

1 **Performance of *Orbicella faveolata* larval cohorts does not align with previously**
2 **observed thermal tolerance of adult source populations**

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45 **Abstract**

46 *Orbicella faveolata*, commonly known as the mountainous star coral, is a dominant reef-building
47 species in the Caribbean, but populations have suffered sharp declines since the 1980s due to
48 repeated bleaching and disease-driven mortality. Prior research has shown that inshore adult *O.*
49 *faveolata* populations in the Florida Keys are able to maintain high coral cover and recover from
50 bleaching faster than their offshore counterparts. However, whether this origin-specific variation
51 in thermal resistance is heritable remains unclear. To address this knowledge gap, we produced
52 purebred and hybrid larval crosses from *O. faveolata* gametes collected at two distinct reefs in
53 the Upper Florida Keys, a nearshore site (Cheeca Rocks, CR) and an offshore site (Horseshoe
54 Reef, HR), in two different years (2019, 2021). We then subjected these aposymbiotic larvae to
55 severe (36 °C) and moderate (32 °C) heat challenges to quantify their thermal tolerance.
56 Contrary to our expectation based on patterns of adult thermal tolerance, HR purebred larvae
57 survived better and exhibited gene expression profiles that were less driven by stress response
58 under elevated temperature compared to purebred CR and hybrid larvae. One potential
59 explanation could be compromised reproductive output of CR adult colonies due to repeated
60 summer bleaching events in 2018 and 2019, as gametes originating from CR in 2019 contained
61 less storage lipids than those from HR. These findings provide an important counter-example to
62 the current selective breeding paradigm, that more tolerant parents will yield more tolerant
63 offspring, and highlight the importance of adopting a holistic approach when evaluating larval
64 quality for conservation and restoration purposes.

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71 **Introduction**

72 Global ecosystems are undergoing unprecedented structural and functional changes as
73 atmospheric CO₂ level and temperature continue to rise in the Anthropocene (Steffen et al.,
74 2007). One ecosystem that is particularly vulnerable to these changes is coral reefs, because
75 most reef-building corals are found in the tropics (Spalding & Brown, 2015) and already live
76 close to their upper thermal limits (Baker et al., 2008). A small temperature increase, as little as
77 1 °C above the maximum monthly mean temperature for a period of four weeks, or four degree
78 heating weeks (Liu et al., 2005), can lead to the breakdown of the symbiotic relationship
79 between the cnidarian animal host and their intracellular photosynthetic dinoflagellate algae.
80 This phenomenon is commonly known as coral bleaching (Hoegh-Guldberg et al., 2007; Lesser,
81 2011). Worldwide, coral cover is estimated to have declined by 20% over the past 30 years and
82 reefs will continue to be threatened by large-scale bleaching events even with climate
83 intervention strategies (Hoegh-Guldberg et al., 2019). Similar to the pattern observed in the
84 wider Caribbean region (Gardner et al., 2003), coral reefs in the Florida Keys have experienced
85 drastic population declines since the early 1980s mostly due to bleaching and disease (Dustan
86 & Halas, 1987; Precht & Miller, 2007). The two most recent large-scale bleaching events to
87 affect this region occurred in 2014 and 2015 when maximum temperatures exceeded local
88 bleaching thresholds for over 4-8 weeks (Eakin et al., 2019; Smith et al., 2019).

89
90 However, not all corals are equally susceptible to bleaching. Coral populations inhabiting
91 thermally-challenging environments, characterized by elevated temperatures and/or greater
92 temperature variabilities, have been repeatedly shown to exhibit higher tolerance to heat stress
93 (Howells et al., 2016; Thomas et al., 2018). Mechanistically, increased temperature tolerance
94 can be the result of adaptation and/or acclimatization on the part of coral hosts, their
95 dinoflagellate endosymbionts, or other members of the microbiome (Ainsworth et al., 2016;
96 Berkelmans & van Oppen, 2006; Palumbi et al., 2014; Santoro et al., 2021). Along the Florida

97 Keys reef tract, inshore patch reefs experience higher annual temperature fluctuations and
98 elevated mean temperature during bleaching-prone summer months in comparison to offshore
99 reefs at similar latitudes (Kenkel et al., 2015; Manzello et al., 2015a, 2015b). This spatially-
100 defined thermal heterogeneity has been theorized to support elevated heat tolerance of inshore
101 corals, which aligns with lab-based experiments and field-based observations of reduced
102 bleaching severity of inshore coral populations (Gintert et al., 2018; Kenkel et al., 2013). During
103 the back-to-back bleaching events in 2014 and 2015, inshore *Orbicella faveolata* colonies in the
104 Upper and Lower Florida Keys demonstrated lower bleaching prevalence and higher recovery
105 rate than colonies at paired offshore sites (Manzello et al., 2019). Due to the lack of distinct
106 genetic structure among inshore and offshore host populations, the increased heat tolerance of
107 the inshore corals was attributed to the significantly greater prevalence of heat tolerant
108 symbionts (*Durussdinium trenchii*) in these corals versus those at offshore sites (Manzello et al.,
109 2019). The host role in shaping holobiont thermotolerance in this system remains unclear.

110

111 Establishing the degree to which the coral host contributes to thermal tolerance is essential for
112 modeling adaptive potential and implementing intervention strategies. Assisted evolution was
113 proposed as a suite of intervention approaches to mitigate the decline and degradation of reef
114 systems, given that adaptive changes that occur naturally might not be able to keep pace with
115 the rapidly-changing climate (Van Oppen & Oliver, 2015). Studies in multiple Indo-Pacific coral
116 species provide compelling evidence for host genomic heritability of traits to enhance the
117 tolerance to heat, ocean acidification, and disease (Dixon et al., 2015; Drury et al., 2022;
118 Howells et al., 2021; Quigley et al., 2020). Additional data is needed to better understand the
119 tradeoffs of selecting for a single trait in corals, given their exposure to multiple environmental
120 challenges (Ladd et al., 2017). This approach may also lead to outbreeding depression and
121 genetic swamping, which can threaten the survival and fitness of their offspring (Aitken &
122 Whitlock, 2013). Significant differences in the decline of coral species have been observed in

123 the Pacific and Caribbean regions (Tebbett et al., 2023)). It has been suggested that impaired
124 colony physiology in corals could contribute to suboptimal larval performance that causes
125 recruitment failure (Hughes & Tanner, 2000; Williams et al., 2008).

126

127 We investigate how the source population affects the performance of *O. faveolata* offspring. To
128 do this, we created larval cohorts sourced from parent colonies living in thermally distinct reef
129 sites in the Upper Florida Keys. We then exposed these aposymbiotic larvae to severe and
130 moderate temperature stress. *O. faveolata* is one of the major reef-building corals in the Florida
131 Keys and its populations are highly connected throughout the wider Caribbean seascape
132 ((Rippe et al., 2017)). However, *O. faveolata* populations have suffered sharp declines in the
133 past few decades (Edmunds, 2015). These declines are largely due to bleaching and disease,
134 making recovery challenging (Gladfelter et al., 1978). *Orbicella faveolata* is a hermaphroditic
135 broadcast-spawning coral species that sexually reproduces during late summer months when
136 water temperatures are maximal (Szmant, 1991). Symbiotic dinoflagellates are acquired from
137 the environment during metamorphosis (Coffroth et al., 2001). By working with aposymbiotic
138 larvae, we can study the physiological and transcriptomic basis for heat tolerance in the animal
139 host without the confounding effects of symbiosis. Understanding the physiological and genetic
140 factors underlying origin-dependent bleaching resistance in *O. faveolata* (and congeners) is
141 crucial because they were listed as threatened under the Endangered Species Act in 2014. By
142 studying this, we can assess the impact on future generations and estimate adaptive capacity,
143 enabling informed conservation efforts.

144

145 **Methods**

146 Sites and Temperature Data

147 Two well-monitored sites in the Upper Florida Keys were chosen for subsequent spawning
148 collections in 2019 and 2021, including Cheeca Rocks (CR, 24.8977°N, 80.6182°W) and

149 Horseshoe Reef (HR, 25.1388°N, 80.3133°W). Temperature was measured every 3 hrs at the
150 Cheeca Rocks Moored-Autonomous pCO₂ buoy (MApCO₂, depth = 1 m) using a conductivity–
151 temperature sensor (Model SBE-16 plus v. 2.2, Seabird Electronics). Data were collected every
152 30 min from HR (depth = 3.4 m) using the following loggers: Pendant (from 1/1/2017 to 8/6/2019
153 09:00 h) and Tidbit MX2204 (8/6/19 09:30 h to 12/31/21). From these data, daily average
154 temperature and the running 30-day mean temperatures were calculated for each site from
155 January 1, 2017 to December 31, 2021.

