuaries Permit #FKNMS-2015-016 and #FKNMS-2016-017). On
22 March 28, 2016, coral colonies were transported from Key West to Fort Pierce in bins filled with
23 aerated seawater and placed in acclimation tanks at the Smithsonian Marine Station (Fort

Sediment stress across location and morphology

1 Pierce, Florida) prior to the start of the experiment. After acclimation, 15 fragments per species
2 were placed in 15-12 L outdoor tanks (3 replicate tanks per treatment and 1 fragment per tank for
3 $n=5$ samples per treatment). Fragments were then exposed to sterilized carbonate sand sediment
4 collected from Key Largo (25°08'22.0"N 80°23'37.6"W). Sediment was filtered through a 63
5 micron sieve to remove debris and obtain a fine grain mixture. Sediment was then run through
6 active carbon filters for a week in seawater to reduce chemical pollution. Sterilization of the
7 sediment stock was performed using an autoclave program for liquid (20 min at 121°C and 2.1
8 bar). The total suspended sediment (TSS) concentration was determined by the technique
9 described in Bahr et al. (2020). On the first day of exposure (May 23, 2016), sediment was added
10 from the sterilized concentrated stock (458 g/L). Sediment was added again on May 24, 25, 27,
11 and 30. There were five total suspended sediment (TSS) 0, 30, 100, 300, 1000 mg/L targets. To
12 approximately obtain these TSS concentrations, 1, 3, 10, and 30 mL of stock was added to each
13 12 L treatment tank. Powerheads (Eheim Universal 300 L/h) were placed at the bottom of each
14 tank to ensure water motion. The temperature exposure was the natural temperature profile in the
15 harbor at the Smithsonian Marine Station (Supplementary Figure 1B). Tanks received natural
16 day/night cycles and under 50% reduction shade. On June 9, coral fragments from all tanks were
17 collected and frozen in liquid nitrogen. Samples were then stored at -80 °C until extraction.

18

19 RNA extraction

20 Prior to RNA extraction, frozen coral samples were crushed with a manual hydraulic press (12
21 tons pressure) and a metal mortar and pestle chilled with liquid nitrogen. Approximately 100 mg
22 of frozen coral powder was used in the RNA extractions starting with 1 mL of TRIzol™ Reagent
23 (Invitrogen) and 100 uL of 0.1 mm ceramic beads in 2.0 mL tubes. Coral tissues were lysed with

Sediment stress across location and morphology

1 two rounds of 20 seconds at 6500 bpm on a tissue homogenizer (FastPrep), and a rest on ice of
2 30 seconds in between rounds. Tissue slurries were then incubated on ice for 5 minutes. After a
3 quick spin down to gather liquid at the bottom of the tube, 300 uL of molecular grade chloroform
4 was added to the slurry. Tubes were hand shaken for 15 seconds and rested for 3 minutes on ice.
5 Phase separation was obtained by centrifugation at 12 xg for 15 minutes at 4°C. Top aqueous
6 phase containing RNA was transferred to new 1.5 mL tubes and mixed again with 200 uL of
7 chloroform. The previous three steps were repeated. An equal volume of chilled 70% molecular
8 grade ethanol was added to the RNA in solution. From this point, the Direct-zol™ RNA
9 MiniPrep (Zymo Research; Cat# R2070) protocol with DNase treatment was followed. Total
10 RNAs were eluted in DEPC-treated water and quantified using the Qubit fluorometer.

11

RNA Sequencing

13 RNA samples were diluted to accommodate for library production starting with 100 ng of total
14 RNA. Samples were then loaded onto Neoprep cards and processed following the TruSeq
15 stranded mRNA Library Prep for NeoPrep kit (Document # 15049725 v03, Illumina) protocol.
16 Quality controlled libraries were then sequenced through HiSeq 50 cycle single read sequencing
17 v4 by the High Throughput Genomics Core Facility at the University of Utah.

18

Bioinformatic analysis

20 This workflow and data is located at <https://github.com/JillAshey/SedimentStress>. Quality
21 checks of raw and trimmed reads were performed using FASTQC (v0.11.8, Java-1.8; Andrews
22 2010) and MultiQC (v1.7, Python-2.7.15; Ewels et al. 2016). Reads that did not pass quality
23 control were trimmed with Trimmomatic (v0.30, Java-1.8; Bolger, Lohse, and Usadel 2014).

Sediment stress across location and morphology

1 Genomes, protein sequences, and transcript sequences were obtained from the following
2 locations: *Acropora cervicornis* (Baums Lab, v1.0_171209;
3 <https://usegalaxy.org/u/skitch/h/acervicornis-genome>); *Montastraea cavernosa* (Matz Lab, July
4 2018 version; https://www.dropbox.com/s/yfqefzntt896xfz/Mcavernosa_genome.tgz); *Montipora*
5 *capitata* (<http://cyanophora.rutgers.edu/montipora/>; Version 2, Stephens et al. 2021); *Orbicella*
6 *faveolata* (NCBI, assembly accession GCF_002042975.1; Prada et al. 2016); *Pocillopora acuta*
7 (http://cyanophora.rutgers.edu/Pocillopora_acuta/; Version 1, (Stephens et al. 2021); *Porites*
8 *lutea* (used to analyze *P. lobata* data in current study; <http://plut.reefgenomics.org/download/>;
9 Version 1.1, Robbins et al. 2019). We have archived all references used for this analysis at
10 <https://osf.io/8qn6c/> (doi = 10.17605/OSF.IO/8QN6C) to enable reproducible analyses for this
11 project. After trimming, reads were mapped to their respective genomes using STAR (v2.5.3;
12 (Dobin et al. 2013). The aligned read files from STAR (BAM file format) were assembled to the
13 references and count data were generated using StringTie (v2.1.1-GCCcore-7.3.0; (Pertea et al.
14 2015). Assembly quality was assessed with gffcompare (v0.11.5; (Pertea and Pertea 2020). The
15 StringTie *prepDE* python (v2.7.15; Pertea et al. 2015) script was used to generate a gene count
16 matrix.
17
18 To generate current gene ontology information for all species, functional annotation was
19 performed on all genomes using the following workflow
20 (<https://github.com/JillAshey/FunctionalAnnotation>). First, protein sequences from each species
21 were identified using BLAST (*blastp*; v2.11.0; Altschul et al. 1990) against NCBI's nr database
22 (1e-5 e-value; database accessed and updated on Oct. 12, 2021) and the Swissport database (1e-5
23 e-value; database accessed and updated on Oct. 22, 2021; Bairoch and Apweiler 1997). The

Sediment stress across location and morphology

1 XML files generated from the BLAST output were then used as input for BLAST2GO (v.5.2.5)
2 to generate gene ontology (GO) terms (Götz et al. 2008). Protein sequences were also used as
3 input to InterProScan (v5.46-81.0; Java v11.0.2), which identified homologous sequences and
4 assigned GO terms (Jones et al. 2014). Using BLAST2GO, the XML file generated from
5 InterProScan was merged with the nr and Swissprot BLAST2GO output, creating final
6 functional annotation tables, which were saved in csv format (<https://osf.io/8qn6c/>).

7 Gene expression and ontology analysis

8 All gene expression and ontology analyses were done in RStudio (v1.3.959) using v4.0.2 of R.
9 First, genes were filtered using *genefilter*'s (v1.70.0; Gentleman et al. 2021) *pOverA* function;
10 genes were retained for expression analysis only if counts were greater than or equal to 5 in at
11 least 85% of the samples, which minimizes differential expression results from low count genes
12 with lower confidence. Because different samples may have been sequenced to different depths,
13 size factors were calculated as the standard median ratio of a sample over a 'pseudosample' (for
14 each gene, the geometric mean of all samples; Anders and Huber 2010; Love, Huber, and Anders
15 2014). After confirming size factors were estimated to be less than 4, filtered gene counts were
16 normalized using *DESeq*'s (v1.28.1) *vst* function (Love, Huber, and Anders 2014). Treatment
17 was set as a factor; in the HI experiment, Time (Day 4 and Day 7) was not significant as a factor
18 and so corals sampled on different days were combined by treatment for further analysis.
19 Differential gene expression was assessed using the *DESeq* function with the Wald likelihood
20 test ratio (p-adjusted < 0.05). PCA plots with differentially expressed genes were generated using
21 the *plotPCA* function (Love, Huber, and Anders 2014).

23

Sediment stress across location and morphology

1 Gene ontology analysis was completed with GOSep (v1.40.0), which corrects for the higher-
2 confidence in differential expression as a function of gene length as follows: Genes that passed
3 the *pOverA* filter and were marked as differentially expressed genes by DESeq above were used
4 to calculate the probability weighting function using function *nullp* with the bias data being gene
5 length (Young et al. 2010). To identify category enrichment amongst differentially expressed
6 genes, the *goseq* function was performed with the Wald method (Love et al. 2010). Significantly
7 enriched GO terms from each of the biological processes (BP), molecular functions (MF), or
8 cellular components (CC) ontologies were denoted as those with an over-represented p-value <
9 0.05. BP GO term analysis is presented here; CC and MF GO term analyses can be found in the
10 supplementary material (See Supplementary Figures 2, 3, 4, 5, 6, and 7). GO term information
11 was organized under their parent GO slim terms (obtained from
12 <http://www.informatics.jax.org/gotools/data/input/map2MGIslim.txt>; accessed on April 4, 2021)
13 for qualitative comparison across species. Overlap of GO terms between species was compared
14 and visualized using ComplexUpset's (v1.3.3) *upset* function.

15

Orthology analysis

17 To test for functional similarities in response to sediment stress across taxa, all species were
18 characterized into orthogroups using OrthoFinder (v2.3.3) with dependencies Diamond
19 (v0.9.22), MCL (v14.137), FastME (v2.1.6.1) and BLAST+ (v2.8.1) and default parameters
20 (Emms and Kelly 2015, 2019). Orthogroups were filtered to those common in all species. The
21 differentially expressed genes (DEGs) in the orthogroups for each species were identified using
22 the DESeq2 results. Commonality in functional orthogroups containing DEGs were compared
23 between species and visualized using ComplexUpset's (v1.3.3) *upset* function.

Sediment stress across location and morphology

1

2 **Results**

3 Read sequencing and quality

4 Sequencing of 77 samples (n=11-15 per species; Table 1) yielded a total of 1,173,800,658 raw
5 reads with an average of $15,244,164 \pm 3,285,387$ raw reads per sample (Supplementary Table 1).
6 Quality filtering and trimming removed an average of $140,560 \pm 119,359$ reads per sample,
7 leaving an average of $15,103,605 \pm 3,290,800$ cleaned reads per sample for analysis
8 (Supplementary Table 1). Reads were aligned to the species-specific genome, and alignment
9 rates ranged from an average of 50.18% in *M. cavernosa* to an average of 70.62% in *A.*
10 *cervicornis* (Table 1; Supplementary Table 1).

11

12 Gene expression

13 For Florida species, there were 215, 62, and 8 unique differentially expressed genes DEGs in *A.*
14 *cervicornis*, *M. cavernosa*, and *O. faveolata*, respectively, between ambient and sediment
15 exposure treatments (Table 2; Supplementary Table 2). Of those DEGs, 67, 19, and 0 were
16 upregulated (Log Fold Change, LFC > 0, padj < 0.05) and 154, 44, and 8 (LFC < 0, padj < 0.05)
17 were downregulated in *A. cervicornis*, *M. cavernosa*, and *O. faveolata*, respectively (Table 2;
18 Supplementary Table 2). In PCA plots of all genes for Hawaiian species, ‘Days’ was not visually
19 separated from ‘Treatment’; therefore, ‘Days’ was dropped as an independent variable in further
20 analyses. For Hawaiian coral species, when comparing all treatments, 157, 263, and 153 unique
21 DEGs were identified in *M. capitata*, *P. acuta*, and *P. lobata*, respectively (Table 2;
22 Supplementary Table 2). 79 genes in *M. capitata*, 129 in *P. acuta* and 32 in *P. lobata* were
23 upregulated (LFC > 0, padj < 0.05), while 78 genes in *M. capitata*, 135 in *P. acuta* and 128 in *P.*

Sediment stress across location and morphology

1 *lobata* were downregulated (LFC < 0, padj < 0.05; Table 2; Supplementary Table 2).
2 Interestingly, there were more downregulated DEGs than upregulated in all species examined,
3 with the exception of *M. capitata*.
4
5 PCA of the DEGs for each species showed that all sediment treatments clustered away from the
6 control on the PC1 axis (PC1 variance explained was 51% for *A. cervicornis*, 46% for *M.*
7 *cavernosa*, 51% for *O. faveolata*, 79% for *M. capitata*, 64% for *P. acuta*, and 50% for *P. lobata*),
8 supporting a concerted gene expression response to sediment stress across taxa (Figure 2A-F).
9 There was slight separation between the mid and high treatments on the PC2 axis for both *P.*
10 *acuta* and *P. lobata*, indicating differential responses to sediment concentration (PC2 variance
11 explained was 13% for *P. acuta* and 22% for *P. lobata*; Figure 2E-F).