156

157 The bleaching thresholds for CR were determined as previously described (Gintert et al., 2018).
158 Bleaching thresholds can be estimated for reef sites in the Florida Keys by taking the average of
159 the maximum monthly mean sea surface temperature (SST) during a non-bleaching year and
160 the minimum monthly mean SST during a bleaching year (Manzello et al., 2007). Every Florida
161 Keys-wide mass bleaching event since 2005 has been predicted in near-real-time by calculating
162 the running 30-day mean SST from the Molasses Reef Coastal Marine Automated Network (C-
163 MAN) station and using that site as a proxy for the rest of the Florida Keys offshore reef sites
164 (Manzello, 2015). This technique has also been used to predict every bleaching event that has
165 occurred at CR since 2012.

166

167 The bleaching threshold for CR is a monthly mean SST $\geq 31.3^{\circ}\text{C}$, such that once the running
168 30-day mean SST reach this value, bleaching has been observed in 2014, 2015, 2018, and
169 2019 (Gintert et al., 2018; Manzello et al., 2019) (Fig. 1). The monthly mean bleaching threshold
170 for offshore reefs in the Florida Keys is $\geq 30.4^{\circ}\text{C}$, nearly 1°C lower than CR. We lack sufficient
171 coverage of both bleaching and non-bleaching year temperature data for HR to determine a
172 local bleaching threshold for this site. However, we can deduce that temperatures experienced
173 at HR are warmer than the values experienced at the far offshore reefs used for calculation of
174 the Florida Keys-wide threshold, but cooler than CR (Fig. 1).

175

176 Spawning, cross design and larval rearing

177 Spawning collections were conducted under permit FKNMS-2018-163-A1. The first spawn
178 occurred on August 22, 2019, seven days after the full moon. Gamete bundles were collected
179 from five adult colonies within each site using spawning tents with 50 mL collection tubes
180 following standard protocols (Marhaver et al. 2017). Collection tubes were removed from tents
181 after ~5 mL of gametes had been collected, immediately capped, and transported back to the
182 boat by divers. Gametes were then diluted to reach a sperm concentration of $\sim 10^6$ cells/mL in
183 individual 5 gallon buckets filled with 0.2 μm filtered seawater (FSW). Three replicate bulk
184 crosses were created on each boat by mixing equivalent aliquots of diluted gametes from each
185 colony ($n = 5$) at each site ($n = 2$). Approximately 1.5 hrs post spawning, diluted gametes
186 released from two CR colonies and three HR colonies were mixed in the laboratory to create
187 three replicate hybrid bulk crosses. Note that given the optimum fertilization window for gametes
188 (~2 hrs), there was not sufficient time to separate eggs from sperm to attempt a diallel-type
189 crossing design. Therefore, the bulk hybrid crosses could include true CR x HR and HR x CR
190 reciprocal larvae as well as CR x CR and HR x HR fertilizations.

191

192 A second round of *O. faveolata* gamete collection from each site occurred six (28 August, HR
193 only) and seven (29 August, CR and HR) days after the full moon in August 2021. Gametes
194 were again obtained from five colonies at HR on day 6, and handled as in 2019, resulting in
195 three bulk crosses of HR x HR cultures. No spawning was observed at CR on day 6. On day 7,
196 only a single colony was observed spawning at each site, which restricted our fertilization
197 design to the creation of hybrid crosses only. Gamete bundles from each site were returned to
198 the Key Largo Marine Research Laboratory and separated into eggs and sperm by filtration
199 through an 80 μm nitex mesh, followed by rinsing with FSW. Individual eggs and sperm were

200 crossed to create three culture replicates of each of two hybrid crosses (CR sperm x HR egg,
201 HR sperm x CR egg).

202

203 For all crosses in each year, successful fertilization was confirmed through observation of initial
204 cell division under 100x magnification following ~2 hrs of incubation. Developing embryos were
205 gently rinsed 3x in FSW to remove excess sperm and transferred to 6 L culture bins at a density
206 of ~1 embryo per mL in FSW. Healthy developing larvae were rinsed and transferred to fresh
207 FSW twice daily until reaching the planula stage, after which water changes were performed
208 every other day. Larval cultures were maintained at 29°C by placing filled bins in shallow 32 L
209 polycarbonate (Rubbermaid) water baths equipped with 100 W aquarium heaters and SL381
210 submersible pumps (Domica) to maintain ambient temperature consistent with field temperature
211 profiles.

212

213 Thermal stress challenges

214 Two thermal stress experiments were conducted at the Key Largo Marine Research Laboratory
215 in 2019 after mature swimming larvae were observed in all cultures (CRxCR, HRxHR, putative
216 hybrid cross), which occurred on day 3 post fertilization: an acute stress at 36 °C and a
217 moderate stress at 32 °C following (Zhang et al., 2022). For each experiment, six replicate 6 L
218 polycarbonate larval bins (Vigors) were filled with 0.2 µm FSW and placed into a set of two
219 shallow 32 L polycarbonate (Rubbermaid) water baths for temperature control (n = 3 larval bins
220 per bath). Each water bath was filled with ~15 L water and equipped with a SL381 submersible
221 water pump to maintain circulation and a 100 W aquarium heater. Each larval rearing bin was
222 fully filled and two of three bins were equipped with HOBO temperature loggers (Onset). For the
223 acute stress experiment, each larval rearing bin received two groups of ten larvae per bulk cross
224 (n = 6 per cross type per treatment) that were aliquoted into floating netwells (70 µm cell
225 strainers, Grenier Bio-One). The control bath remained at 29 °C and the treatment bath was

226 heated up to 36 °C over 24 hrs (Fig. S1a). Mortality was assessed every 24 hrs by counting the
227 number of surviving larvae in the netwells. The experiment was terminated once mortality
228 reached more than 50% for the majority of the netwells. For the moderate stress experiment, a
229 total of 20 larvae from each bulk cross replicate were aliquoted to each netwell. The control bin
230 remained at 29 °C and the treatment bin was heated up to 32 °C over 24 hrs (Fig. S1a). After 4
231 days of exposure, swimming larvae were removed from each netwell, flash frozen in liquid
232 nitrogen, and stored at -80 °C for RNA extraction.

233

234 Similar acute and moderate stress experiments were conducted with larvae reared in 2021 (HR
235 x HR, CR x HR, HR x CR). All cultures were transported to the Experimental Reef Lab (ERL) at
236 The University of Miami's Cooperative Institute for Marine and Atmospheric Studies (CIMAS) on
237 day 4 post fertilization. Larvae were packed into 50 mL centrifuge tubes with no air bubbles and
238 stored in coolers at ambient temperature during transit. Upon arrival at ERL, larvae were re-
239 distributed into 6 L culture bins (n = 3 per cross) filled with 0.2 µm FSW. Temperature control
240 was accomplished using the ERL aquaria with individual treatment tanks serving as water baths
241 (Enochs et al., 2018). For the stress experiments, netwells were floated directly in the
242 temperature-controlled flow-through aquaria. For each bulk cross, two groups of 10 larvae were
243 allocated to each treatment tank in the acute stress treatment (n = 6 per cross type per
244 treatment) and three groups of 20 larvae were allocated to each treatment tank in the moderate
245 stress (n = 9 per cross type per treatment). The control temperature for both experiments was
246 set to 27 °C. Heat ramps started 5 days post fertilization (note this represented different
247 calendar days for HR x HR vs. CR x HR and HR x CR to account for differences in
248 developmental age) and target temperatures were reached over 48 hrs (Fig. S1b). Lights were
249 maintained at 180 µmol s⁻¹·m⁻² for a 12:12 hr light dark cycle. The acute stress assay was
250 monitored every 12 hrs for survival and the experiment was terminated once mortality reached
251 more than 50% for the majority of the netwells. For the moderate duration experiment, larvae

252 were retrieved from 1 netwell per cross per replicate tank following 4 days of exposure using a
253 pipette, counted, and transferred to a cryovial for RNA extraction. Excess seawater was
254 removed and larvae were snap frozen in liquid nitrogen and stored at -80 °C until processing.
255 Additional replicate samples were taken from the remaining 2 netwells for protein and lipid
256 analyses.

257

258 Physiological assays

259 Gametes from both years, as well as 2021 fertilized larvae, were collected for lipid analyses.

260 Gametes from each parent colony were sampled in duplicate. Gametes and larvae were
261 counted under a dissection microscope before being transferred to combusted glass tubes and
262 frozen at -20 °C until lipid processing. Total lipids were extracted and determined gravimetrically
263 using a modified Folch method (Folch et al., 1957), as described in (Keister et al., 2023).

264 Subsequently, 100% chloroform was added to total lipids to achieve a 10 mg mL⁻¹
265 concentration, to standardize samples. Lipid classes were quantified by spotting, in duplicate, 1
266 µL of extracted lipids on silica Chromarods[®] before being developed using thin-layer
267 chromatography via a two-step solvent system (Conlan et al., 2014, 2017; Nichols et al., 2001),
268 as described in (Keister et al., 2023). Developed rods were then dried at 100 °C for 10 min in an
269 IsoTemp Oven (Fisher Scientific) before being run on an Iatroscan MK 6S thin-layer flame
270 ionization detector (TLC-FID). Known concentrations of lipid compounds, ranging from 0.1–10.0
271 mg mL⁻¹, were used to calibrate the Iatroscan for the following lipid classes:

272 phosphatidylethanolamine (PE), phosphatidylserine and phosphatidylinositol (PS-PI),
273 phosphatidylcholine (PC), lysophosphatidylcholine (LPC), wax esters (WAX), triacylglycerols
274 (TAG), sterols (ST), and diacylglycerols (DAG). All phospholipid lipid classes (PE, PS-PI, PC,
275 LPC) were grouped and analyzed as one unit. All total lipid and lipid class values were
276 presented as µg per gamete or µg per larva, respectively.

277

278 Protein samples were thawed on ice and homogenized by back pipetting in FSW for at least 60
279 s until no visible cellular debris was present. Total homogenate volume was recorded. Soluble
280 host protein was quantified in triplicate with a BCA Protein Assay Kit II (BioVision) following the
281 manufacturer's protocol. Final protein concentration was multiplied by the initial homogenate
282 volume and then standardized by the number of total sampled larvae.

283

284 Statistical analysis of physiological trait data and survival under acute stress

285 All statistical analyses were conducted in R 4.2.1 (R Core Team, 2022). Traits were evaluated
286 for normality using the Shapiro-Wilks test and log-transformed if not normally distributed. A two-
287 way anova was performed to analyze the effects of treatment (levels: control and heat) and
288 origin (levels: CR x CR, HR x HR) on 2019 gamete lipid content. No statistical analysis of 2021
289 gametes was possible given that only one CR colony spawned. Larval protein and lipid content
290 for the 2021 cohorts were analyzed using a two-way anova to analyze the effect of origin
291 (levels: HR x HR, CR x HR, HR x CR) and treatment (levels: control and heat). A mixed effects
292 Cox model (Therneau, 2018) was used to model time of death for the acute stress experiment
293 from both years as a function of treatment and origin, including a random effect of replicate bulk
294 cross. Due to low mortality rate, HR x HR and control were chosen as the reference level for
295 each fixed effect to calculate hazard ratio for other levels. A hazard ratio above 1 means a
296 higher mortality risk and below 1 means a lower mortality risk compared to the reference levels.
297 The null hypothesis was rejected at an alpha of 0.05.

298

299 RNA extraction, library preparation, and sequencing

300 Total RNA was extracted from frozen samples using the Aurum Total RNA Mini Kit (Bio-Rad).
301 Samples were back pipetted to ensure complete homogenization after Lysis Solution was
302 added. Genomic DNA was removed by adding DNase I on-column according to the
303 manufacturer's instructions. RNA concentration was quantified using a Take2 plate on a

304 Synergy H1 microplate reader (Biotek) and only samples with > 10 ng/μl concentration were
305 used to generate tag-based RNA-seq libraries, following protocols modified for sequencing on
306 the Illumina platform https://github.com/ckenkel/tag-based_RNAseq.

307
308 Libraries from 2019 (n = 45 samples) were sequenced on the NextSeq 550 in 2020 by the USC
309 Genome Core using a 1x75bp HO kit. Libraries from 2021 (n = 30 samples) were sequenced on
310 the NextSeq 2000 in 2022 in two replicate separate runs by the USC Norris Comprehensive
311 Cancer Center Molecular Genomics core using a NextSeq 2000 P2 Reagents v3 kit, after which
312 reads were concatenated to reach a comparable sequencing depth. The average sequencing
313 depth per sample for the two libraries was 5.9 M reads (± 0.2 M) and 5.4 M reads (± 0.3 M) for
314 the 2019 and 2021 datasets, respectively.

315

316 Bioinformatics pipelines

317 Downstream bioinformatic processing was performed on USC's Center for Advanced Research
318 Computing (CARC) following protocols described in [https://github.com/ckenkel/tag-](https://github.com/ckenkel/tag-based_RNAseq)
319 [based_RNAseq](https://github.com/ckenkel/tag-based_RNAseq). Briefly, a custom perl script was used to discard PCR duplicates and reads
320 missing adapter sequences. Poly-A tails and adapter sequences were trimmed and only high
321 quality reads (PHRED score ≥ 20 over 90% of the read) were retained. A total of 2.7 M (± 0.09
322 M) and 2.4 M (± 0.1 M) reads per sample remained for the 2019 and 2021 datasets respectively
323 after quality filtering. Filtered reads were mapped to an adult *Orbicella faveolata* transcriptome
324 (Supplemental Materials) using the gmapper command from SHRiMP2 (Rumble et al., 2009).
325 Read counts were summed by isogroup using a custom perl script, resulting in a per sample
326 mapped read average of 1.5 M (± 0.05 M) and 1.3 M (± 0.07 M), for the 2019 and 2021 datasets
327 respectively.

328

329 Statistical analysis of gene expression

330 Larval samples from the two years were first compiled, where isogroups shared between the
331 two year cohorts that had fewer than 2 counts across 90% of the samples were discarded,
332 resulting in a total of 28,675 high quality isogroups remaining. Counts were then transformed
333 using the *rlog* function in DESeq2 (Love et al., 2014) and outlier samples were identified through
334 a sample network with a standardized connectivity score of less than -2.5 , based on which one
335 sample from the 2019 dataset and one from the 2021 data set were filtered. The filtered dataset
336 was then separated by year to form two subsets, on which principal component analysis (PCA)
337 was applied to visualize global gene expression pattern for each larval cohort using the R
338 package FactoMineR (Lê et al., 2008).

339

340 Gene co-expression analysis was conducted using the WGCNA package in R (Langfelder &
341 Horvath, 2008) following the standard tutorial (Langfelder & Horvath 2016) and a pipeline for
342 performing meta-analysis of two microarray datasets (Miller 2011). Briefly, a signed gene co-
343 expression network was constructed with a soft power of 4. Module assignment was generated
344 based on the 2019 dataset and imposed onto the 2021 dataset to assess how well those
345 modules were preserved across datasets. The eigengenes of these shared modules were
346 correlated with origin (levels for 2019: CR x CR, HR x HR, Hybrid; levels for 2021: HR x HR, CR
347 x HR, HR x CR) and treatment (levels: control, heat).

348

349 DESeq2 (Love et al., 2014) was used to examine genes that were differentially expressed in
350 larvae by origin and treatment, with each year's cohort being analyzed separately. Isogroups
351 that had fewer than 2 counts across 90% of the samples were discarded, resulting in a total of
352 27,894 high quality reads in the 2019 dataset and 28,960 reads in the 2021 dataset. A two-
353 factor grouping was created by combining origin and treatment and fed into the DESeq2 model
354 as the design, from which results were extracted with specific contrasts to obtain differentially
355 expressed genes (DEGs) in response to treatment in a particular larval origin (e.g. CR x CR

356 heat vs. CR x CR control). Significance testing was determined using a Wald test after
357 independent filtering using a false discovery rate-corrected (FDR) threshold of 0.1. Multiple test
358 correction was applied to raw p-values following Benjamini & Hochberg (Benjamini & Hochberg,
359 1995) and adjusted p-values less than 0.1 were considered significant. Signed log-p values
360 were generated based on the adjusted p-values to serve as input for gene ontology (GO)
361 enrichment analysis described below.

362

363 To explore the possibility that heat-responsive genes were front- or back-loaded in the CR x CR
364 or HR x HR populations which would preclude their identification as significantly differentially
365 expressed genes (Barshis et al., 2013), we identified the overlap between significantly up-
366 regulated/down-regulated genes in response to heat in one population and the significantly up-
367 regulated/down-regulated genes between groups in control conditions (i.e. constitutively
368 differentially expressed) as candidate front-/back-loaded genes in the other population. For
369 example, genes identified as significantly up-regulated in CR x CR larvae but not HR x HR
370 larvae under heat that were also significantly up-regulated under control in HR x HR vs CR x CR
371 larvae were identified as potentially front-loaded genes in HR x HR larvae. The same approach
372 was repeated for determining back-loaded genes as well as front-loaded/back-loaded genes in
373 CR x CR larvae.

374

375 To understand the functional implications of conserved gene modules identified in WGCNA, a
376 categorical gene ontology (GO) enrichment analysis was performed using binary values (1 or 0)
377 to indicate module membership in the WGCNA set followed by a Fisher's exact test and false
378 discovery rate correction. For heat-responsive DEGs by larval origin, signed log p-values using
379 adaptive clustering of GO categories and a two-sided Mann-Whitney U-test was applied,
380 followed by a false discovery rate correction. GO scripts can be found at

381 https://github.com/z0on/GO_MWU. Heatmaps of hierarchically clustered GO terms were
382 generated using the pheatmap package (Kolde, 2012) in R.
383
384 Discriminant analysis of principal components (DAPC) was performed to explore the relative
385 changes in global expression between treatment in different populations from 2019 and 2021
386 using the R package adegenet (Jombart, 2008). Variance stabilized data (VSD) were used to
387 create the model and the number of PCs was chosen to capture at least 80% of transcriptional
388 variance. Distribution of samples grouped by origin and treatment was visualized in density
389 plots.