13 Gene ontology

14 Significantly enriched gene ontology (GO) terms were identified in the DEGs of the species in
15 the present study. In the Florida experiment, in which corals were exposed to sterilized carbonate
16 sediment for 18 days, *A. cervicornis*, *M. cavernosa*, and *O. faveolata*, had 278, 158, and 27
17 unique GO terms assigned to their DEGs, respectively (Table 2; Supplementary Table 2). *A.*
18 *cervicornis* shared GO terms related to developmental processes and signal transduction with
19 both *M. cavernosa* and *O. faveolata* (Table 3; Figure 3; Supplementary Table 3). Specifically,
20 ovarian cumulus expansion (GO:0001550), positive regulation of skeletal muscle tissue
21 development (GO:0048643) and regulation of Rho protein signal transduction (GO:0035023)
22 were shared between *A. cervicornis* and *M. cavernosa*, while chondrocyte development
23 (GO:0002063) and positive regulation of signal transduction (GO:0009967) were shared between

Sediment stress across location and morphology

1 *A. cervicornis* and *O. faveolata* (Table 3; Figure 3; Supplementary Table 3). Despite similar
2 morphologies, *M. cavernosa* and *O. faveolata* did not share any GO terms.
3
4 In the Hawaiian species that were exposed to unsterilized red terrigenous sediment, 237, 380,
5 and 198 GO terms were assigned to DEGs to *M. capitata*, *P. acuta*, and *P. lobata*, respectively
6 (Table 2; Supplementary Table 2). Grouping by GO slim term, we found that GO enrichment
7 occurred in functions related to developmental processes, protein metabolism, cell organization
8 and biogenesis, signal transduction, and stress response, among others (Table 3; Figure 3;
9 Supplementary Table 3). Only one shared GO term was found in all three of *M. capitata*, *P.*
10 *acuta*, and *P. lobata*, namely microtubule-based processes (GO:0007017; Table 3; Figure 3;
11 Supplementary Table 3). *M. capitata* and *P. acuta*, both of which were branching corals, shared
12 20 GO terms, primarily relating to developmental processes and signal transduction (Table 3;
13 Figure 3; Supplementary Table 3). This was the largest number of GO terms shared between
14 species. *P. acuta* and *P. lobata* shared functional responses related to cell organization and
15 biogenesis, cell-cell signaling, and developmental processes. Developmental processes included
16 5 shared GO terms, specifically embryonic axis specifications (GO:0000578), embryonic organ
17 development (GO:0048568), eye development (GO:0001654), neural crest cell fate specification
18 (GO:0014036), and neural plate anterior/posterior regionalization (GO:0021999). *P. acuta* had
19 the most GO terms identified across both experiments (Table 3; Figure 3; Supplementary Table
20 3). Aside from the microtubule-based process, *M. capitata* and *P. lobata* did not share other GO
21 terms.
22

Sediment stress across location and morphology

1 There was overlap of GO terms between location and morphology (Table 3; Figures 4 & 5). *M.*
2 *capitata* was present in all in common three-way interactions; it shared positive regulation of
3 skeletal muscle tissue development (GO:0048643) with *M. cavernosa* and *A. cervicornis*,
4 riboflavin transport (GO:0032218) with *M. cavernosa* and *P. acuta*, and microtubule-based
5 process (GO:0007017) with *P. acuta* and *P. lobata*. *M. capitata* and *M. cavernosa* also shared
6 three GO terms related to developmental processes (positive regulation of neuron projection
7 development (GO:0010976) and retina development in camera-type eye (GO:0060041; Table 3;
8 Supplementary Table 3). Among others, *M. cavernosa* from the FL experiment, shared GO terms
9 relating to ectoderm development (GO:0007398) and regulation of protein serine/threonine
10 kinase activity (GO:0071900) with *P. acuta* and *P. lobata*, respectively. *A. cervicornis* and *P.*
11 *acuta*, both branching corals, shared 11 GO terms, including Mo-molybdopterin cofactor
12 biosynthetic process (GO:0006777) and Wnt receptor signaling pathway (GO:0016055). With
13 the exception of *O. faveolata*, the massive corals, *M. cavernosa* and *P. lobata*, shared GO terms
14 relating to regulation of cell junction assembly (GO:1901888), regulation of protein
15 serine/threonine kinase activity (GO:0071900), and somatic muscle development (GO:0007525).
16 In total, more significantly enriched GO terms were identified at HI and in the branching corals
17 (Figures 4 & 5).

18

19 Orthogroups

20 In total, 21,688 total orthogroups were identified and 9216 were common to all species. Amongst
21 the Florida species, 119, 31, and 2 DEGs from *A. cervicornis*, *M. cavernosa*, and *O. faveolata*,
22 respectively, were found in the common orthogroups (Table 2; Supplementary Table 4).

Sediment stress across location and morphology

1 Amongst the Hawai*i* species, 87, 123, and 66 DEGs from *M. capitata*, *P. acuta*, and *P. lobata*,
2 respectively, were identified in the common orthogroups (Table 2; Supplementary Table 4).

3

4 Similarly to GO terms, there was overlap in orthogroups between species. There was one 4-way
5 interaction, in which *A. cervicornis*, *M. capitata*, *P. acuta*, and *P. lobata* shared two orthogroups
6 (Figure 6; Supp. Table 5). *A. cervicornis* and *P. acuta* shared 3 orthogroups, the most
7 orthogroups shared between a set of species (Figure 6; Supplementary Table 5). Although there
8 may not have been a high number of overlapping orthogroups, the function of those overlapping
9 orthogroups were similar between species. For example, one orthogroup (OG0000487) contained
10 DEGs related to microtubule-based processes and structural constituent of cytoskeleton in *M.*
11 *capitata* and *P. acuta* (Supplementary Table 5).

12

13

14

15

16

17

18

19

20

21

22

23

Sediment stress across location and morphology

1 **Discussion**

2 Sediment type, size, deposition rate, and exposure time varies greatly across coral sediment
3 stress studies, making it difficult to synthesize coral responses to sediment stress. In this study,
4 we conducted two separate experiments to characterize the molecular underpinnings of corals
5 with differing morphological characteristics responding to sediment stressors in two locations.
6 The methodological differences prevent us from making direct statistical comparisons between
7 the experiments. However, it is still relevant to highlight general biological processes and
8 mechanisms related to sediment stress responses across morphology and location.

9

10 *Responses to unsterilized red sediment in Hawai'i*

11 Corals in Hawai'i have become increasingly exposed to sedimentation, agricultural runoff, and
12 pollution that contains organic material (Pastorok and Bilyard 1985; Hunter and Evans 1995;
13 Ogston and Field 2010; Abaya et al. 2018). The increased prevalence of organic matter in
14 sediment and outflows can negatively affect corals (Loiola, Oliveira, and Kikuchi 2013). When
15 exposed to silty nutrient-rich sediments, it was more difficult for *Montipora peltiformis* to
16 remove these particles and photosynthetic yields decreased significantly after 36 hours (Weber et
17 al. 2012). Organic enrichment of sediment also increased the amount of microbial activity in the
18 sediment, causing anoxia and increased disease prevalence in corals (Hodgson 1990; Weber,
19 Lott, and Fabricius 2006; Weber et al. 2012). In this study, *M. capitata*, *P. acuta* and *P. lobata*
20 were exposed to unsterilized red clay sediment for up to 7 days. The sediment was obtained from
21 the highest point on Moku O Lo'e (Coconut Island) and was not sterilized.

22

Sediment stress across location and morphology

1 It is well established that morphology can play a role in modulating a coral's response to
2 sediment stress, but it is unknown if gene expression varies in corals with differing
3 morphological traits. Our study found that, across morphologies in Hawaiian corals,
4 developmental processes were primarily affected by unsterilized sediment. Previous work has
5 demonstrated that sediment deposition, high turbidity, and low light have been shown to
6 adversely affect development and reproductive output across morphologies (Kojis and Quinn
7 1984; Gilmour 1999; Humphrey et al. 2008), though *M. capitata* development and fecundity are
8 not always negatively affected by high sedimentation rates (Padilla-Gamiño et al. 2014). In the
9 present study, the branching coral, *P. acuta*, shared a relatively high number of overlapping
10 developmental processes-related responses with corals with intermediate (*M. capitata*; 20 shared
11 terms) and massive (*P. lobata*; 11 shared terms) morphology. In our study, shared responses
12 corresponded to reproduction and developmental processes like embryonic axis specifications,
13 embryonic organ development, eye development, neural crest cell fate specification, neural plate
14 anterior/posterior regionalization, embryonic hindlimb morphogenesis and others. Unexpectedly,
15 these terms primarily correspond to vertebrate developmental processes, complicating our ability
16 to interpret these results. Given our use of specific annotation databases (i.e., NCBI, SwissProt,
17 InterPro), it is possible that these databases are all dominated by vertebrate-related annotations
18 and thus, invertebrate protein sequences are assigned vertebrate-centric annotations. While it is
19 clear that developmental processes are affected by short term sediment stress across
20 morphologies, it is unclear how vertebrate specific gene functions may translate to
21 developmental processes in basal metazoans. More work is necessary to evaluate equivalency
22 across annotation softwares.

23

Sediment stress across location and morphology

1 During the year, *M. capitata*, *P. acuta*, and *P. lobata* develop their gametes and then release
2 when conditions are optimal, typically June and July in Hawai'i (Stimson 1978; Richmond and
3 Jokiel 1984; Padilla-Gamiño and Gates 2012; Brown et al. 2020). Because the corals in our study
4 were exposed during this time period, it is possible that we captured gene expression signatures
5 unique to corals at or near the peak of their reproductive phenotype. Although the corals were
6 not sampled at different times of year under the same experimental conditions, these results
7 suggest that because the corals were sampled during their reproductive period, signals of
8 developmental processes may be higher than they might have been at other points in the year.
9 Given the high number of shared and unique developmental process GO terms in the Hawai'i
10 corals, however, it is likely that sediment has the potential to have reproductive and
11 developmental consequences, which can have subsequent impacts on population growth and
12 dynamics.

13

14 The only specific shared response across the three Hawaiian species was the downregulation of
15 microtubule-based processes. Microtubules, tubulin polymers that maintain structure and shape
16 to eukaryotic cells, are major components of cilia, which coral utilize for activities such as
17 feeding and clearing sediment (Westbroek, Yanagida, and Isa 1980; Rogers 1990; Erfteimeijer et
18 al. 2012). Downregulation of processes relating to cilia biogenesis/degradation and motility was
19 also observed in the *P. acuta* host and its endosymbionts in response to combined acute heat and
20 sediment stress (Poquita-Du et al. 2019, 2020). The authors of these studies hypothesized that the
21 corals were likely diverting energy resources away from feeding and active sediment clearing in
22 order to prioritize energy for basic homeostasis. While our experiment did not include heat
23 stress, it is probable that downregulation of microtubule or cilia-related processes is a generally

Sediment stress across location and morphology

1 conserved response to sediment stress. Exhaustion of sediment-clearing activity of corals and
2 eventual loss/breakdown of cilia cells can also be a consequence of continued exposure to high
3 levels of sediment stress (Stafford-Smith and Ormond 1992; Stafford-Smith 1993; Erftemeijer et
4 al. 2012). Given the acute nature of the stress in the Hawaiꞏi experiment (i.e., sediment was
5 added at day 0 and day 4), the corals here may have exhausted their ciliary action abilities, thus
6 leading to downregulation of microtubule-based processes.

7 8 Responses to sterilized white sediment in Florida

9 Similarly to Hawaiꞏi, corals in Florida are subject to an increased number of sedimentation
10 events, in particular through dredging and coastal development (Barnes et al. 2015; Miller et al.
11 2016; Cuning et al. 2019). These activities disturb bottom/seafloor sediments, which reduces
12 light availability and can smother corals, leading to a multitude of adverse effects (Erftemeijer et
13 al. 2012; Jones et al. 2016). The Florida experiment was designed to mimic movement of sand
14 caused by dredging over a 18 day period; sand sediment was added periodically to simulate
15 repeated dredgings. In contrast to the Hawaiꞏi sediment, the Florida experimental sediment was
16 sterilized prior to exposures. Despite being sterilized, the sediment still had the capacity to
17 induce adverse effects in corals. Previous studies have found that sterilized sediment can deplete
18 energy reserves, induce immune responses, and decrease photosynthetic capabilities (Browne et
19 al. 2014; Junjie et al. 2014; Sheridan et al. 2014). Thus, it is likely that the sterilized sediment
20 stressor used in the Florida experiment induced an ecologically relevant response.

21
22 No specific responses were shared among *A. cervicornis*, *M. cavernosa*, and *O. faveolata*,
23 despite being exposed to the same sediment for the same amount of time. Given the length of

Sediment stress across location and morphology

1 time of this experiment, it is possible that we only captured the longer-term stress responses of
2 these corals and that their responses may have been more similar at the beginning of the
3 exposures. Morphology also may have contributed to the divergent responses between these
4 species. In previous sediment stress studies, morphology contributed to physiological response
5 variability (Stafford-Smith 1993; Fabricius 2005; Erftemeijer et al. 2012). For example, Rogers
6 (1979) evaluated the effect of shading (as a proxy for turbidity) for five weeks on several coral
7 species from San Cristobal Reef, Puerto Rico. The branching coral, *A. cervicornis*, had entirely
8 bleached, while the massive coral, *M. cavernosa*, was visibly unaffected and appeared to have
9 little response. On the other hand, *M. annularis* had substantial bleaching after 5 weeks, despite
10 being a close relative of *M. cavernosa* (Rogers 1990). Different branching coral species exhibit a
11 wide range of sediment tolerances; after 12 weeks of exposure, the lowest sediment treatments
12 that caused full colony mortality were 30 mg/L⁻¹ for *M. aequituberculata* and 100 mg/L⁻¹ for *A.*
13 *millepora* (Flores et al. 2012). Stafford-Smith (1993) examined sediment rejection efficiency in
14 22 species of Australian corals from a range of morphologies, finding that there was a wide
15 range of active-rejection efficiencies between species. For instance, *Gardineroseris planulata* is
16 a competent rejector of a variety of sediment sizes, but only for a short period of time. *Favia*
17 *stelligera* and *Leptoria phrygia* had moderate clearance rates, but tissue mortality occurred
18 within one to two days. *M. aequituberculata* and *Porites* spp. had low rejection efficiency and
19 had bleached tissues, but no tissue mortality for up to 8 days. Morphology was a driving factor in
20 those results, as branching corals had faster clearing rates than massive corals (Stafford-Smith
21 1993). Thus, it is likely that morphology played a role in driving differences in long-term
22 response in the Floridian corals, highlighting the challenge of predicting responses across
23 species.