390

391 **Results**

392

393 Temperatures and bleaching history at study sites

394 SST was always cooler at HR than CR (Fig. 1) during the data collection period, although
395 bleaching occurred at both sites in 2014 and 2015 (Gintert et al., 2018; Manzello et al., 2019).
396 Bleaching was observed at CR in 2018 and during gamete collection in 2019 at CR, but not at
397 HR. CR gametes were collected from colonies that didn't show visual bleaching. In 2021,
398 bleaching was not observed at either site, in line with the cooler temperature patterns (Fig. 1). In
399 total, from the start of the most recent mass global bleaching event in 2014 to the 2021 spawn,
400 CR experienced four bleaching events, whereas HR only experienced two events. Notably, HR
401 did not experience bleaching between 2015 and 2021.

402

403 Larval survival under acute heat stress

404 To achieve a reasonable separation of survivorship among the different groups, the 36 °C acute
405 temperature stress lasted for 96 hrs in 2019 and 141 hrs in 2021. The average survival rate at
406 the end of each year's experiment for the control vs. heat group was 77% vs. 47% and 86% vs.

407 33% respectively (Fig. 2). Exposure to 36 °C significantly increased larval mortality in both
408 years, resulting in hazard ratios of 3.1 ($z = 4.16$, $p < 0.001$) and 28.7 ($z = 4.11$, $p < 0.001$) for
409 heat-treated larvae from 2019 and 2021, respectively. Larval origin also had a significant effect
410 on survival in both years. In 2019, the CR x CR cross experienced almost double the mortality
411 risk (HR = 1.9, $z = 2.18$, $p < 0.05$) in comparison to the HR x HR cross, and in 2021 the CR x
412 HR cross experienced more than 10 times the risk (HR = 11.8, $z = 2.18$, $p < 0.001$) of the HR x
413 HR cross.

414

415 Gamete (2019 and 2021) and larval ecophysiology (2021 only)

416 Total lipid content of gametes collected from CR and HR during 2019 spawning season did not
417 differ between the two sites (Fig. 3b). However, HR gametes contained 2.2 times more wax
418 esters ($F = 6.27$, $df = 1$, $p < 0.05$) and 1.5 times more phospholipids ($F = 6.07$, $df = 1$, $p < 0.05$)
419 than CR gametes (Fig. 3a). No differences in triacylglycerol content by origin were apparent.
420 Qualitatively, lipid content tended to be higher in the CR gametes in 2021 (Fig. 3b), but as only
421 one colony spawned, formal significance was not evaluated. No differences were detected in
422 total lipid content, lipid classes, or total soluble protein content of 2021 larvae between origin,
423 treatment conditions or their interaction (Fig. S2, S3; Table S2). Protein and lipid data were
424 unavailable for 2019 larvae exposed to the 32 °C moderate stress.

425

426 Major drivers of transcriptional variation

427 Principal component analysis (PCA) on all high expression genes (count > 2 in 90% samples)
428 within each dataset showed that larval origin was an important driver of transcriptional variance
429 in addition to temperature treatment (Fig. 4). Samples clustered by larval origin along PC2 in the
430 2019 dataset, which accounted for 5.6% of the overall variance in expression (Fig. 4a). Putative
431 hybrid samples generally clustered mid-way between HR x HR and CR x CR origin larvae. In
432 the 2021 dataset, larval origin was the primary driver of expression variation, as cross types

433 were clustered along the first PC which accounted for 10.4% of the overall variance (Fig. 4b).
434 No significant differences in clustering were apparent between the reciprocal hybrids (CR x HR
435 vs. HR x CR) but both were distinct from the HR x HR origin larvae. Clustering by temperature
436 treatment was apparent along PC2 in the 2021 dataset, which explained 9.6% of the variance in
437 overall expression (Fig. 4b), and along PC3 in the 2019 dataset, which explained 5.1% of
438 transcriptional variance (Fig. S4).

439

440 Conservation of expression networks and their functional significance

441 WGCNA was used to investigate whether and to what degree expression patterns were
442 conserved between the two datasets and to further explore their relationships with experimental
443 factors. Four modules, pink (n = 400 genes), purple (n = 268), magenta (n = 283), and red (n =
444 514), were highly correlated with origin in both years, although the directions of the correlations
445 were not always conserved (Fig. 5). Specifically, genes within the pink and purple modules were
446 strongly negatively correlated, or down-regulated, in 2019 CR x CR larvae (pink: Pearson's $r = -$
447 0.87 , $p_{cor} = 5e^{-15}$; purple: Pearson's $r = -0.74$, $p_{cor} = 8e^{-9}$) and up-regulated in 2019 HR x HR
448 larvae (pink: Pearson's $r = 0.81$, $p_{cor} = 1e^{-11}$; purple: Pearson's $r = 0.69$, $p_{cor} = 1e^{-7}$), while the
449 opposite relationship was observed in the magenta (CR x CR: Pearson's $r = 0.71$, $p_{cor} = 6e^{-8}$;
450 HR x HR: Pearson's $r = -0.86$, $p_{cor} = 7e^{-14}$) and red modules (CR x CR: Pearson's $r = 0.88$, $p_{cor} =$
451 $2e^{-15}$; HR x HR: Pearson's $r = -0.76$, $p_{cor} = 1e^{-9}$). In comparison, the magnitude and direction of
452 module-trait correlations remained similar for the pink and red modules in 2021, with pink
453 module genes again showing strong up-regulation in HR x HR larvae and red module genes
454 showing strong down-regulation (pink: Pearson's $r = 0.98$, $p_{cor} = 3e^{-21}$; red: Pearson's $r = -0.98$,
455 $p_{cor} = 3e^{-21}$). While there were still highly significant correlations observed for the purple and
456 magenta modules with respect to origin, the direction of the association was completely
457 reversed, with strong down-regulation of genes in the purple module (Pearson's $r = -0.98$, $p_{cor} =$
458 $9e^{-20}$) and up-regulation of genes in the magenta module (Pearson's $r = 0.96$, $p_{cor} = 5e^{-17}$).

459 Modules significantly associated with treatment exhibited more moderate correlation coefficients
460 (Pearson's r range: ± 0.4 to ± 0.6), but their expression patterns were more strongly conserved
461 across years. Genes in the black ($n = 408$ genes), greenyellow ($n = 268$), salmon ($n = 160$), and
462 yellow ($n = 574$) modules were consistently up-regulated in heat-treated larvae whereas genes
463 in the cyan ($n = 152$) module were consistently down-regulated in heat-treated larvae across
464 years.

465

466 Categorical functional enrichment analysis of genes assigned to the significant origin modules
467 (pink, purple, red, magenta) revealed few significant GO terms, likely due to small module size.

468 Two terms in the molecular function category (GO:0031683, G-protein beta/gamma-subunit
469 complex binding; GO:0050780, dopamine receptor binding) and one term in the cellular
470 components category (GO:0005834; GO:1905360, GTPase complex) were significantly
471 enriched ($p \leq 0.05$) among genes in the pink module (Table S3). GO analysis of treatment
472 modules identified enrichment of genes associated with heat response pathways, such as
473 immune response, defense response, and response to external stimulus in the black module
474 which was up-regulated in response to heat treatment ($p < 0.01$, Fig. S5, Table S3). Whereas
475 the cyan module, which was down-regulated in response to heat, showed enrichment of genes
476 associated with amide and peptide metabolic and biosynthetic processes ($p < 0.05$, Fig. S6,
477 Table S3). These processes were also enriched among genes in the salmon module, which was
478 up-regulated under heat (Table S3). One term in the molecular function category (GO:0031210;
479 GO:0050997, phosphatidylcholine binding) was significantly enriched ($p = 0.05$) among genes in
480 the yellow module, and was also up-regulated under heat treatment (Table S3). No significant
481 functional enrichments were detected for the greenyellow module.

482

483 Origin-specific responses to thermal stress

484 DESeq2 analysis of the 2019 dataset showed that a total of 561 genes were up-regulated and
485 436 were down-regulated in heat-treated larvae relative to controls (Table S4). When further
486 partitioned by origin, 133 heat-responsive genes were up-regulated and 144 were down-
487 regulated in 2019 CR x CR larvae, whereas 376 were up-regulated and 158 were down-
488 regulated in 2019 HR x HR larvae. The putative hybrid larvae up-regulated 12 genes and down-
489 regulated 9 genes under heat and thus were excluded from the downstream functional
490 enrichment analysis due to a limited number of differentially expressed genes (DEGs). Among
491 these origin-specific heat-responsive genes, 28 (11 annotated) were front-loaded and 47 (31)
492 were back-loaded in 2019 HR x HR larvae, while 82 (48) were front-loaded and 23 (14) were
493 back-loaded in 2019 CR x CR larvae (Table S5). In the 2021 dataset, 706 heat-responsive
494 genes were up-regulated and 953 were down-regulated in HR x HR larvae, 228 were up-
495 regulated and 54 were down-regulated in CR x HR larvae, and 124 were up-regulated and 101
496 were down-regulated in HR x CR larvae.