Sediment stress across location and morphology

1
2 Despite differing morphologies, the branching coral, *A. cervicornis*, shared responses related to
3 developmental processes and signal transduction with both massive corals, *M. cavernosa* and *O.*
4 *faveolata*, independently. Interestingly, no specific responses were shared between *M. cavernosa*
5 and *O. faveolata*, despite similar morphologies. Downregulation of Rho protein signal
6 transduction was observed in *A. cervicornis* and *M. cavernosa* (with the exception of
7 upregulation in the *M. cavernosa* T2vT3 treatment comparison). Rho proteins are part of a
8 superfamily of signaling GTPase proteins, which typically control the assembly and organization
9 of the cytoskeleton, as well as participate in functions such as cell adhesion, contraction,
10 migration, morphogenesis, and phagocytosis (Mackay and Hall 1998; Moon and Zheng 2003;
11 Phuyal and Farhan 2019). In corals, Rho GTPases participate in cytoskeleton remodeling during
12 phagocytosis, as well as cell division of endosymbionts within symbiotic gastrodermal cells (Li
13 et al. 2014). Rho GTPase pathways have been identified in coral polyp bailout responses to heat
14 stress and hyperosmosis, as well as in bleached corals in response to low flow environments
15 (Chuang and Mitarai 2020; Fifer et al. 2021; Gösler et al. 2021). In this study, downregulation of
16 Rho protein pathways suggests that minimal cytoskeleton maintenance, assembly and
17 organization is occurring and that the corals may not be able to properly maintain their cellular
18 structures under sedimentation.
19
20 *A. cervicornis* and *O. faveolata* both downregulated chondrocyte development, a developmental
21 response. Chondrocytes are cells in cartilage that make up the cytoskeletal matrix in humans and
22 other animals, including some marine invertebrates (Philpott and Person 1970; Cowden and
23 Fitzharris 1975; Libbin et al. 1976; Archer and Francis-West 2003; Kamisan et al. 2013). In

Sediment stress across location and morphology

1 corals, the cytoskeleton matrix is made up of calcium carbonate, as opposed to cartilage, which
2 forms through rapid accretion of protein rich skeletal organic matrix and extracellular calcium
3 carbonate crystals to form a stony skeleton (Vandermeulen and Watabe 1973; Akiva et al. 2018).
4 Skeletal matrix formation begins when planktonic coral larvae settle and begin to secrete calcium
5 carbonate, which helps to anchor the coral to the substrate (Akiva et al. 2018). The skeleton
6 grows as the coral animal continues to secrete calcium carbonate, building up a large and
7 intricate 3D skeletal structure (Tambutté et al. 2011). Sedimentation can hinder coral skeletal
8 growth by depositing sediment on the tissue and diverting energy away from growth, as well as
9 decreasing the amount of light that reaches the coral, thus affecting the possibility for light-
10 enhanced calcification, which is responsible for most of the skeletal growth in corals (Dodge,
11 Aller, and Thomson 1974; Erfteimeijer et al. 2012). Downregulation of chondrocyte development
12 may be related to the disruption of skeletal matrix formation, which may have an effect on
13 skeletal density and growth. Decreased skeletal density was observed in corals from inshore reefs
14 which experience higher levels of sedimentation as compared to corals from offshore reefs
15 (Lough and Barnes 1992). Therefore, the downregulation of genes relating to chondrocyte
16 development across morphologies, suggests that corals with a range of morphological
17 characteristics may have decreased skeletal growth in response to sediment stress.

18

Differences in response based on morphological characteristics

20 This study combined data from two independent experiments in order to characterize the
21 molecular mechanisms that corals use to respond to different sedimentation stressors. The
22 experiments used different sediment types (Hawaii: unsterilized red clay sediment, Florida:
23 sterilized carbonate sand sediment) and lengths of exposure (Hawaii: up to 7 days, Florida: 18

Sediment stress across location and morphology

1 days). In both experiments, branching, massive and intermediate coral morphologies were
2 exposed. Differences in experimental methodology prevent us from making direct statistical
3 comparisons between the two experiments; however, the shared morphologies between
4 experiments enables us to identify broad generalizations about conserved gene regulation in
5 response to sediment stress. These comparisons are relevant, as knowledge of what genes and
6 biological processes are broadly affected by sediment stress can help coral reef management.
7
8 We did not identify a generalized response across morphology nor gene expression patterns
9 across taxa. There were two groups of 3 species (*M. capitata*, *M. cavernosa*, *P. acuta* and *A.*
10 *cervicornis*, *M. capitata*, *M. cavernosa*) that shared specific responses. However, no group
11 contained a single morphology and gene expression patterns varied among species. For example,
12 *M. capitata* (intermediate), *M. cavernosa* (massive), and *P. acuta* (branching) all expressed
13 genes relating to riboflavin transport, the transport of certain vitamins in cells, though they had
14 very different expression patterns. Riboflavin transport genes were upregulated in *P. acuta*, but
15 downregulated in *M. cavernosa*; *M. capitata*, on the other hand, differentially expressed two
16 genes relating to riboflavin transport, one of which was upregulated and the other
17 downregulated. Thus, even though a shared response was identified, the direction of the response
18 varies greatly. This result demonstrates that responses may be shared across morphologies and
19 sediment types/locations, but it may be difficult to predict the directionality of the response. The
20 other group, which contained *A. cervicornis* (branching), *M. capitata* (intermediate), and *M.*
21 *cavernosa* (massive), all downregulated genes relating to positive regulation of skeletal muscle
22 tissue development. This term refers to the activation, maintenance, or increase of the rate of
23 skeletal muscle tissue development (Buckingham et al. 2003; Grefte et al. 2007). Given that this

Sediment stress across location and morphology

1 term is downregulated, it means that there is little to no activation, maintenance, or increase of
2 the rate of skeletal muscle tissue development in these corals. These gene expression patterns
3 highlight the complexity of characterizing responses to different kinds of sedimentation stress in
4 species with different morphotypes.

5
6 Some terms were shared across both morphology and location, indicating a generalized sediment
7 stress response. For example, DNA methylation-dependent heterochromatin assembly was
8 shared between *A. cervicornis* (branching) and *P. lobata* (massive). Opposite expression patterns
9 were again observed, in which upregulation occurred in *A. cervicornis* and downregulation in *P.*
10 *lobata*. DNA methylation-dependent heterochromatin assembly refers to repression of
11 transcription by DNA methylation leading to the formation of heterochromatin (Jones and
12 Wolffe 1999; Grewal and Moazed 2003). In the case of *A. cervicornis*, upregulation suggests that
13 repression of transcription by DNA methylation and subsequent heterochromatin formation is
14 occurring. Therefore, certain portions of DNA cannot be accessed, giving *A. cervicornis* more
15 stringent control on gene expression. On the other hand, the downregulation of these genes in *P.*
16 *lobata* means less repression of transcription by DNA methylation occurring and
17 heterochromatin is not being fully formed, leaving much of the DNA accessible to transcription
18 machinery and ultimately, reducing the amount of control that *P. lobata* has on gene expression.
19 To date, no work has examined how sediment stress affects epigenetic mechanisms, such as
20 DNA methylation, in coral. However, previous studies have found changes to DNA methylation
21 in response to stress and environmental change (Putnam, Davidson, and Gates 2016; Liew et al.
22 2018; Dimond and Roberts 2020; Rodríguez-Casariago, Mercado-Molina, and Garcia-Souto
23 2020). Additionally, it has been observed in corals that genes with weak methylation signals are

Sediment stress across location and morphology

1 more likely to demonstrate differential gene expression (Dixon, Bay, and Matz 2014;
2 Entrambasaguas et al. 2021). Epigenetic modifications and their regulation of gene transcription
3 are highly species and context dependent. Furthermore, the directionality of epigenetic regulation
4 on gene expression or repression can vary depending on the underlying genetic machinery and
5 the environment. This is exemplified in the two species with overlapping response terms.
6 Namely, *A. cervicornis* exhibits a more regulated control on gene expression in contrast to *P.*
7 *lobata*, which exhibits a less regulated profile of gene regulation. Thus, it is likely that
8 sedimentation stress from each location impacted DNA methylation and heterochromatin in
9 different ways, causing opposing expression patterns.

10

11 More general responses were shared over morphology and location, as identified by the
12 orthogroup analysis. The orthogroup analysis grouped homologous gene sequences in different
13 species related to one another by linear descent. The resulting ‘orthogroup’ represents a group of
14 similar gene sequences across multiple species (Emms and Kelly 2015, 2019). Orthogroups were
15 shared across morphology and location, though in relatively low numbers (sharing between 1 and
16 3 orthogroups). This sharing may represent a group of orthogroups that form a core group of
17 genes in response to sediment stress. Although we cannot directly compare the genes or
18 orthogroups between experiments, the shared orthogroups represent potential overlap in
19 sedimentation response over location and morphology.

20

21 The branching corals, *A. cervicornis* from the Florida experiment and *P. acuta* from the Hawaiꞏi
22 experiment, shared a metabolic response, ‘Mo-molybdopterin cofactor biosynthetic process’,
23 which describes the creation of the Mo-molybdopterin cofactor, an essential component for

Sediment stress across location and morphology

1 catalytic activity of certain enzymes (Kisker, Schindelin, and Rees 1997; Mendel 2013).
2 Molybdenum (Mo) is a trace metal synthesized *de novo* through GTP-based processes; cyclic
3 pyranopterun monophosphate (cPMP) is initially formed, which is then converted to the
4 molybdopterin cofactor (Mendel 2013). This essential cofactor catalyzes the oxidation and
5 reduction of molecules in enzymatic processes regulating nitrogen, sulfur, and carbon (Daniels et
6 al. 2008; Iobbi-Nivol and Leimkühler 2013). Similarly to other results in the present study, both
7 species demonstrated opposing differential gene expression for this term. Mo-molybdopterin
8 cofactor biosynthetic process was downregulated in *A. cervicornis*, but upregulated in *P. acuta*,
9 suggesting that while sediment type can induce similar differentially expressed genes, it can
10 produce different expression patterns for those genes. Molybdopterin are co-factors for
11 oxidoreductases, a family of enzymes that catalyze the transfer of electrons between molecules
12 (Kisker, Schindelin, and Rees 1997). Upregulation of molybdopterin synthesis may suggest that
13 these types of enzymes are more metabolically active. Stressful conditions have made these
14 kinds of enzymes more active in plants and corals (Bouchard and Yamasaki 2008; Zdunek-
15 Zastocka and Sobczak 2013). Upregulation of metabolism related genes was also observed in a
16 study that examined transcriptomic responses of corals in response to two different sediment
17 experiments (Bollati et al. 2021). Downregulation of Mo-molybdopterin cofactor synthesis may
18 indicate that the coral does not have enough energy to synthesize molybdopterin which in turn
19 makes it so the activity of these specific enzymes is decreased or stopped altogether, suggesting a
20 decrease in metabolism for *A. cervicornis*. It is also possible that the unsterilized sediment was
21 providing molybdopterin to *P. acuta*, making it necessary for *P. acuta* to upregulate genes
22 relating to molybdopterin processing to manage the influx (Fujimoto and Sherman 1951; Siebert
23 et al. 2015). Because the sediment was sterilized in the Florida experiment, no molybdenum

Sediment stress across location and morphology

1 would be present in the sediment, indicating that molybdenum-related enzymes may not have
2 been functioning at a high level, leading to downregulation.

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

Sediment stress across location and morphology

1 **Conclusion**

2 This study incorporated data from two separate experiments to more fully characterize the
3 molecular mechanisms induced by sedimentation in corals. We found that developmental
4 processes are primarily impacted in branching corals across location, highlighting potential
5 future research avenues with regards to sediment stress and reproductive potential and output.
6 While few specific genes were shared across morphology and location, orthogroup analysis
7 uncovered potential overlap in generalized sediment stress responses. Direct comparisons across
8 species are necessary to further elucidate the genetic basis of coral susceptibility to sediment
9 stress.

10

11

12

13

14

15

16

17

18

19

20

21

22

23

Sediment stress across location and morphology

1 **Acknowledgments**

2 This study was supported by National Fish and Wildlife Foundation, NSF HDR Awards
3 #1939795 and #1939263, and an NSF GRFP. We thank Erin Chille for assistance with
4 bioinformatic analyses. We thank the Hawai'i Institute of Marine Biology (HIMB) in
5 Kāneohe Bay, Oahu, Hawai'i and the Smithsonian Marine Station (SMS) in Fort Pierce,
6 Florida for facilities support.