497

498 Subsequent ontology analysis of these DEGs showed that biological processes including
499 immune response, peptide hormone processing, defense response, and response to stimulus
500 were enriched among heat-responsive up-regulated genes in 2019 CR x CR larvae (FDR <
501 0.01), while RNA metabolic process and macromolecule biosynthetic process were enriched
502 among down-regulated genes (FDR < 0.01, Fig. 6a, Table S6). In the 2019 HR x HR larvae,
503 phospholipid catabolic process was enriched among genes up-regulated in heat (FDR < 0.05),
504 while microtubule-based movement/process and protein-DNA complex subunit organization
505 were enriched among downregulated genes (FDR < 0.01, Fig. 6b, Table S7). Discriminant
506 analysis of principal components (DAPC) for heat responsive genes revealed a greater
507 transcriptional response in 2019 HR x HR larvae compared to the CR x CR larvae from the
508 same year (Fig. 6c).

509

510 In the 2021 cohort, genes up-regulated in response to heat in HR x HR larvae were enriched for
511 biological processes including DNA metabolic process, organelle localization, and cellular
512 response to DNA damage stimulus (FDR < 0.01, Fig. S8, Table S8). Processes enriched among
513 down-regulated genes included a suite of metabolic processes of small and large molecules,
514 such as organic acids, fatty acids, and lipids (FDR < 0.01). For the two hybrid crosses,
515 upregulated genes were enriched for NF- κ B signaling regulation, immune and defense
516 response, as well as regulation of cell death (FDR < 0.05, Fig. S9, S10, Table S9, S10).
517 Downregulated genes were enriched for lipids and protein metabolic/catabolic processes (FDR
518 < 0.05). DAPC for heat responsive genes identified a greater response in HR x CR larvae
519 compared to the HR x HR and CR x HR larvae from the same year (Fig. S11).

520

521 **Discussion**

522 Adult *O. faveolata* from Cheeca Rocks exhibit elevated thermal tolerance in response to natural
523 bleaching events (Gintert et al., 2018; Manzello et al., 2019) suggesting that they have
524 acclimatized or adapted to naturally higher and more variable temperatures characteristic of
525 inshore reef sites in the Florida Keys (Kenkel et al., 2015; Kenkel & Matz, 2016). Contrary to the
526 current paradigm of inherited and/or enhanced thermal tolerance in adults experiencing more
527 extreme thermal regimes (Dixon et al., 2015; Putnam & Gates, 2015; Strader & Quigley, 2022),
528 we found that offspring of these more tolerant inshore colonies were more susceptible to
529 thermal stress, exhibiting reduced survival and stronger expression signatures of a stress
530 response in comparison to larvae from offshore colonies (Fig. 2, 6). The observed total lipid data
531 suggest robust bleaching resistance in adults may come at the cost of reproductive investment
532 (Fig. 3), although patterns are inconsistent across years. These findings represent an important
533 counter-example to the rationale underpinning selective breeding approaches (Drury et al.,
534 2022): that tolerant parents can be counted on to produce tolerant offspring.

535

536 Impaired larval performance may result from reduced reproductive investment

537 Reef origin (or cross type) and temperature treatment played important roles in driving
538 physiological traits in *O. faveolata* larvae. As expected, larvae were more likely to die under heat
539 treatment than in the ambient control (Fig. 2), but the origin response was unexpected.
540 Horseshoe Reef (HR) purebred larvae appeared to be the best performers in both years, while
541 Cheeca Rocks (CR) purebred larvae in 2019 and 2021 CR x HR hybrids experienced
542 significantly higher mortality in comparison (Fig. 2). Previous studies conducted on *Acropora*
543 *millepora* and *Acropora tenuis* from the Great Barrier Reef showed that larval survival under a
544 similar level of acute thermal stress (35.5 °C) was enhanced in offspring of parents from warmer
545 source reefs (Dixon et al., 2015; Strader & Quigley, 2022). Yet, in our study system, the HR x
546 HR cross, produced by the less tolerant parents sourced from a cooler reef environment,
547 repeatedly outperformed the CR cross produced by more tolerant parents from a warmer reef
548 environment (Gintert et al., 2018; Manzello et al., 2019). This suggests that the host genetic
549 contribution to thermal tolerance may be minimal or overpowered by other factors such as
550 recent and concurrent heat stress, or maternal provisioning.

551
552 Prior thermal stress and bleaching has been linked to negative reproductive outcomes, including
553 both fecundity and gamete quality, in multiple coral species (Jones & Berkelmans, 2011;
554 Szmant & Gassman, 1990; Ward et al., 2002). Although the severity of reproductive impacts is
555 thought to be related to the severity of bleaching and rate of recovery, resistant/resilient colonies
556 do not necessarily exhibit latent effects (Leinbach et al., 2021; Szmant, 1991). Notably, the HR
557 source population suffered from both a reduced number of spawning colonies and total gametes
558 released as a result of the back-to-back bleaching events in 2014 and 2015 (Fisch et al., 2019).
559 Previous research consistently demonstrates that adult colonies from nearshore reef
560 environments display higher resistance and resilience to thermal stress compared to those from
561 offshore reefs (Gintert et al., 2018; Manzello et al., 2019). Based on these findings, we

562 anticipated that the offspring of these resilient adults would also exhibit enhanced thermal
563 tolerance, in line with previous studies. At the time of spawning collections in 2019, temperature
564 at CR had surpassed the local bleaching threshold and a majority of the colonies showed some
565 degree of paling (i.e., onset of bleaching), while HR experienced cooler temperatures and all the
566 colonies appeared fully pigmented (Fig. 1). Recently, (Leinbach et al., 2021) showed that
567 *Acropora hyacinthus* which resisted bleaching maintained higher reproductive capacity than
568 recovered coral. Although we only collected gametes from CR colonies without apparent signs
569 of physiological stress in 2019 (i.e. thermally resistant adults), it is likely that the energetic state
570 of individual colonies was already compromised due to the heat stress, which may have
571 impacted maternal provisioning.

572

573 Scleractinian coral gametes are largely composed of lipids with wax esters, phospholipids and
574 triacylglycerols being the most abundant classes (Figueiredo et al., 2012). Mean total lipids
575 were higher for 2019 HR gametes compared to CR gametes (Fig. 3b). Lipid class analyses, of
576 the 2019 gametes, showed CR gametes contained less wax esters and phospholipids than HR
577 gametes (Fig. 3a). Wax esters take longer to metabolize, supporting their role in long-term
578 energy storage, and are important for larval development (Lee et al., 2006; Richmond, 1987;
579 Rivest et al., 2017), indicating 2019 HR gametes had greater energy reserves available for
580 dispersal and settlement. Additionally, greater wax ester stores lower gamete/larval density,
581 possibly contributing to extended dispersal (Richmond, 1987). Phospholipids were significantly
582 different across sites in 2019, but this lipid class encompasses many structural lipid compounds,
583 thus muddling the implications. While triacylglycerols can be rapidly hydrolyzed and likely
584 support immediate energetic needs (Figueiredo et al., 2012; Sewell, 2005), no significant
585 differences were detected across sites in either 2019 or 2021. Similar lipid class patterns were
586 not observed in the 2021 cohort, possibly due to lack of replication in CR gametes (Fig. 3a). The
587 general deficiency of 2019 CR gametes in all these classes suggests that bleaching tolerance

588 and resilience of CR adults may come at the cost of reproductive investment, potentially
589 contributing to reduced performance of CR larvae.

590

591 Further supporting a compromised reproductive output of nearshore coral populations was the
592 absence of mass spawning on CR in the summer of 2021, despite no observed bleaching (Fig.
593 1), which limited our ability to re-assess performance of CR x CR larvae and additional hybrid
594 crosses. This implies that reproductive capacity of CR adults (likely among other marginal
595 nearshore populations) could be jeopardized by the persistent latent effects of accumulated
596 stress. As thermal anomalies increase in magnitude and frequency in the Florida Reef Tract
597 (Manzello, 2015), it is important to consider any latent effects beyond visible stress responses
598 that could affect the next generation and ultimately the persistence of reef communities.

599

600 Stress tolerance rather than a front-loaded stress response is associated with enhanced larval
601 survival

602 In addition to origin-specific physiological differences, larvae originating from different reef sites
603 mounted different transcriptional responses after four days of 32 °C heat challenge. In the 2019
604 CR purebred larvae, we observed functional enrichment in a wide range of stress responses
605 (e.g., immunity, defense, inflammation) among the significantly upregulated genes, whereas
606 metabolic and biosynthetic processes (e.g., various RNA molecules, peptide, cellular nitrogen
607 compounds) were enriched among downregulated genes (Fig. 6a). Similarly, underperforming
608 CR x HR larvae in 2021 showed pronounced upregulation of defense and immune response
609 pathways and downregulation of metabolic processes (Fig. S9). Upregulation of stress response
610 genes and concomitant downregulation of growth related processes, such as rRNA metabolism,
611 is a hallmark of the environmental stress response (López-Maury et al., 2008). In contrast, fewer
612 enriched GO categories were identified among the heat-responsive DEGs in the 2019 and 2021
613 HR purebred crosses (Fig. 6b, S7), and more interestingly, the enrichments did not highlight

614 stress response pathways like their CR x CR (2019) or hybrid (2021) counterparts, but cellular
615 homeostatic processes instead (Fig. 6b).

616

617 The lack of an apparent stress response in HR purebreds does not appear to be due to an
618 inability to detect differential expression as a result of transcriptional front-loading, or higher
619 baselines expression of stress response genes (Barshis et al., 2013) We tested for front-loading
620 and found that comparatively fewer genes were front-loaded in HR x HR larvae and none of
621 those were annotated as stress response genes (Table S5). Among the annotated genes that
622 were front-loaded in CR x CR larvae, two were related to protein ubiquitination (Ube2g1 and
623 ZNF598, Table S4), which may indicate constitutive expression of stress response pathways.
624 Additionally, HR purebreds had a more robust expression response to thermal stress, exhibiting
625 lower baseline expression, but achieving the same magnitude of overall transcriptional plasticity
626 in response to thermal stress as 2019 CR x CR larvae (Fig. 6c). In 2021 HR x HR and CR x HR
627 larvae showed similarly elevated levels of baseline expression in comparison to HR x CR hybrid
628 (Fig. S11). Taken together, this suggests the 2019 HR x HR larvae may be more resistant to
629 thermal stress not because they were pre-conditioned for stressful conditions, but because they
630 were able to strongly and rapidly acclimate their physiology, possibly as a result of having more
631 energy reserves to devote towards a stress response (Fig. 3). Such a robust response may also
632 be followed by a rapid return to baseline expression, or transcriptomic resilience (Rivera et al.,
633 2021), when stress abates, although additional time-course data are needed to test this
634 hypothesis.

635

636 Conserved transcriptomic signatures of population origin and response to treatment

637 In addition to survival and gamete lipid content, global gene expression profiles of aposymbiotic
638 *O. faveolata* larvae also revealed a strong signature of reef origin. The 2019 samples were
639 organized into 3 distinct clusters based on origin along PC2, while origin was the predominant

640 driver of clustering in the 2021 samples (Fig. 4). Treatment appeared to be a weaker driver in
641 both datasets, clustering samples along PC3 in 2019 and PC2 in 2021 (Fig. S4, 4b). Despite a
642 difference in developmental age of the two larval cohorts at the time of sampling (7 vs. 9 days
643 post fertilization for 2019 vs. 2021), a WGCNA meta-analysis identified highly conserved gene
644 modules significantly correlated with larval origin although the magnitude and direction of
645 expression patterns was not always conserved (Fig. 5). Corals exhibit waves of transcription
646 during early development consistent with zygotic genome activation and degradation of
647 maternal transcripts (Chille et al., 2021; Hayward et al., 2011). Differences in the magnitude and
648 direction of select modules between the 2019 and 2021 datasets (purple and magenta, Fig. 5)
649 may reflect these temporal transcriptional waves, although a more thorough time course is
650 needed to test this hypothesis. Nevertheless, evidence of such strong module preservation
651 implies the existence of a core group of origin-specific genes that have a lasting effect
652 throughout the organisms' development and these modules may be linked to baseline
653 differences in larval physiology. However, little information on the functional implications of
654 these gene sets was retrieved from the GO enrichment analysis, which may be attributable to
655 small module size and/or an insufficient number of annotated genes. This could be a worthwhile
656 investigation for future studies as annotations improve and better enrichment methods become
657 available.

658
659 WGCNA meta-analysis also identified highly conserved gene modules that were significantly
660 associated with temperature treatment. Similar to the origin-specific modules, the majority of
661 these treatment modules lacked significant functional enrichment. For the two modules that did
662 have sufficient enrichment, biological processes including immune response, defense response,
663 and response to external stimulus were enriched in genes associated with the black module,
664 which were up-regulated in heat (Fig. S5), while amide and peptide metabolic and biosynthetic
665 processes were enriched in genes associated with the cyan module that were downregulated in

666 heat (Fig. S6). Therefore, upregulation of genes involved in stress response pathways and
667 downregulation of metabolic processes related to growth and development in heat-treated
668 larvae aligns with the cellular stress and cellular homeostasis response profiles identified in
669 corals (and other marine and terrestrial organisms) during short- to medium-term physiological
670 stress (Kenkel et al., 2014; Kültz, 2005).

671

672 Implications for adaptive management

673 The documented resilience of adult corals to recurrent heat stress and bleaching in
674 environments with high and variable temperatures may come with highly consequential trade-
675 offs. In this study, we show that larvae from a site that routinely experiences and recovers from
676 heat-induced bleaching, significantly underperformed relative to larvae from a cooler, less
677 variable reef site. The unexpected outcome that more thermally tolerant *O. faveolata* produced
678 poorer quality offspring challenges the prevailing paradigm that breeding vulnerable populations
679 with thermally tolerant individuals can contribute to genetic rescue (Bay et al., 2017), which also
680 serves as the theoretical justification for selective breeding approaches (Drury et al., 2022).
681 Moreover, natural adaptive capacity is likely already impaired as CR x CR larvae and CR x HR
682 hybrids exhibited greater mortality risk even under ambient conditions (Fig. 2). These findings
683 also align with the long-term pattern of recruitment failure in Caribbean coral (Hughes & Tanner,
684 2000; Williams et al., 2008)

685

686 **Data availability**

687 Raw sequence data can be found under NCBI BioProject PRJNA981197. Scripts and input files
688 are available at DOI: 10.5281/zenodo.8025715.

689

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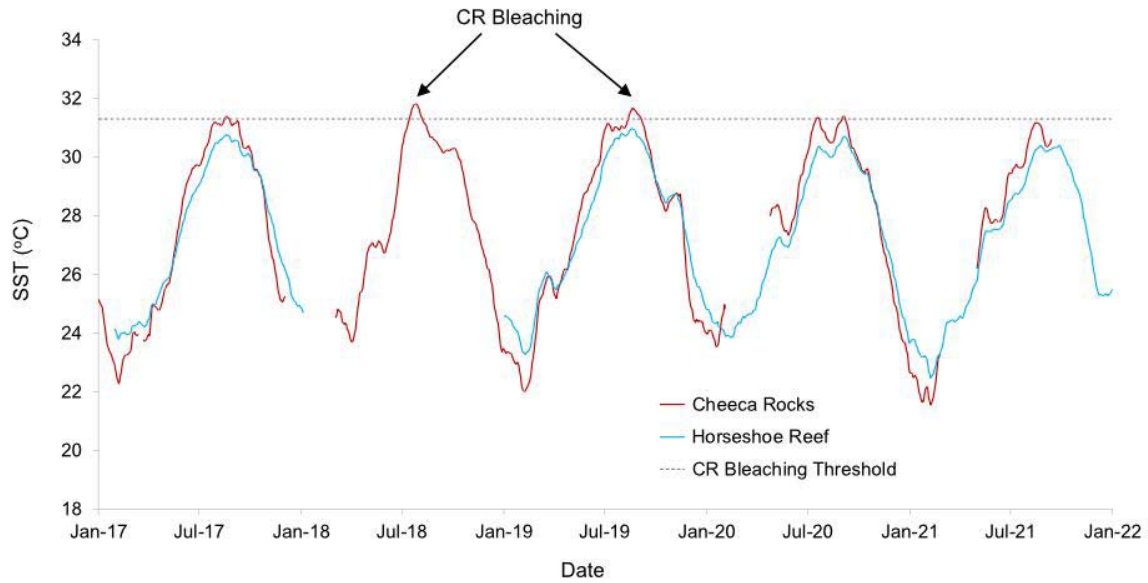


Fig. 1 Time series of running 30-day mean sea surface temperature (SST) for Cheeca Rocks (red line) and Horseshoe Reef (blue line) from 2017 to 2021. Bleaching threshold for Cheeca Rocks (monthly mean SST $\geq 31.3^{\circ}\text{C}$) shown as dashed black line.