7

8 **Conflicts of interest**

9 On behalf of all authors, the corresponding author states there is no conflict of interest.

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

Sediment stress across location and morphology

1 References

- 2 Abaya, L. M., T. N. Wiegner, J. P. Beets, S. L. Colbert, K. M. Carlson, and K. L. Kramer. 2018.
3 “Spatial Distribution of Sewage Pollution on a Hawaiian Coral Reef.” *Marine Pollution*
4 *Bulletin* 130 (May): 335–47. <https://doi.org/10.1016/j.marpolbul.2018.03.028>.
- 5 Akiva, A., M. Neder, K. Kahil, R. Gavriel, I. Pinkas, G. Goobes, and T. Mass. 2018. “Minerals
6 in the Pre-Settled Coral *Stylophora pistillata* Crystallize via Protein and Ion Changes.”
7 *Nature Communications* 9 (1): 1880. <https://doi.org/10.1038/s41467-018-04285-7>.
- 8 Altschul, S. F., W. Gish, W. Miller, E. W. Myers, and D. J. Lipman. 1990. “Basic Local
9 Alignment Search Tool.” *Journal of Molecular Biology* 215 (3): 403–10.
10 [https://doi.org/10.1016/S0022-2836\(05\)80360-2](https://doi.org/10.1016/S0022-2836(05)80360-2).
- 11 Anders, S., and W. Huber. 2010. “Differential Expression Analysis for Sequence Count Data.”
12 *Genome Biology* 11 (10): R106. <https://doi.org/10.1186/gb-2010-11-10-r106>.
- 13 Andrews, S. 2010. “A Quality Control Tool for High Throughput Sequence Data” [Online].
14 Available online at: <http://www.bioinformatics.babraham.ac.uk/projects/fastqc>.
- 15 Anthony, K. R. N., and K. E. Fabricius. 2000. “Shifting Roles of Heterotrophy and Autotrophy
16 in Coral Energetics under Varying Turbidity.” *Journal of Experimental Marine Biology and*
17 *Ecology* 252 (2): 221–53. [https://doi.org/10.1016/S0022-0981\(00\)00237-9](https://doi.org/10.1016/S0022-0981(00)00237-9).
- 18 Archer, C. W., and P. Francis-West. 2003. “The Chondrocyte.” *The International Journal of*
19 *Biochemistry & Cell Biology* 35 (4): 401–4. [https://doi.org/10.1016/s1357-2725\(02\)00301-](https://doi.org/10.1016/s1357-2725(02)00301-1)
20 1.
- 21 Bahr, K.D., K.S. Rodgers, P.L. Jokiel, N.G. Prouty, and C.D. Storlazzi. 2020. “Pulse sediment
22 event does not impact the metabolism of a mixed coral reef community”. *Ocean and*
23 *Coastal Management* 184 (105007).
24 <https://doi.org/10.1016/j.ocecoaman.2019.105007>.
- 25 Bairoch, A., and R. Apweiler. 1997. “The SWISS-PROT Protein Sequence Data Bank and Its
26 Supplement TrEMBL.” *Nucleic Acids Research* 25 (1): 31–36.
27 <https://doi.org/10.1093/nar/25.1.31>.
- 28 Bak, R. P. M. 1978. “Lethal and Sublethal Effects of Dredging on Reef Corals.” *Marine*
29 *Pollution Bulletin* 9 (1): 14–16. [https://doi.org/10.1016/0025-326X\(78\)90275-8](https://doi.org/10.1016/0025-326X(78)90275-8).
- 30 Barnes, B. B., C. Hu, C. Kovach, and R. N. Silverstein. 2015. “Sediment Plumes Induced by the
31 Port of Miami Dredging: Analysis and Interpretation Using Landsat and MODIS Data.”
32 *Remote Sensing of Environment* 170: 328. <https://doi.org/10.1016/j.rse.2015.09.023>.
- 33 Bessell-Browne, P., A. P. Negri, R. Fisher, P. L. Clode, A. Duckworth, and R. Jones. 2017.
34 “Impacts of Turbidity on Corals: The Relative Importance of Light Limitation and
35 Suspended Sediments.” *Marine Pollution Bulletin* 117 (1): 161–70.
36 <https://doi.org/10.1016/j.marpolbul.2017.01.050>.
- 37 Bolger, A. M., M. Lohse, and B. Usadel. 2014. “Trimmomatic: A Flexible Trimmer for Illumina
38 Sequence Data.” *Bioinformatics* 30 (15): 2114–20.
39 <https://doi.org/10.1093/bioinformatics/btu170>.
- 40 Bollati, E., Y. Rosenberg, N. Simon-Blecher, R. Tamir, O. Levy, and D. Huang. 2021.
41 “Untangling the Molecular Basis of Coral Response to Sedimentation.” *Molecular Ecology*,
42 November. <https://doi.org/10.1111/mec.16263>.
- 43 Bouchard, J. N., and H. Yamasaki. 2008. “Heat Stress Stimulates Nitric Oxide Production in
44 Symbiodinium *Microadriaticum*: A Possible Linkage between Nitric Oxide and the Coral
45 Bleaching Phenomenon.” *Plant & Cell Physiology* 49 (4): 641–52.
46 <https://doi.org/10.1093/pcp/pcn037>.

Sediment stress across location and morphology

- 1 Brodie, J., J. Waterhouse, B. Schaffelke, M. Furnas, J. Maynard, C. Collier, S. Lewis, M.
2 Warne, K. Fabricius, M. Devlin, L. McKenzie, H. Yorkston, L. Randall, J. Bennett, and V.
3 Brando. 2013. “Relative Risks to the Great Barrier Reef from Degraded Water Quality.” In
4 *2013 Scientific Consensus Statement*. Reef Water Quality Protection Plan Secretariat, The
5 State of Queensland.
- 6 Browne, N. K., E. Precht, K. S. Last, and P. A. Todd. 2014. “Photo-Physiological Costs
7 Associated with Acute Sediment Stress Events in Three near-Shore Turbid Water Corals.”
8 *Marine Ecology Progress Series* 502: 129–43. <https://doi.org/10.3354/meps10714>.
- 9 Brown, N. P., Z. H. Forsman, K. T. Tisthammer, and R. H. Richmond. 2020. “A Resilient
10 Brooding Coral in the Broadcast Spawning *Porites lobata* Species Complex: A New
11 Endemic, Introduced Species, Mutant, or New Adaptive Potential?” *Coral Reefs* 39 (3):
12 809–18. <https://doi.org/10.1007/s00338-020-01922-w>.
- 13 Buckingham, M., L. Bajard, T. Chang, P. Daubas, J. Hadchouel, S. Meilhac, D. Montarras, D.
14 Rocancourt, and F. Relaix. 2003. “The Formation of Skeletal Muscle: From Somite to
15 Limb.” *Journal of Anatomy* 202 (1): 59–68. [https://doi.org/10.1046/j.1469-
16 7580.2003.00139.x](https://doi.org/10.1046/j.1469-7580.2003.00139.x).
- 17 Chuang, P., and S. Mitarai. 2020. “Signaling Pathways in the Coral Polyp Bail-out Response.”
18 *Coral Reefs* 39 (6): 1535–48. <https://doi.org/10.1007/s00338-020-01983-x>.
- 19 Costanza, R., R. Groot, P. Sutton, S. Ploeg, S. J. Anderson, I. Kubiszewski, S. Farber, and R. K.
20 Turner. 2014. “Changes in the Global Value of Ecosystem Services.” *Global Environmental*
21 *Change: Human and Policy Dimensions* 26 (May): 152–58.
22 <https://doi.org/10.1016/j.gloenvcha.2014.04.002>.
- 23 Cowden, R. R., and T. Fitzharris. 1975. “The Histochemistry and Structure of Tentacle Cartilage
24 Tissues in the Marine Polychaete, *Sabella melanostigma*.” *Histochemistry* 43 (1): 1–10.
25 <https://doi.org/10.1007/BF00490150>.
- 26 Cunning, R., R. N. Silverstein, B. B. Barnes, and A. C. Baker. 2019. “Extensive Coral Mortality
27 and Critical Habitat Loss Following Dredging and Their Association with Remotely-Sensed
28 Sediment Plumes.” *Marine Pollution Bulletin* 145 (August): 185–99.
29 <https://doi.org/10.1016/j.marpolbul.2019.05.027>.
- 30 Dallmeyer, D. G., J. W. Porter, and G. J. Smith. 1982. “Effects of Particulate Peat on the
31 Behavior and Physiology of the Jamaican Reef-Building Coral *Montastrea annularis*.”
32 *Marine Biology* 68 (3): 229–33. <https://doi.org/10.1007/BF00409589>.
- 33 Daniels, J. N., M. M. Wuebbens, K. V. Rajagopalan, and H. Schindelin. 2008. “Crystal Structure
34 of a Molybdopterin Synthase–Precursor Z Complex: Insight into Its Sulfur Transfer
35 Mechanism and Its Role in Molybdenum Cofactor Deficiency,.” *Biochemistry* 47 (2): 615–
36 26. <https://doi.org/10.1021/bi701734g>.
- 37 Davy, S. K., D. Allemand, and V. M. Weis. 2012. “Cell Biology of Cnidarian-Dinoflagellate
38 Symbiosis.” *Microbiology and Molecular Biology Reviews: MMBR* 76 (2): 229–61.
39 <https://doi.org/10.1128/MMBR.05014-11>.
- 40 DeSalvo, M. K., S. Sunagawa, C. R. Voolstra, and M. Medina. 2010. “Transcriptomic Responses
41 to Heat Stress and Bleaching in the Elkhorn Coral *Acropora palmata*.” *Marine Ecology*
42 *Progress Series* 402: 97–113. <http://www.jstor.org/stable/24873809>.
- 43 DeSalvo, M. K., C. R. Voolstra, S. Sunagawa, J. A. Schwarz, J. H. Stillman, M. A. Coffroth, A.
44 M. Szmant, and M. Medina. 2008. “Differential Gene Expression during Thermal Stress
45 and Bleaching in the Caribbean Coral *Montastraea faveolata*.” *Molecular Ecology* 17 (17):
46 3952–71. <https://doi.org/10.1111/j.1365-294X.2008.03879.x>.

Sediment stress across location and morphology

- 1 Dimond, J. L., and S. B. Roberts. 2020. “Convergence of DNA Methylation Profiles of the Reef
2 Coral *Porites astreoides* in a Novel Environment.” *Frontiers in Marine Science* 6 (January):
3 1–12. <https://doi.org/10.3389/fmars.2019.00792>.
- 4 Dixon, G. B., L. K. Bay, and M. V. Matz. 2014. “Bimodal Signatures of Germline Methylation
5 Are Linked with Gene Expression Plasticity in the Coral *Acropora millepora*.” *BMC*
6 *Genomics* 15 (1): 1–11. <https://doi.org/10.1186/1471-2164-15-1109>.
- 7 Dobin, A., C. A. Davis, F. Schlesinger, J. Drenkow, C. Zaleski, S. Jha, P. Batut, M. Chaisson,
8 and T. R. Gingeras. 2013. “STAR: Ultrafast Universal RNA-Seq Aligner.” *Bioinformatics*
9 29 (1): 15–21. <https://doi.org/10.1093/bioinformatics/bts635>.
- 10 Dodge, R. E., R. C. Aller, and J. Thomson. 1974. “Coral Growth Related to Resuspension of
11 Bottom Sediments” 247 (5442): 574–77. <https://doi.org/10.1038/247574a0>.
- 12 Emms, D. M., and S. Kelly. 2015. “OrthoFinder: Solving Fundamental Biases in Whole Genome
13 Comparisons Dramatically Improves Orthogroup Inference Accuracy.” *Genome Biology* 16
14 (August): 157. <https://doi.org/10.1186/s13059-015-0721-2>.
- 15 ———. 2019. “OrthoFinder: Phylogenetic Orthology Inference for Comparative Genomics.”
16 *Genome Biology* 20 (1): 238. <https://doi.org/10.1186/s13059-019-1832-y>.
- 17 Entrambasaguas, L., M. Ruocco, K. J. F. Verhoeven, G. Procaccini, and L. Marín-Guirao. 2021.
18 “Gene Body DNA Methylation in Seagrasses: Inter- and Intraspecific Differences and
19 Interaction with Transcriptome Plasticity under Heat Stress.” *Scientific Reports* 11 (1):
20 14343. <https://doi.org/10.1038/s41598-021-93606-w>.
- 21 Erfteemeijer, P. L. A., A. Riegl, B. W. Hoeksema, and P. A. Todd. 2012. “Environmental Impacts
22 of Dredging and Other Sediment Disturbances on Corals: A Review.” *Marine Pollution*
23 *Bulletin* 64 (9): 1737–65. <https://doi.org/10.1016/j.marpolbul.2012.05.008>.
- 24 Ewels, P., M. Magnusson, S. Lundin, and M. Källér. 2016. “MultiQC: Summarize Analysis
25 Results for Multiple Tools and Samples in a Single Report.” *Bioinformatics* 32 (19): 3047–
26 48. <https://doi.org/10.1093/bioinformatics/btw354>.
- 27 Fabricius, K. E. 2005. “Effects of Terrestrial Runoff on the Ecology of Corals and Coral Reefs:
28 Review and Synthesis.” *Marine Pollution Bulletin* 50 (2): 125–46.
29 <https://doi.org/10.1016/j.marpolbul.2004.11.028>.
- 30 Fabricius, K. E., G. De’ath, C. Humphrey, I. Zagorskis, and B. Schaffelke. 2013. “Intra-Annual
31 Variation in Turbidity in Response to Terrestrial Runoff on near-Shore Coral Reefs of the
32 Great Barrier Reef.” *Estuarine, Coastal and Shelf Science* 116 (January): 57–65.
33 <https://doi.org/10.1016/j.ecss.2012.03.010>.
- 34 Fifer, J., B. Bentlage, S. Lemer, A. G. Fujimura, M. Sweet, and L. J. Raymundo. 2021. “Going
35 with the Flow: How Corals in High-Flow Environments Can Beat the Heat.” *Molecular*
36 *Ecology* 30 (9): 2009–24. <https://doi.org/10.1111/mec.15869>.
- 37 Flores, F., M. O. Hoogenboom, L. D. Smith, T. F. Cooper, D. Abrego, and A. P. Negri. 2012.
38 “Chronic Exposure of Corals to Fine Sediments: Lethal and Sub-Lethal Impacts.” *PloS One*
39 7 (5): 1–12. <https://doi.org/10.1371/journal.pone.0037795>.
- 40 Fujimoto, G., and G. D. Sherman. 1951. “Molybdenum Content of Typical Soils and Plants of
41 the Hawaiian Islands.” *Agronomy Journal* 43 (9): 424–29.
42 <https://doi.org/10.2134/agronj1951.00021962004300090003x>.
- 43 Gentleman, R., V. J. Carey, W. Huber, and F. Hahne. 2021. *Genefilter: Genefilter: Methods for*
44 *Filtering Genes from High-Throughput Experiments* (version 1.70.0). R package.
- 45 Gilmour, J. 1999. “Experimental Investigation into the Effects of Suspended Sediment on
46 Fertilisation, Larval Survival and Settlement in a Scleractinian Coral.” *Marine Biology* 135:

Sediment stress across location and morphology

- 1 451–62.
- 2 Gösser, F., A. Raulf, A. Mosig, R. Tollrian, and M. Schweinsberg. 2021. “Signaling Pathways of
- 3 Heat- and Hypersalinity-Induced Polyp Bailout in *Pocillopora acuta*.” *Coral Reefs* 40 (6):
- 4 1713–28. <https://doi.org/10.1007/s00338-021-02191-x>.
- 5 Götz, S., J. M. García-Gómez, J. Terol, T. D. Williams, S. H. Nagaraj, M. J. Nueda, M. Robles,
- 6 M. Talón, J. Dopazo, and A. Conesa. 2008. “High-Throughput Functional Annotation and
- 7 Data Mining with the Blast2GO Suite.” *Nucleic Acids Research* 36 (10): 3420–35.
- 8 <https://doi.org/10.1093/nar/gkn176>.
- 9 Grefte, S., A. M. Kuijpers-Jagtman, R. Torensma, and J. W. Von den Hoff. 2007. “Skeletal
- 10 Muscle Development and Regeneration.” *Stem Cells and Development* 16 (5): 857–68.
- 11 <https://doi.org/10.1089/scd.2007.0058>.
- 12 Grewal, S. I. S., and D. Moazed. 2003. “Heterochromatin and Epigenetic Control of Gene
- 13 Expression.” *Science* 301 (5634): 798–802. <https://doi.org/10.1126/science.1086887>.
- 14 Hashimoto, K., T. Shibuno, E. Murayama-Kayano, H. Tanaka, and T. Kayano. 2004. “Isolation
- 15 and Characterization of Stress-Responsive Genes from the Scleractinian Coral *Pocillopora*
- 16 *damicornis*.” *Coral Reefs* 23 (4): 485–91. <https://doi.org/10.1007/s00338-004-0410-1>.
- 17 Hodgson, G. 1990. “Tetracycline Reduces Sedimentation Damage to Corals.” *Marine Biology*
- 18 104 (3): 493–96. <https://doi.org/10.1007/BF01314355>.
- 19 Hoegh-Guldberg, O., P. J. Mumby, A. J. Hooten, R. S. Steneck, P. Greenfield, E. Gomez, C. D.
- 20 Harvell, et al. 2007. “Coral Reefs under Rapid Climate Change and Ocean Acidification.”
- 21 *Science* 318 (5857): 1737–42. <https://doi.org/10.1126/science.1152509>.
- 22 Houlbrèque, F., and C. Ferrier-Pagès. 2009. “Heterotrophy in Tropical Scleractinian Corals.”
- 23 *Biological Reviews of the Cambridge Philosophical Society* 84 (1): 1–17.
- 24 <https://doi.org/10.1111/j.1469-185X.2008.00058.x>.
- 25 Humphrey, C., M. Weber, C. Lott, T. Cooper, and K. Fabricius. 2008. “Effects of Suspended
- 26 Sediments, Dissolved Inorganic Nutrients and Salinity on Fertilisation and Embryo
- 27 Development in the Coral *Acropora millepora* (Ehrenberg, 1834).” *Coral Reefs* 27 (4):
- 28 837–50. <https://doi.org/10.1007/s00338-008-0408-1>.
- 29 Hunter, C. L., and C. W. Evans. 1995. “Coral Reefs in Kāneohe Bay, Hawai'i: Two Centuries
- 30 of Western Influence and Two Decades of Data.” *Bulletin of Marine Science* 57 (2): 501–
- 31 15.
- 32 Iobbi-Nivol, C., and S. Leimkühler. 2013. “Molybdenum Enzymes, Their Maturation and
- 33 Molybdenum Cofactor Biosynthesis in *Escherichia coli*.” *Biochimica et Biophysica Acta*
- 34 1827 (8-9): 1086–1101. <https://doi.org/10.1016/j.bbabi.2012.11.007>.
- 35 Jones, P., D. Binns, H. Chang, M. Fraser, W. Li, C. McAnulla, H. McWilliam, et al. 2014.
- 36 “InterProScan 5: Genome-Scale Protein Function Classification.” *Bioinformatics* 30 (9):
- 37 1236–40. <https://doi.org/10.1093/bioinformatics/btu031>.
- 38 Jones, P. L., and A. P. Wolffe. 1999. “Relationships between Chromatin Organization and DNA
- 39 Methylation in Determining Gene Expression.” *Seminars in Cancer Biology* 9 (5): 339–47.
- 40 <https://doi.org/10.1006/scbi.1999.0134>.
- 41 Jones, R., P. Bessell-Browne, R. Fisher, W. Klonowski, and M. Slivkoff. 2016. “Assessing the
- 42 Impacts of Sediments from Dredging on Corals.” *Marine Pollution Bulletin* 102 (1): 9–29.
- 43 <https://doi.org/10.1016/j.marpolbul.2015.10.049>.
- 44 Jones, R., R. Fisher, and P. Bessell-Browne. 2019. “Sediment Deposition and Coral
- 45 Smothering.” *PloS One* 14 (6): e0216248. <https://doi.org/10.1371/journal.pone.0216248>.
- 46 Jones, R., N. Giofre, H. M. Luter, T. L. Neoh, R. Fisher, and A. Duckworth. 2020. “Responses of

Sediment stress across location and morphology

- 1 Corals to Chronic Turbidity.” *Scientific Reports* 10 (1): 1–13.
2 <https://doi.org/10.1038/s41598-020-61712-w>.
- 3 Junjie, R. K., N. K. Browne, P. L. A. Erfteimeijer, and P. A. Todd. 2014. “Impacts of Sediments
4 on Coral Energetics: Partitioning the Effects of Turbidity and Settling Particles.” *PloS One*
5 9 (9): e107195. <https://doi.org/10.1371/journal.pone.0107195>.
- 6 Kamisan, N., S. V. Naveen, R. E. Ahmad, and T. Kamarul. 2013. “Chondrocyte Density,
7 Proteoglycan Content and Gene Expressions from Native Cartilage Are Species Specific
8 and Not Dependent on Cartilage Thickness: A Comparative Analysis between Rat, Rabbit
9 and Goat.” *BMC Veterinary Research* 9 (1): 62. <https://doi.org/10.1186/1746-6148-9-62>.
- 10 Kaniewska, P., P. R. Campbell, D. I. Kline, M. Rodriguez-Lanetty, D. J. Miller, S. Dove, and O.
11 Hoegh-Guldberg. 2012. “Major Cellular and Physiological Impacts of Ocean Acidification
12 on a Reef Building Coral.” *PloS One* 7 (4). <https://doi.org/10.1371/journal.pone.0034659>.
- 13 Kisker, C., H. Schindelin, and D. C. Rees. 1997. “MOLYBDENUM-COFACTOR–
14 CONTAINING ENZYMES: Structure and Mechanism.” *Annual Review of Biochemistry* 66
15 (1): 233–67. <https://doi.org/10.1146/annurev.biochem.66.1.233>.
- 16 Kojis, B. L., and N. J. Quinn. 1984. “Seasonal and Depth Variation in Fecundity of *Acropora*
17 *palifera* at Two Reefs in Papua New Guinea.” *Coral Reefs* 3 (3): 165–72.
18 <https://doi.org/10.1007/BF00301961>.
- 19 LaJeunesse, T. C., J. E. Parkinson, P. W. Gabrielson, H. J. Jeong, J. D. Reimer, C. R. Voolstra,
20 and S. R. Santos. 2018. “Systematic Revision of Symbiodiniaceae Highlights the Antiquity
21 and Diversity of Coral Endosymbionts.” *Current Biology: CB* 28 (16): 2570–80.e6.
22 <https://doi.org/10.1016/j.cub.2018.07.008>.
- 23 Lasker, H. R. 1980. “Sediment Rejection by Reef Corals: The Roles of Behavior and
24 Morphology in *Montastrea cavernosa* (Linnaeus).” *Journal of Experimental Marine*
25 *Biology and Ecology* 47 (1): 77–87. [https://doi.org/10.1016/0022-0981\(80\)90139-2](https://doi.org/10.1016/0022-0981(80)90139-2).
- 26 Libbin, R. M., R. Ozer, P. Person, and A. Hirschman. 1976. “In Vitro Accumulation of Mineral
27 Components by Invertebrate Cartilage.” *Calcified Tissue Research* 22 (1): 67–75.
28 <https://doi.org/10.1007/BF02010347>.
- 29 Liew, Y. J., D. Zoccola, Y. Li, E. Tambutté, A. A. Venn, C. T. Michell, G. Cui, et al. 2018.
30 “Epigenome-Associated Phenotypic Acclimatization to Ocean Acidification in a Reef-
31 Building Coral.” *Science Advances* 4 (6): eaar8028. <https://doi.org/10.1126/sciadv.aar8028>.
- 32 Li, H., Z. Huang, S. Ye, C. Lu, P. Cheng, S. Chen, and C. Chen. 2014. “Membrane Labeling of
33 Coral Gastrodermal Cells by Biotinylation: The Proteomic Identification of Surface Proteins
34 Involving Cnidaria-Dinoflagellate Endosymbiosis.” *PloS One* 9 (1): e85119.
35 <https://doi.org/10.1371/journal.pone.0085119>.
- 36 Loiola, M., M. D. M. Oliveira, and R. K. P. Kikuchi. 2013. “Tolerance of Brazilian Brain Coral
37 *Mussismilia braziliensis* to Sediment and Organic Matter Inputs.” *Marine Pollution Bulletin*
38 77 (1-2): 55–62. <https://doi.org/10.1016/j.marpolbul.2013.10.033>.
- 39 Lough, J. M., and D. J. Barnes. 1992. “Comparisons of Skeletal Density Variations in *Porites*
40 from the Central Great Barrier Reef.” *Journal of Experimental Marine Biology and Ecology*
41 155 (1): 1–25. [https://doi.org/10.1016/0022-0981\(92\)90024-5](https://doi.org/10.1016/0022-0981(92)90024-5).
- 42 Love, M. I., W. Huber, and S. Anders. 2014. “Moderated Estimation of Fold Change and
43 Dispersion for RNA-Seq Data with DESeq2.” *Genome Biology* 15 (12): 1–21.
44 <https://doi.org/10.1186/s13059-014-0550-8>.
- 45 Mackay, D. J., and A. Hall. 1998. “Rho GTPases.” *The Journal of Biological Chemistry* 273
46 (33): 20685–88. <https://doi.org/10.1074/jbc.273.33.20685>.

Sediment stress across location and morphology

- 1 Mendel, R. R. 2013. “The Molybdenum Cofactor.” *The Journal of Biological Chemistry* 288
2 (19): 13165–72. <https://doi.org/10.1074/jbc.R113.455311>.
- 3 Miller, M. W., J. Karazsia, C. E. Groves, S. Griffin, T. Moore, P. Wilber, and K. Gregg. 2016.
4 “Detecting Sedimentation Impacts to Coral Reefs Resulting from Dredging the Port of
5 Miami, Florida USA.” *PeerJ* 4 (November): e2711. <https://doi.org/10.7717/peerj.2711>.
- 6 Moon, S. Y., and Y. Zheng. 2003. “Rho GTPase-Activating Proteins in Cell Regulation.” *Trends*
7 *in Cell Biology* 13 (1): 13–22. [https://doi.org/10.1016/s0962-8924\(02\)00004-1](https://doi.org/10.1016/s0962-8924(02)00004-1).
- 8 Morgan, M. B., S. E. Edge, and T. W. Snell. 2005. “Profiling Differential Gene Expression of
9 Corals along a Transect of Waters Adjacent to the Bermuda Municipal Dump.” *Marine*
10 *Pollution Bulletin* 51 (5-7): 524–33. <https://doi.org/10.1016/j.marpolbul.2005.09.023>.
- 11 Muscatine, L., P. G. Falkowski, J. W. Porter, and Z. Dubinsky. 1984. “Fate of Photosynthetic
12 Fixed Carbon in Light- and Shade-Adapted Colonies of the Symbiotic Coral *Stylophora*
13 *pistillata*.” *Proc. R. Soc. Land. B* 222: 181–202.
- 14 Muscatine, L., and J. W. Porter. 1977. “Reef Corals: Mutualistic Symbioses Adapted to Nutrient-
15 Poor Environments.” *Bioscience* 27 (7): 454–60. <https://doi.org/10.2307/1297526>.
- 16 Ogston, A.S., and M. E. Field. 2010. “Predictions of Turbidity Due to Enhanced Sediment
17 Resuspension Resulting from Sea-Level Rise on a Fringing Coral Reef: Evidence from
18 Molokāi, Hawaiī.” *Journal of Coastal Research* 26 (6): 1027–37.
19 <https://doi.org/10.2112/JCOASTRES-D-09-00064.1>.
- 20 Padilla-Gamiño, J. L., and R. D. Gates. 2012. “Spawning Dynamics in the Hawaiian Reef-
21 Building Coral *Montipora capitata*.” *Marine Ecology Progress Series* 449 (March 2014):
22 145–60. <https://doi.org/10.3354/meps09530>.
- 23 Pastorok, R. A., and G. R. Bilyard. 1985. “Effects of Sewage Pollution on Coral-Reef
24 Communities.” *Marine Ecology Progress Series* 21 (1/2): 175–89.
25 <http://www.jstor.org/stable/24816929>.
- 26 PBS&J. 2008. “Best Management Practices (BMPs) for Construction, Dredge and Fill and Other
27 Activities Adjacent to Coral Reefs.” The Southeast Florida Coral Reef Initiative Maritime
28 Industry and Coastal Construction Impacts Focus Team.
- 29 Perteau, G., and M. Perteau. 2020. “GFF Utilities: GffRead and GffCompare.” *F1000Research* 9
30 (304): 304. <https://doi.org/10.12688/f1000research.23297.2>.
- 31 Perteau, M., G. M. Perteau, C. M. Antonescu, T. Chang, J. T. Mendell, and S. L. Salzberg. 2015.
32 “StringTie Enables Improved Reconstruction of a Transcriptome from RNA-Seq Reads.”
33 *Nature Biotechnology* 33 (3): 290–95. <https://doi.org/10.1038/nbt.3122>.
- 34 Philipp, E., and K. Fabricius. 2003. “Photophysiological Stress in Scleractinian Corals in
35 Response to Short-Term Sedimentation.” *Journal of Experimental Marine Biology and*
36 *Ecology* 287 (1): 57–78. [https://doi.org/10.1016/S0022-0981\(02\)00495-1](https://doi.org/10.1016/S0022-0981(02)00495-1).
- 37 Philpott, D. E., and P. Person. 1970. “The Biology of Cartilage. II. Invertebrate Cartilages: Squid
38 Head Cartilage.” *Journal of Morphology* 131 (4): 417–30.
39 <https://doi.org/10.1002/jmor.1051310405>.
- 40 Phuyal, S., and H. Farhan. 2019. “Multifaceted Rho GTPase Signaling at the Endomembranes.”
41 *Frontiers in Cell and Developmental Biology* 7 (July): 127.
42 <https://doi.org/10.3389/fcell.2019.00127>.
- 43 Piniak, G. A., and E. K. Brown. 2008. “Growth and Mortality of Coral Transplants (*Pocillopora*
44 *damicornis*) along a Range of Sediment Influence in Mau‘i, Hawai‘i.” *Pacific Science* 62
45 (1): 39–55. [https://doi.org/10.2984/1534-6188\(2008\)62\[39:gamoct\]2.0.co;2](https://doi.org/10.2984/1534-6188(2008)62[39:gamoct]2.0.co;2).
- 46 Pollock, F. J., J. B. Lamb, S. N. Field, S. F. Heron, B. Schaffelke, G. Shedrawi, D. G. Bourne,

Sediment stress across location and morphology

- 1 and B. L. Willis. 2014. “Sediment and Turbidity Associated with Offshore Dredging
2 Increase Coral Disease Prevalence on Nearby Reefs.” *PloS One* 9 (7): e102498.
3 <https://doi.org/10.1371/journal.pone.0102498>.
- 4 Poquita-Du, R. C., Y. L. Goh, D. Huang, L. M. Chou, and P. A. Todd. 2020. “Gene Expression
5 and Photo-Physiological Changes in *Pocillopora acuta* Coral Holobiont Following Heat
6 Stress and Recovery.” *Microorganisms* 8 (8): 1227.
7 <https://doi.org/10.3390/microorganisms8081227>.
- 8 Poquita-Du, R. C., D. Huang, L. M. Chou, Mrinalini, and P. A. Todd. 2019. “Short Term
9 Exposure to Heat and Sediment Triggers Changes in Coral Gene Expression and Photo-
10 Physiological Performance.” *Frontiers in Marine Science* 6 (MAR): 1–15.
11 <https://doi.org/10.3389/fmars.2019.00121>.
- 12 Prada, C., B. Hanna, A. F. Budd, C. M. Woodley, J. Schmutz, J. Grimwood, R. Iglesias-Prieto, et
13 al. 2016. “Empty Niches after Extinctions Increase Population Sizes of Modern Corals.”
14 *Current Biology: CB* 26 (23): 3190–94. <https://doi.org/10.1016/j.cub.2016.09.039>.
- 15 Putnam, H. M., J. M. Davidson, and R. D. Gates. 2016. “Ocean Acidification Influences Host
16 DNA Methylation and Phenotypic Plasticity in Environmentally Susceptible Corals.”
17 *Evolutionary Applications* 9 (9): 1165–78. <https://doi.org/10.1111/eva.12408>.
- 18 Reaka-Kudla, M. L. 1997. “The Global Biodiversity of Coral Reefs: A Comparison with Rain
19 Forests.” *Biodiversity II: Understanding and Protecting Our Biological Resources*.
20 <https://doi.org/10.2307/1791071>.
- 21 Richmond, R. H., and P. L. Jokiel. 1984. “Lunar Periodicity in Larva Release in the Reef Coral
22 *Pocillopora damicornis* at Enewetak and Hawaii.” *Bulletin of Marine Science* 34 (2):
23 280–87.
24 <https://www.ingentaconnect.com/content/umrsmas/bullmar/1984/00000034/00000002/art00>
25 010.
- 26 Riegl, B., and G. M. Branch. 1995. “Effects of Sediment on the Energy Budgets of Four
27 Scleractinian (Bourne 1900) and Five Alcyonacean (Lamouroux 1816) Corals.” *Journal of*
28 *Experimental Marine Biology and Ecology* 186 (2): 259–75. <https://doi.org/10.1016/0022->
29 0981(94)00164-9.
- 30 Robbins, S. J., C. M. Singleton, C. X. Chan, L. F. Messer, A. U. Geers, H. Ying, A. Baker, et al.
31 2019. “A Genomic View of the Reef-Building Coral *Porites lutea* and Its Microbial
32 Symbionts.” *Nature Microbiology* 4 (12): 2090–2100. <https://doi.org/10.1038/s41564-019->
33 0532-4.
- 34 Rodríguez-Casariego, Javier A., A. E. Mercado-Molina, and D. Garcia-Souto. 2020. “Genome-
35 Wide DNA Methylation Analysis Reveals a Conserved Epigenetic Response to Seasonal
36 Environmental Variation in the Staghorn Coral *Acropora cervicornis*.” *Frontiers in Marine*
37 *Science* 7 (September): 1–17. <https://doi.org/10.3389/fmars.2020.560424>.
- 38 Rogers, C.S. 1979. “The Effect of Shading on Coral Reef Structure and Function”. *Journal of*
39 *Experimental Marine Biology and Ecology* 41 (3): 269-288. <https://doi.org/10.1016/0022->
40 0981(79)90136-9.
- 41 Rogers, C. S. 1983. “Sublethal and Lethal Effects of Sediments Applied to Common Caribbean
42 Reef Corals in the Field.” *Marine Pollution Bulletin* 14 (10): 378–82.
43 [https://doi.org/10.1016/0025-326X\(83\)90602-1](https://doi.org/10.1016/0025-326X(83)90602-1).
- 44 Rogers, C. S. 1990. “Responses of Coral Reefs and Reef Organisms to Sedimentation.” *Marine*
45 *Ecology Progress Series* 62: 185–202. <https://doi.org/10.3354/meps062185>.
- 46 Roth, M. S. 2014. “The Engine of the Reef: Photobiology of the Coral-Algal Symbiosis.”

Sediment stress across location and morphology

- 1 *Frontiers in Microbiology* 5 (August): 422. <https://doi.org/10.3389/fmicb.2014.00422>.
- 2 Rushmore, M. E., C. Ross, and N. D. Fogarty. 2021. “Physiological Responses to Short-Term
3 Sediment Exposure in Adults of the Caribbean Coral *Montastraea cavernosa* and Adults
4 and Recruits of *Porites astreoides*.” *Coral Reefs*, July. [https://doi.org/10.1007/s00338-021-](https://doi.org/10.1007/s00338-021-02156-0)
5 02156-0.
- 6 Schlaefer, J. A., S. B. Tebbett, and D. R. Bellwood. 2021. “The Study of Sediments on Coral
7 Reefs: A Hydrodynamic Perspective.” *Marine Pollution Bulletin* 169 (August): 112580.
8 <https://doi.org/10.1016/j.marpolbul.2021.112580>.
- 9 Sheridan, C., P. Grosjean, J. Leblud, C. V. Palmer, A. Kushmaro, and I. Eeckhaut. 2014.
10 “Sedimentation Rapidly Induces an Immune Response and Depletes Energy Stores in a
11 Hard Coral.” *Coral Reefs* 33 (4): 1067–76. <https://doi.org/10.1007/s00338-014-1202-x>.
- 12 Siebert, C., J. C. Pett-Ridge, S. Opfergelt, R. A. Guicharnaud, A. N. Halliday, and K. W. Burton.
13 2015. “Molybdenum Isotope Fractionation in Soils: Influence of Redox Conditions, Organic
14 Matter, and Atmospheric Inputs.” *Geochimica et Cosmochimica Acta* 162 (August): 1–24.
15 <https://doi.org/10.1016/j.gca.2015.04.007>.
- 16 Stafford-Smith, M. G. 1993. “Sediment-Rejection Efficiency of 22 Species of Australian
17 Scleractinian Corals.” *Marine Biology* 115 (2): 229–43.
18 <https://doi.org/10.1007/BF00346340>.
- 19 Stafford-Smith, M. G., and R. F. G. Ormond. 1992. “Sediment-Rejection Mechanisms of 42
20 Species of Australian Scleractinian Corals.” *Marine and Freshwater Research* 43 (4): 683–
21 705. <https://doi.org/10.1071/mf9920683>.
- 22 Stephens, T. G., E. L. Strand, A. R. Mohamed, A. Williams, E. N. Chiles, X. Su, D.
23 Bhattacharya, and H. M. Putnam. 2021. “Ploidy Variation and Its Implications for
24 Reproduction and Population Dynamics in Two Sympatric Hawaiian Coral Species.”
25 *bioRxiv*. <https://doi.org/10.1101/2021.11.21.469467>.
- 26 Stimson, J. S. 1978. “Mode and Timing of Reproduction in Some Common Hermatypic Corals
27 of Hawaii and Enewetak.” *Marine Biology* 48 (2): 173–84.
28 <https://doi.org/10.1007/BF00395017>.
- 29 Studivan, M. S., A. M. Rossin, E. Rubin, N. Soderberg, D. M. Holstein, and I. C. Enochs. 2022.
30 “Reef Sediments Can Act As a Stony Coral Tissue Loss Disease Vector.” *Frontiers in*
31 *Marine Science* 8. <https://doi.org/10.3389/fmars.2021.815698>.
- 32 Sumaila, U. R., and A. M. Cisneros-Montemayor. 2010. “A Global Estimate of Benefits from
33 Ecosystem-Based Marine Recreation: Potential Impacts and Implications for Management.”
34 *Journal of Bioeconomics* 12 (3): 245–68. <https://doi.org/10.1007/s10818-010-9092-7>.
- 35 Tambutté, S., M. Holcomb, C. Ferrier-Pagès, S. Reynaud, E. Tambutté, D. Zoccola, and D.
36 Allemand. 2011. “Coral Biomineralization: From the Gene to the Environment.” *Journal of*
37 *Experimental Marine Biology and Ecology* 408 (1): 58–78.
38 <https://doi.org/10.1016/j.jembe.2011.07.026>.
- 39 Tomanek, L.. 2015. “Proteomic Responses to Environmentally Induced Oxidative Stress.” *The*
40 *Journal of Experimental Biology* 218 (12): 1867–79. <https://doi.org/10.1242/jeb.116475>.
- 41 Vandermeulen, J. H., and N. Watabe. 1973. “Studies on Reef Corals. Skeleton Formation by
42 Newly Settled Planula Larva of *Pocillopora damicornis*.” *Marine Biology* 23 (1): 47–57.
43 <https://doi.org/10.1007/bf00394111>.
- 44 Weber, M., D. de Beer, C. Lott, L. Polerecky, K. Kohls, R. M. M. Abed, T. G. Ferdelman, and
45 K. E. Fabricius. 2012. “Mechanisms of Damage to Corals Exposed to Sedimentation.”
46 *Proceedings of the National Academy of Sciences of the United States of America* 109 (24):

Sediment stress across location and morphology

- 1 E1558–67. <https://doi.org/10.1073/pnas.1100715109>.
- 2 Weber, M., C. Lott, and K. E. Fabricius. 2006. “Sedimentation Stress in a Scleractinian Coral
3 Exposed to Terrestrial and Marine Sediments with Contrasting Physical, Organic and
4 Geochemical Properties.” *Journal of Experimental Marine Biology and Ecology* 336 (1):
5 18–32. <https://doi.org/10.1016/j.jembe.2006.04.007>.
- 6 Westbroek, P., J. Yanagida, and Y. Isa. 1980. “Functional Morphology of Brachiopod and Coral
7 Skeletal Structures Supporting Ciliated Epithelia.” *Paleobiology* 6 (3): 313–30.
8 <http://www.jstor.org/stable/2400348>.
- 9 Wiens, M., M. S. Ammar, A. H. Nawar, C. Koziol, H. M. Hassanein, M. Eisinger, I. M. Müller,
10 and W. E. Müller. 2000. “Induction of Heat-Shock (stress) Protein Gene Expression by
11 Selected Natural and Anthropogenic Disturbances in the Octocoral *Dendronephthya*
12 *klunzingeri*.” *Journal of Experimental Marine Biology and Ecology* 245 (2): 265–76.
13 [https://doi.org/10.1016/s0022-0981\(99\)00167-7](https://doi.org/10.1016/s0022-0981(99)00167-7).
- 14 Young, M. D., M. J. Wakefield, G. K. Smyth, and A. Oshlack. 2010. “Gene Ontology Analysis
15 for RNA-Seq: Accounting for Selection Bias.” *Genome Biology* 11 (2).
16 <https://doi.org/10.1186/gb-2010-11-2-r14>.
- 17 Zdunek-Zastocka, E., and M. Sobczak. 2013. “Expression of *Pisum Sativum* PsAO3 Gene,
18 Which Encodes an Aldehyde Oxidase Utilizing Abscisic Aldehyde, Is Induced under
19 Progressively but Not Rapidly Imposed Drought Stress.” *Plant Physiology and*
20 *Biochemistry: PPB / Societe Francaise de Physiologie Vegetale* 71 (October): 57–66.
21 <https://doi.org/10.1016/j.plaphy.2013.06.027>.

22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46

Sediment stress across location and morphology

1 **Data availability**

2 All code for analyses is on GitHub (<https://github.com/JillAshey/SedimentStress> and
3 <https://github.com/JillAshey/FunctionalAnnotation>). Archived references used in this study are
4 located at <https://osf.io/8qn6c/> (DOI = 10.17605/OSF.IO/8QN6C). Raw sequences are housed
5 under the NCBI BioProject Accession PRJNA911752.

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

Sediment stress across location and morphology

1 **Author contributions**

2 FS, RHR, and VP conceptualized and performed the experiments. JA, HM, JF, and PS analyzed
3 and validated the data. JA, LC, RHR, VP, FS, and HMP acquired funding. LC, RHR, VP, FS,
4 and HMP provided experimental and computational resources. JA wrote the original draft of the
5 manuscript. All authors edited and wrote subsequent drafts of the manuscript.

6
7
8

9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41

Sediment stress across location and morphology

1 **Tables**

2 Table 1: Summary of read sequencing, quality and alignment by species. See Supplementary

3 Table 1 for more details.

Species	Location	# of samples	Avg. Raw Reads	Avg. Clean Reads	Avg. # of Reads Removed	Avg. % of Reads Mapped
<i>A. cervicornis</i>	Florida	13	16,630,621	16,499,909	130,712	70.62
<i>M. cavernosa</i>	Florida	15	15,921,959	15,740,195	181,765	50.18
<i>O. faveolata</i>	Florida	14	14,111,857	13,905,076	206,781	65.54
<i>M. capitata</i>	Hawaii	11	16,360,670	16,280,212	80,459	63.75
<i>P. acuta</i>	Hawaii	12	14,441,259	14,359,654	81,605	68.94
<i>P. lobata</i>	Hawaii	12	13,995,393	13,858,883	136,510	52.83

4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33

Sediment stress across location and morphology

- 1 Table 2: Summary of differentially expressed genes (DEGs), gene ontology (GO) terms and
 2 orthogroups by species. See Supplementary Tables 2, 3, 4 & 5 for more details.