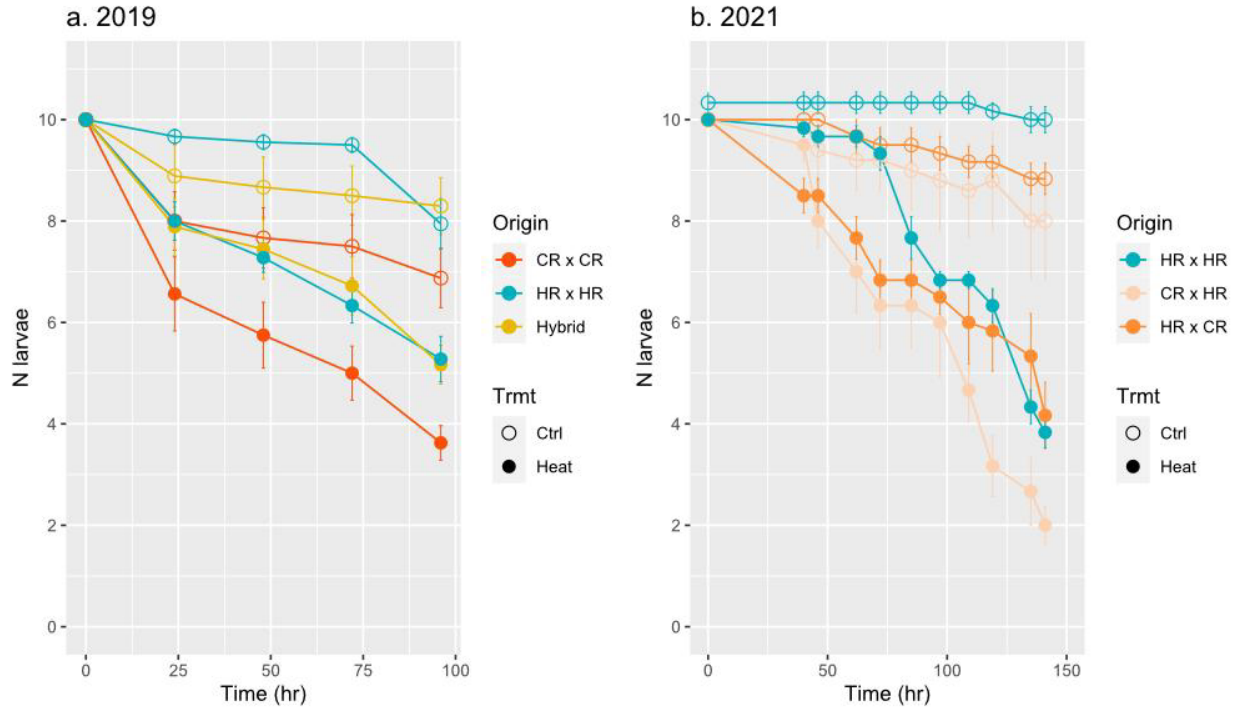


Fig. 2 Number of surviving larvae (mean \pm standard error of the mean [SEM]) across time in the acute temperature stress experiment for a) 2019 and b) 2021 cohorts. Survivorship was grouped by larval origin and treatment condition.

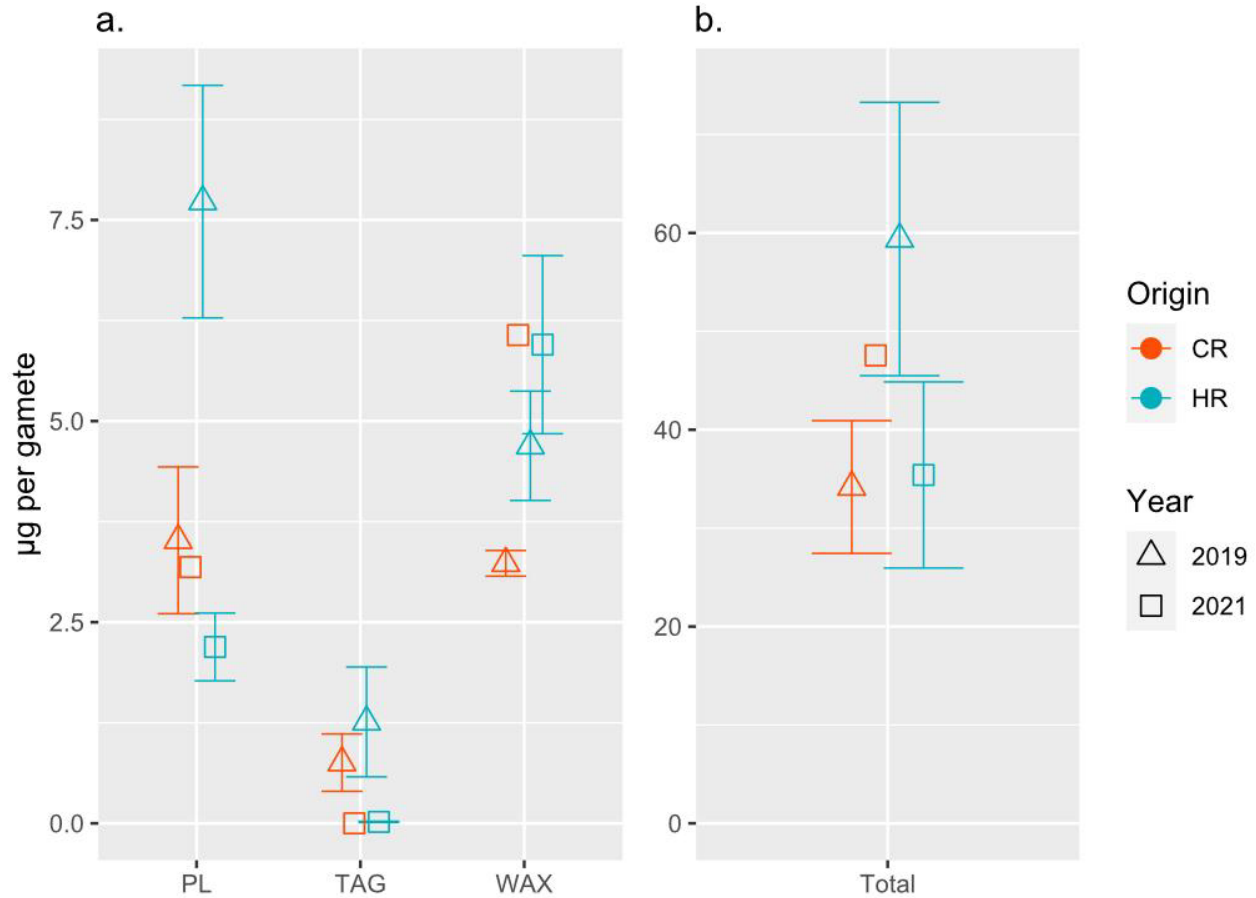


Fig. 3 Concentration (mean \pm SEM) of (a) different lipid classes (PL: phospholipid, TAG: triacylglycerol, WAX: wax ester) and (b) total lipids standardized by individual gamete bundle collected from Cheeca Rocks (n = 5) and Horseshoe Reef (n = 5) in 2019 and 2021. No replicates were available for CR gametes in 2021.

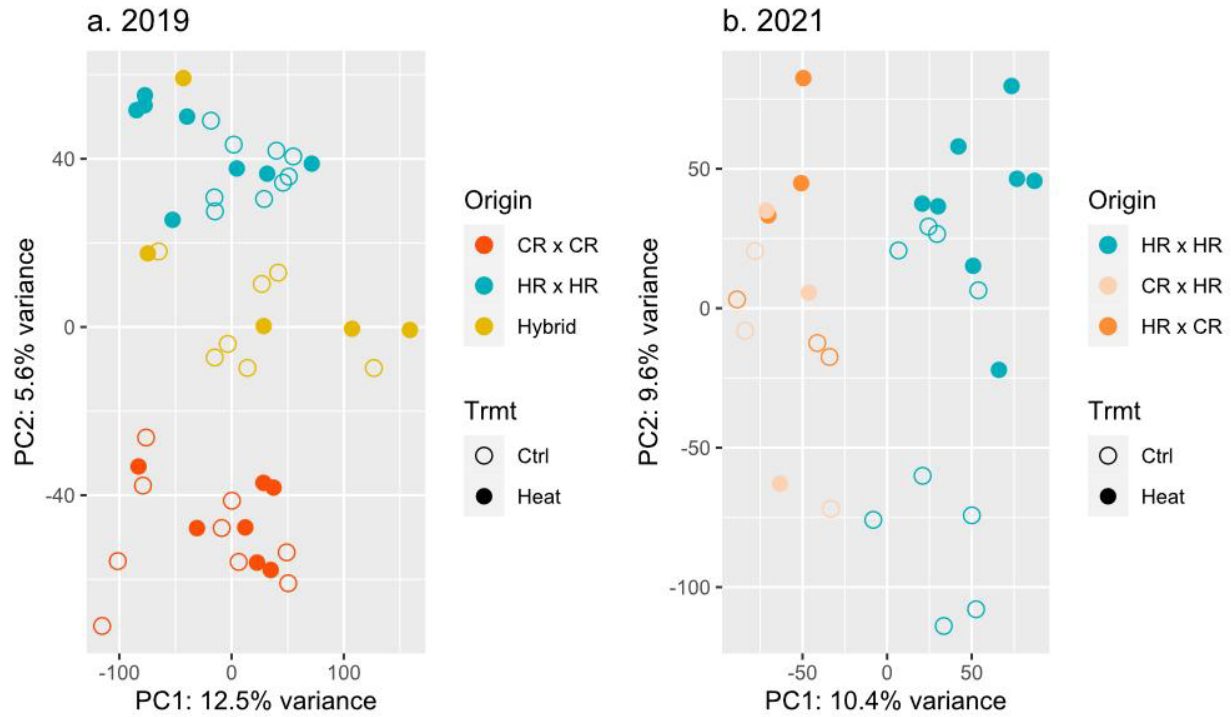


Fig. 4 Principal component analysis (PCA) on *rlog*-transformed read counts in a) 2019 and b) 2021 larval datasets. Points are colored by origin and shaped by treatment. The percentage variance explained by each PC is reflected on the axis label.