Species	Location	Morphology	Unique DEGs	Up regulated DEGs	Down-regulated DEGs	Unique GO terms assigned to DEGs	Orthogroups containing DEGs	DEGs in orthogrou
<i>A. cervicornis</i>	Florida	Branching	215	67	154	278	102	119
<i>M. cavernosa</i>	Florida	Massive	62	19	44	158	28	31
<i>O. faveolata</i>	Florida	Massive	8	0	8	27	2	2
<i>M. capitata</i>	Hawaiꞏi	Intermediate	157	79	78	237	80	87
<i>P. acuta</i>	Hawaiꞏi	Branching	263	129	135	380	100	123
<i>P. lobata</i>	Hawaiꞏi	Massive	153	32	128	198	56	66

3
 4
 5
 6
 7
 8
 9
 10
 11
 12
 13
 14
 15
 16
 17
 18
 19
 20
 21
 22
 23
 24
 25
 26
 27
 28

Sediment stress across location and morphology

- 1 Table 3: Shared Biological Processes gene ontology (GO) terms between species. See
- 2 Supplementary Tables 2 & 3 for more details.

Species	# of shared GO terms	GO slim Category	GO term
<i>A. cervicornis</i> + <i>M. cavernosa</i> + <i>M. capitata</i>	1	Developmental processes	positive regulation of skeletal muscle tissue development (GO:0048643)
<i>M. capitata</i> + <i>M. cavernosa</i> + <i>P. acuta</i>	1	Transport	riboflavin transport (GO:0032218)
<i>M. capitata</i> + <i>P. acuta</i> + <i>P. lobata</i>	1	Other biological processes	Microtubule-based process (GO:0007017)
<i>A. cervicornis</i> + <i>M. cavernosa</i>	11	cell-cell signaling & transport	regulation of neurotransmitter uptake (GO:0051580)
		developmental processes	ovarian cumulus expansion (GO:0001550)
		other metabolic processes	diadenosine tetraphosphate biosynthetic process (GO:0015966)
		signal transduction	regulation of Rho protein signal transduction (GO:0035023)
		No GO slim category identified	negative regulation of chondrocyte proliferation (GO:1902731)
			positive regulation of cell junction assembly (GO:1901890)
			negative regulation of histone H4-K16 acetylation (GO:2000619)
			plasma membrane raft assembly (GO:0044854)
cellular response to exogenous dsRNA (GO:0071360)			
cellular response to platelet-derived growth factor stimulus (GO:0036120)			

Sediment stress across location and morphology

			positive regulation of extrinsic apoptotic signaling pathway via death domain receptors (GO:1902043)
<i>A. cervicornis</i> + <i>O. faveolata</i>	2	developmental processes	chondrocyte development (GO:0002063)
		signal transduction	positive regulation of signal transduction (GO:0009967)
<i>A. cervicornis</i> + <i>M. capitata</i>	1	No GO slim category identified	positive regulation of adipose tissue development (GO:1904179)
<i>A. cervicornis</i> + <i>P. acuta</i>	5	other biological processes	maintenance of gastrointestinal epithelium (GO:0030277)
		signal transduction	Wnt receptor signaling pathway (GO:0016055)
		protein metabolism	Mo-molybdopterin cofactor biosynthetic process (GO:0006777)
		stress response	cellular response to starvation (GO:0009267)
		No GO slim category identified	mesenchymal stem cell maintenance involved in nephron morphogenesis (GO:0072038)
<i>A. cervicornis</i> + <i>P. lobata</i>	3	RNA metabolism	positive regulation of transcription by RNA polymerase II (GO:0045944) DNA methylation-dependent heterochromatin assembly (GO:0006346)
		cell organization and biogenesis	DNA methylation-dependent heterochromatin assembly (GO:0006346)
<i>M. capitata</i> + <i>M. cavernosa</i>	4	cell organization and biogenesis	positive regulation of neuron projection development (GO:0010976)
		developmental processes	positive regulation of neuron projection development (GO:0010976) retina development in camera-type eye (GO:0060041)

Sediment stress across location and morphology

		other biological processes	negative regulation of mitochondrial membrane potential (GO:0010917)
		No GO slim category identified	ventricular compact myocardium morphogenesis (GO:0003223)
<i>M. capitata</i> + <i>P. acuta</i>	20	cell adhesion	heterophilic cell adhesion (GO:0007157)
		developmental processes	embryonic hindlimb morphogenesis (GO:0035116)
			positive regulation of osteoclast differentiation (GO:0045672)
		other biological processes	positive regulation of bone resorption (GO:0045780)
			positive regulation of cellular pH reduction (GO:0032849)
			oxidation reduction (GO:0055114)
			response to activity (GO:0014823)
			response to glucagon stimulus (GO:0033762)
			response to pH (GO:0009268)
		RNA metabolism	positive regulation of transcription factor activity (GO:0051091)
		signal transduction	blue light signaling pathway (GO:0009785)
		stress response	response to pain (GO:0048265)
		transport	carbon dioxide transport (GO:0015670)
			secretion (GO:0046903)

Sediment stress across location and morphology

		No GO slim category identified	angiotensin-activated signaling pathway (GO:0038166)
			cellular response to retinoic acid (GO:0071300)
			positive regulation of dipeptide transmembrane transport (GO:2001150)
			positive regulation of mitochondrial membrane permeability (GO:0035794)
			negative regulation of glucocorticoid secretion (GO:2000850)
			regulation of chloride transport (GO:2001225)
<i>M. cavernosa</i> + <i>P. acuta</i>	2	developmental processes	ectoderm development (GO:0007398)
		No GO slim category identified	anterior head development (GO:0097065)
<i>M. cavernosa</i> + <i>P. lobata</i>	3	developmental processes	somatic muscle development (GO:0007525)
		No GO slim category identified	regulation of cell junction assembly (GO:1901888)
			regulation of protein serine/threonine kinase activity (GO:0071900)
<i>O. faveolata</i> + <i>P. acuta</i>	1	other biological processes	response to stimulus (GO:0050896)
<i>P. acuta</i> + <i>P. lobata</i>	11	cell-cell signaling	Spemann organizer formation (GO:0060061)
		cell organization and biogenesis	neural crest cell fate specification (GO:0014036)
		developmental processes	embryonic axis specification (GO:0000578)
embryonic organ development (GO:0048568)			

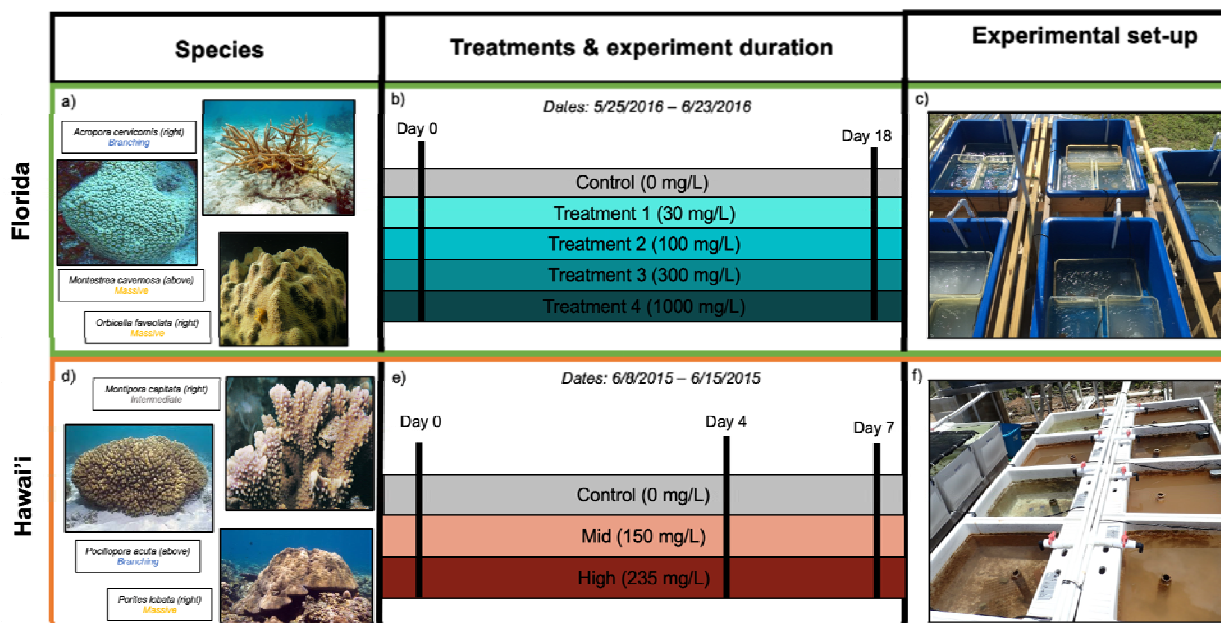
Sediment stress across location and morphology

			eye development (GO:0001654)
			neural crest cell fate specification (GO:0014036)
			neural plate anterior/posterior regionalization (GO:0021999)
		other biological processes	regulation of circadian rhythm (GO:0042752)
		No GO slim category identified	canonical Wnt signaling pathway involved in neural crest cell differentiation (GO:0044335)
			cellular hypotonic response (GO:0071476)
			endocardial cushion development (GO:0003197)
			negative regulation of cardiac cell fate specification (GO:2000044)
<i>A. cervicornis</i>	256	See Supplementary Tables S2 & S3	
<i>M. capitata</i>	209	See Supplementary Tables S2 & S3	
<i>M. cavernosa</i>	136	See Supplementary Tables S2 & S3	
<i>O. faveolata</i>	24	See Supplementary Tables S2 & S3	
<i>P. acuta</i>	339	See Supplementary Tables S2 & S3	
<i>P. lobata</i>	181	See Supplementary Tables S2 & S3	

1
2
3
4
5
6
7

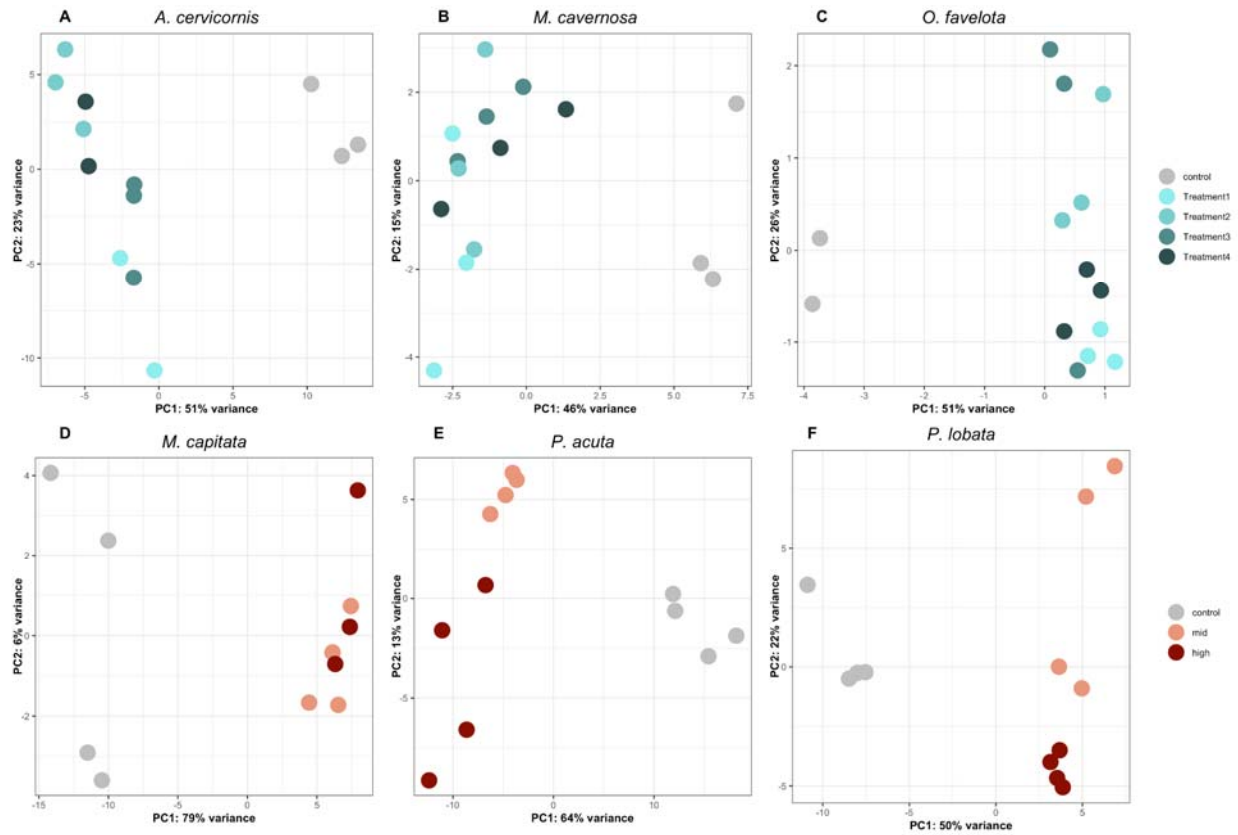
Sediment stress across location and morphology

1 Figures



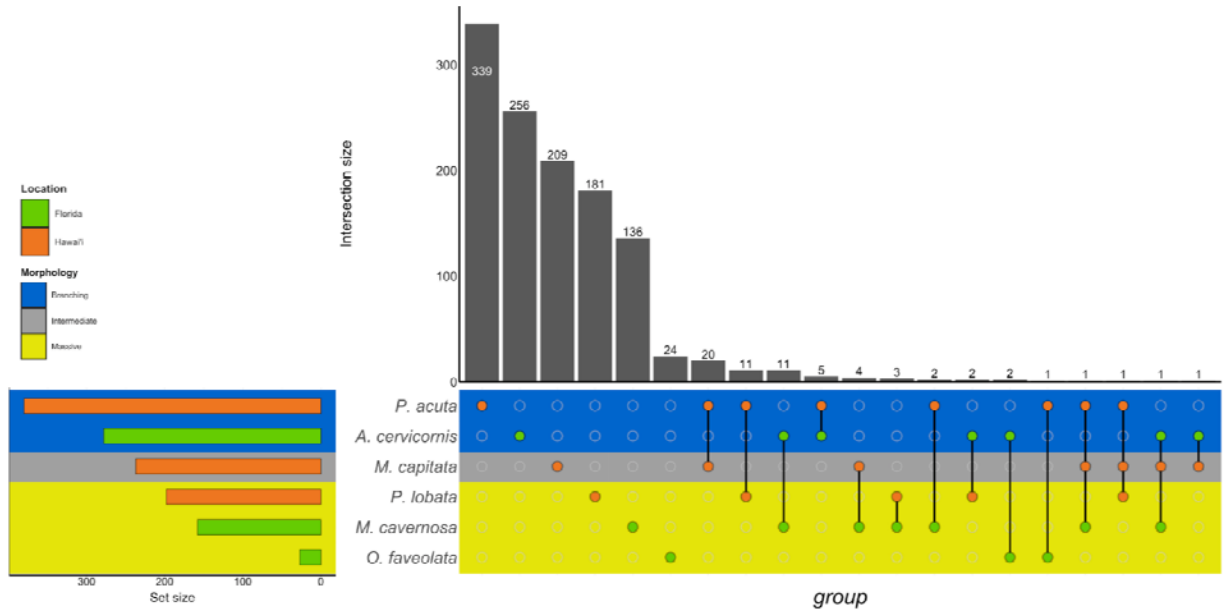
2
3 Figure 1. Experimental design for Florida (a-c) and Hawai'i (d-f) experiments. (a) Species used
4 and their morphologies in parentheses for Florida study: *Acropora cervicornis* (branching
5 morphology), *Montastrea cavernosa* (massive morphology) and *Orbicella faveolata* (massive
6 morphology). (b) Illustration of timeline and sediment treatments. (c) Experimental tanks from
7 the Florida experiment (image by Francois Seneca). (d) Species used and their morphologies in
8 parentheses for Hawai'i study: *Montipora capitata* (intermediate morphology), *Pocillopora*
9 *acuta* (branching morphology) and *Porites lobata* (massive morphology) (e) Illustration of
10 timeline and sediment treatments. (f) Experimental tanks from the Hawai'i experiment (image by
11 Francois Seneca). Images of coral species were obtained from Corals of the World
12 (<http://www.coralsoftheworld.org>)
13
14
15
16
17

Sediment stress across location and morphology



1
2 Figure 2. Principal component analysis (PCA) of differentially expressed genes in (A) *A.*
3 *cervicornis*, (B) *M. cavernosa*, (C) *O. faveolata*, (D) *M. capitata*, (E) *P. acuta*, and (F) *P. lobata*.
4
5
6
7

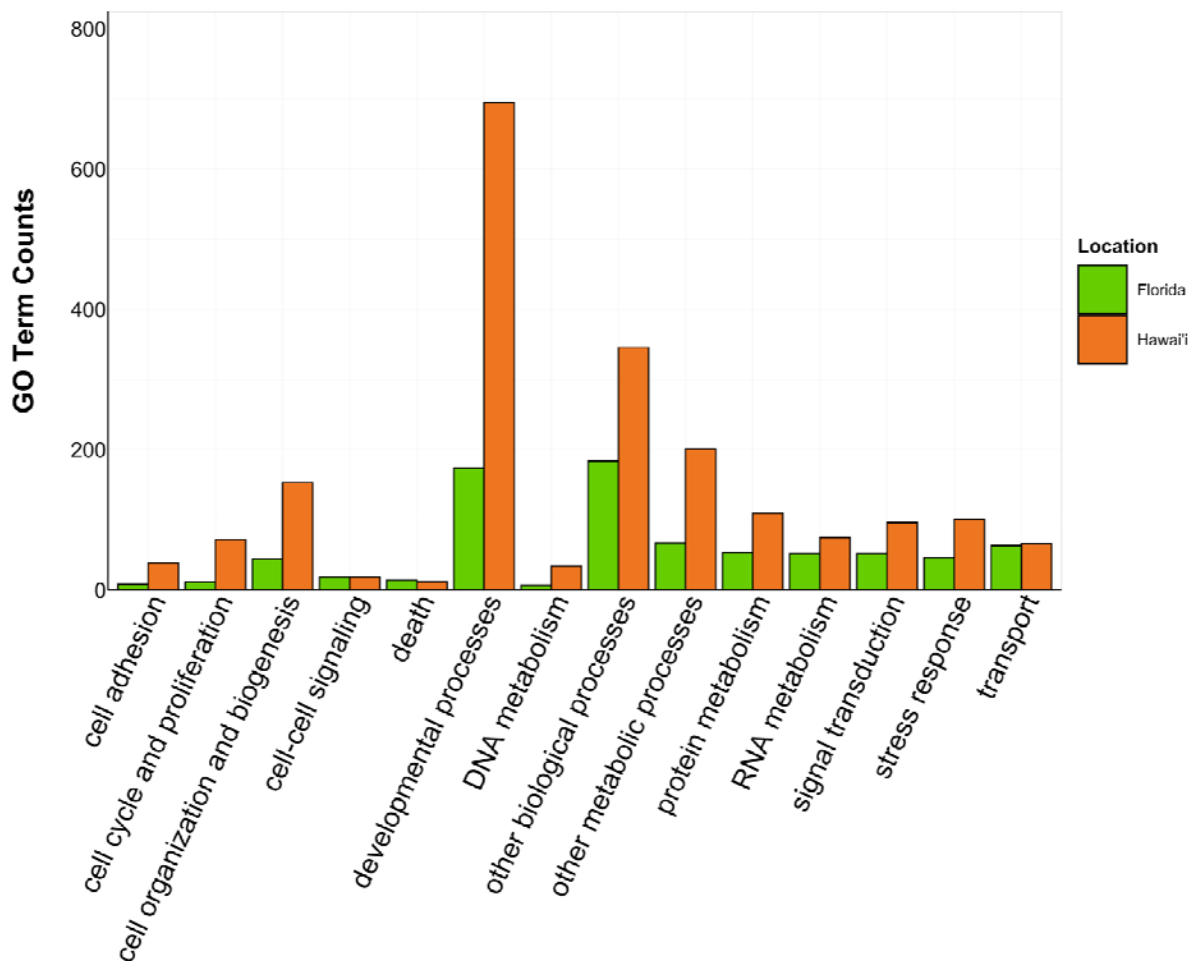
Sediment stress across location and morphology



1
 2 Figure 3. UpSet plot of the intersection of Biological Process GO terms across location (green
 3 bar or circle = Florida; orange bar or circle = Hawaii) and morphology (blue stripe =
 4 branching; gray stripe = intermediate; yellow stripe = massive).

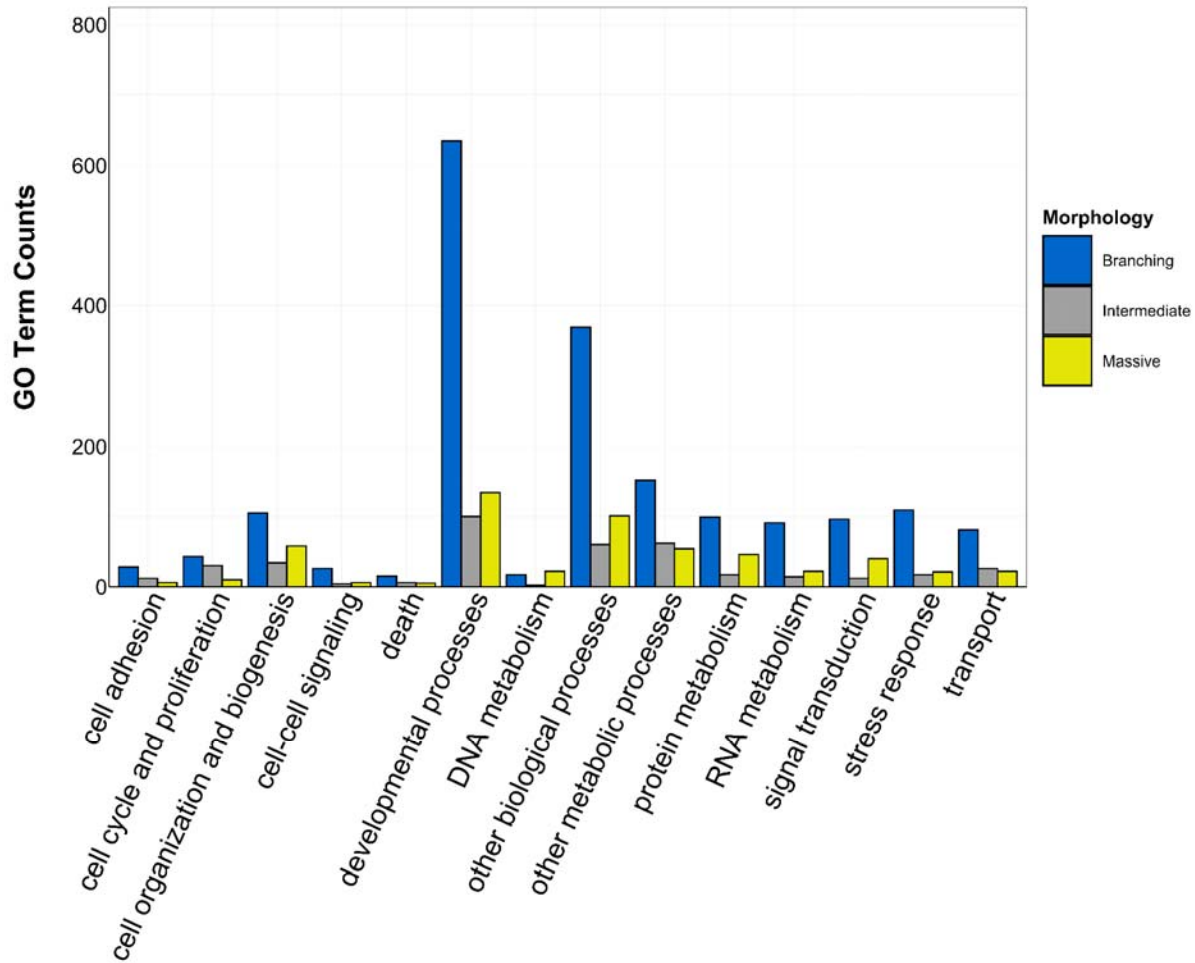
5
 6
 7

Sediment stress across location and morphology



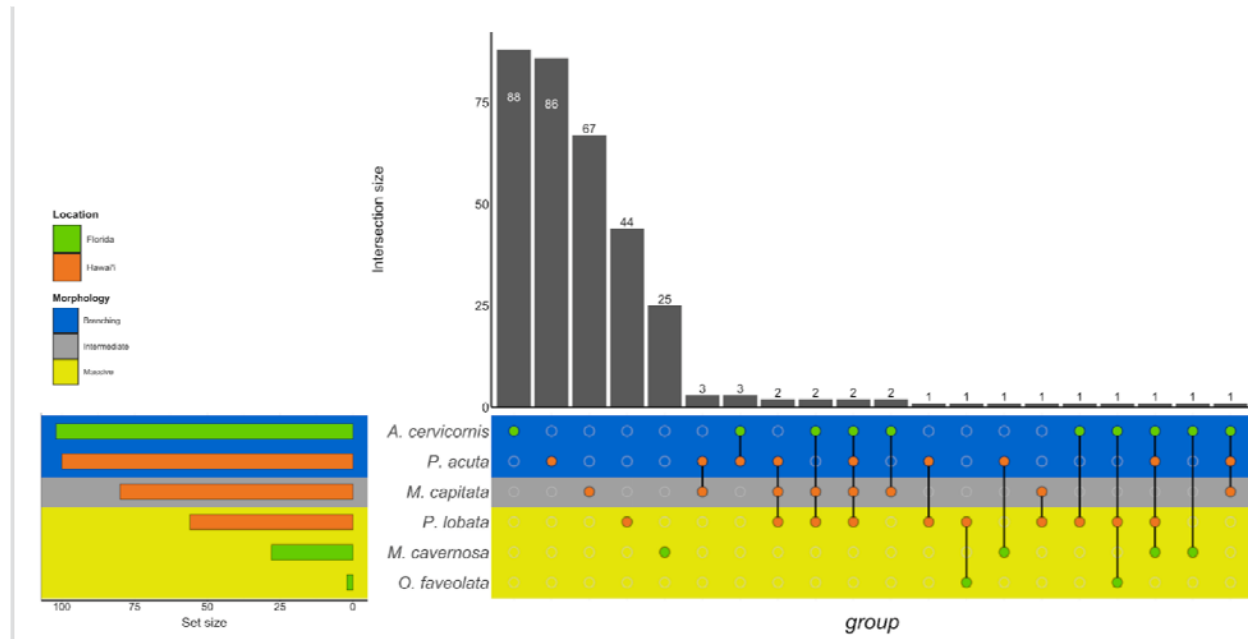
1
2 Figure 4. Counts of Biological Process GO terms grouped under GO slim categories by location.
3 GO slim categories are on the x-axis, while the number of Biological Processes GO terms in
4 each GO slim category is on the y-axis. The bars are colored by location: green bar = Florida,
5 orange bar = Hawai'i.
6

Sediment stress across location and morphology



1
2 Figure 5. Counts of Biological Process GO terms grouped under GO slim categories by
3 morphology. GO slim categories are on the x-axis, while the number of Biological Processes GO
4 terms in each GO slim category is on the y-axis. The bars are colored by morphology: blue bar =
5 branching, gray bar = intermediate, yellow bar = massive.
6
7
8
9

Sediment stress across location and morphology



1
2 Figure 6. UpSet plot of the intersection of orthogroups across location (green bar or circle =
3 Florida; orange bar or circle = Hawaii) and morphology (blue stripe = branching; gray stripe =
4 intermediate; yellow stripe = massive).

5
6
7
8
9
10