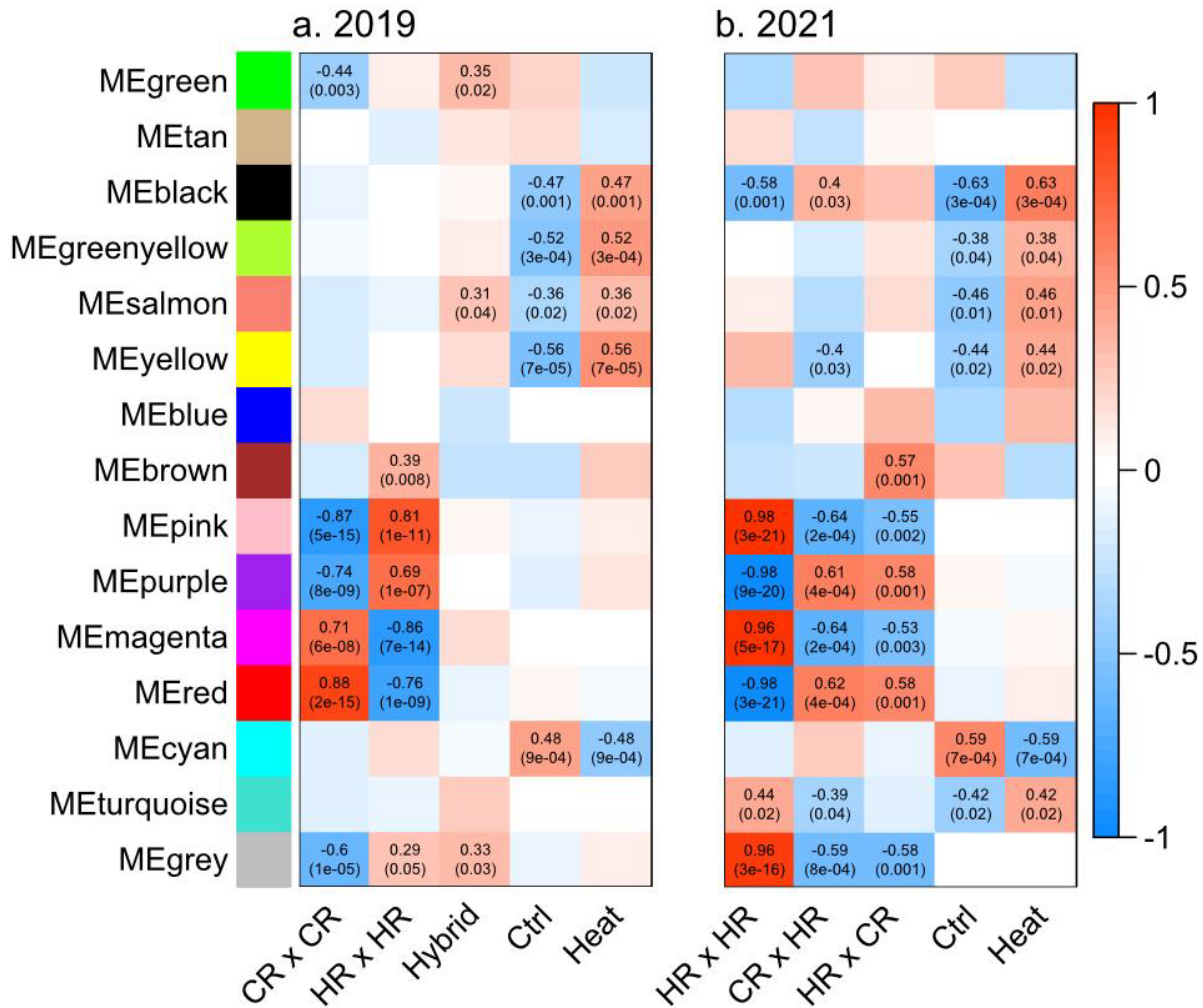


Fig. 5 Weighted gene co-expression network analysis (WGCNA) module-trait relationships identified in a) 2019 and b) 2021 larval cohort. Correlation values range from 1 (red) to -1 (blue) and the associated p values were included in the parenthesis below for modules showing a significant trait association, with color of each block determined by strength and direction of the correlation between given module and trait.

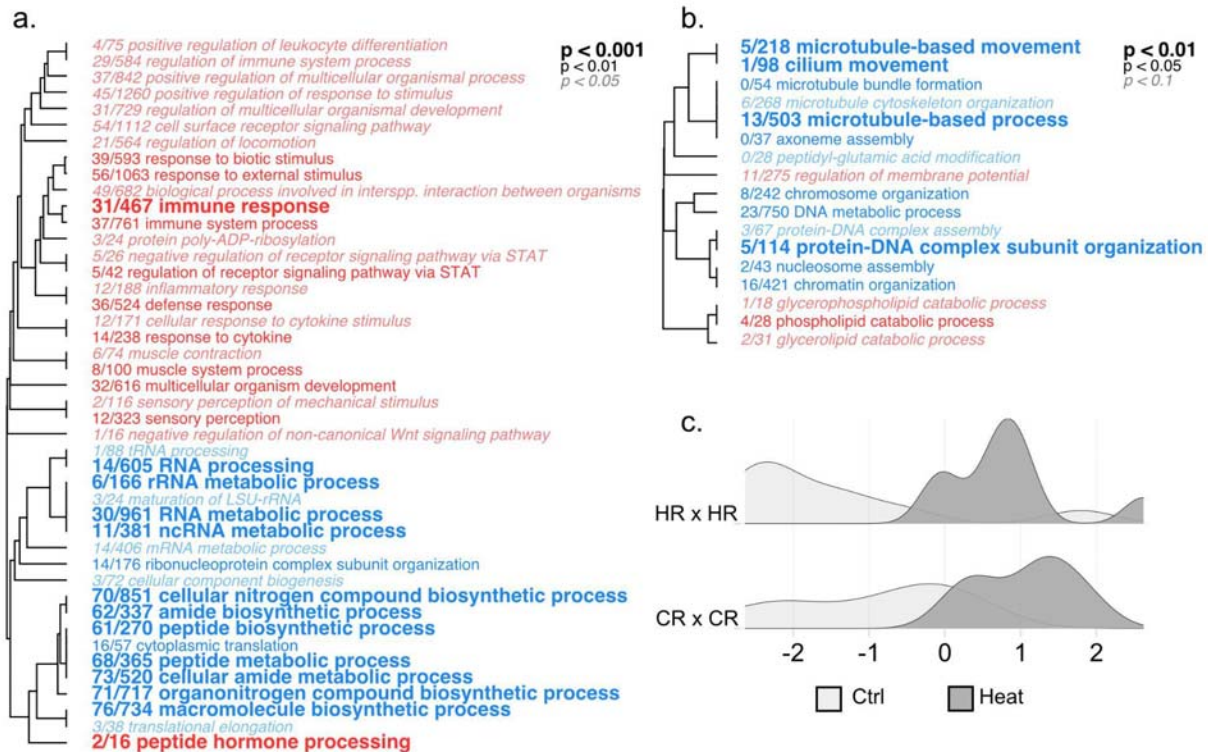


Fig. 6 Hierarchical clustering of ontology terms enriched by genes up-regulated (red) or down-regulated (blue) in 2019 heat-treated (a) CR x CR larvae and (b) HR x HR larvae compared to their respective untreated control, summarized by biological process (BP). Font size indicates level of statistical significance (FDR-corrected). Term names are preceded by fractions indicating the number of individual genes within each term differentially regulated with respect to treatment (unadjusted $p < 0.05$). (c) Density plots showing distribution of global expression across samples from the two origins along the temperature responsive axis (linear discriminant 2 [LD2], Fig. S7) based on discriminant analysis of principal components (DAPC) performed on variance stabilized data (VSD) grouped by treatment and origin.